

1 **A validation standard for Area of Habitat maps for terrestrial** 2 **birds and mammals**

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4 Prabhat R. Dahal^{1,2}, Maria Lumbierres^{1,2}, Stuart H. M. Butchart^{2,3}, Paul F. Donald^{2,3}, Carlo
5 Rondinini¹

6 **Affiliations**

7
8 1- Global Mammal Assessment Program, Department of Biology and Biotechnologies, Sapienza
9 University of Rome, Viale dell'Università 32, 00185 Rome, Italy

10 2- BirdLife International, David Attenborough Building, Pembroke Street, Cambridge CB2 3QZ,
11 UK

12 3- Department of Zoology, University of Cambridge, Downing Street, Cambridge CB2 3EJ, UK

13
14 corresponding author: prabhatraj.dahal@uniroma1.it
15

16 **Abstract**

17
18 Area of Habitat (AOH) is a deductive model which maps the distribution of suitable habitat at
19 suitable altitudes for a species inside its broad geographical range. AOH maps have been validated
20 using presence-only data for small subsets of species for different taxonomic groups, but no
21 standard validation method exists when absence data are not available. We develop a novel two-step
22 validation protocol for AOH which includes first a model-based evaluation of model prevalence
23 (i.e, the proportion of suitable habitat within a species' range), and second a validation using species
24 point localities (presence-only) data. We applied the protocol to AOH maps of terrestrial birds and
25 mammals. In the first step we built logistic regression models to predict expected model prevalence
26 (the proportion of the range retained as AOH) as a function of each species' elevation range, mid-
27 point of elevation range, number of habitats, realm and, for birds, seasonality. AOH maps with large
28 difference between observed and predicted model prevalence were identified as outliers and used to
29 identify a number of sources of systematic error which were then corrected when possible. For the
30 corrected AOH, only 1.7% of AOH maps for birds and 2.3% of AOH maps for mammals were
31 flagged as outliers in terms of the difference between their observed and predicted model
32 prevalence. In the second step we calculated point prevalence, the proportion of point localities of a
33 species falling in pixels coded as suitable in the AOH map. We used 48,336,141 point localities for
34 4889 bird species and 107,061 point localities for 420 mammals. Where point prevalence exceeded
35 model prevalence, the AOH was a better reflection of species' distribution than random. We also
36 found that 4689 out of 4889 (95.9%) AOH maps for birds, and 399 out of 420 (95.0%) AOH maps

37 for mammals were better than random. Possible reasons for the poor performance of a small
38 proportion of AOH maps are discussed.

39

40 **Introduction**

41

42 An accurate estimate of the distribution of species is central to ecological and conservation research
43 and action. There are three different classes of information on the distribution of species (Rondinini
44 and Boitani, 2006). These are 1) point localities (latitude and longitude) of individuals; 2)
45 geographic ranges, which are derived by mapping the extent of known point localities along with
46 expert knowledge; and 3) species distribution models, which use environmental and other relevant
47 variables associated with the species to refine geographical ranges. Species distribution models are
48 of two types (Stoms et al., 1992). The first are deductive models, which use expert-based
49 information on species' habitat use to model the suitable areas for the species. The second type are
50 inductive models, in which the environmental conditions at point localities where the species were
51 recorded are interpolated over wider areas.

52 Area of Habitat (AOH; also known as Extent of Suitable Habitat, ESH) is a deductive model which
53 maps the distribution of suitable habitat for a species inside its broad geographical range (Brooks et
54 al., 2019). It aims to reduce commission errors present in the range map while minimizing omission
55 errors. Several sets of AOH maps for different taxonomic groups at continental and global scales
56 have already been produced (Rondinini et al., 2005; Rondinini et al., 2006; Catullo et al., 2008;
57 Jenkins and Giri, 2008; Rondinini et al., 2011; Ficetola et al., 2015; Tracewski et al., 2016;
58 Lumbierres et al., 2021b).

59 Habitat models are prone to two major types of errors: omission errors occur when suitable habitat
60 areas for the species are wrongly mapped as being unsuitable, commission errors occur when areas
61 unsuitable for the species are wrongly mapped as being suitable. Quantification of these errors is
62 one of the key parts of the habitat modeling process and is done by validation. The omission and
63 commission errors could both be quantified only when independent presence and absence data on
64 the species are available. In such cases standard validation metrics such as True Skill Statistics
65 (TSS) (Allouche et al., 2006) and the Boyce Index (Boyce et al., 2002) are used. In case of AOH
66 maps produced for large taxonomic groups when true absence data are not available, no standard
67 validation method exists.

68 Rondinini et al. (2011) and Ficetola et al. (2015) used point localities from GBIF (Global
69 Biodiversity Information Facility) (www.gbif.org) to validate AOH maps for mammals and
70 amphibians respectively. AOH maps for South Asian mammals (Catullo et al., 2008) and African

71 vertebrates (Rondinini et al., 2005) were also validated using point localities. Brooks et al. (2019)
72 recommend using point localities for validation and inclusion of AOH maps for IUCN
73 (International Union for Conservation of Nature) Red List assessment. However, point localities are
74 often not available for many species and are biased towards certain taxonomic group and well-
75 studied areas.

76 In this paper, we developed a novel two-step validation protocol for AOH which includes: a) a
77 model-based evaluation of model prevalence (i.e., the proportion of a species' range that comprises
78 AOH), and b) a validation using species point localities (presence-only) data. We demonstrate the
79 use of this approach by validating a new set of AOH maps produced by Lumbierres et al. (2021b)
80 for all terrestrial birds and mammals. The validation method developed here is an iterative process
81 whereby systematic errors in the production of AOH (e.g. in the matching of habitat classes to land
82 cover maps) were identified using logistic regression models, then corrected where possible and a
83 new set of AOH maps produced. Then we employed a point validation analysis for the subset of
84 species for which point localities were available to assess the performance of the AOH maps.
85 Finally, we assessed the extent to which the subset of species for which point locality data were
86 available were representative of those for which no point data were available.

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88 **2. Methods**

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90 The new set of AOH maps (Lumbierres et al., 2021b) was produced at a resolution of 100 m using a
91 novel habitat-land cover model (Lumbierres et al., 2021a) which associated the different land cover
92 classes in the Copernicus land cover map (Buchhorn et al., 2019) with the Level-1 habitat classes of
93 the IUCN habitat classification scheme (IUCN, 2012). The IUCN habitat classification scheme is a
94 hierarchy of habitat classes, and each species assessed in the IUCN Red List is assigned to one or
95 more of these habitat classes, based on available information in the literature, unpublished reports
96 and expert knowledge. The habitat-land cover model (Lumbierres et al., 2021a) has the provision of
97 associating IUCN habitat classes to land cover classes using three different thresholds (1, 2 and 3).
98 Lower thresholds permit weaker associations between land cover and habitat classes. Therefore,
99 with threshold 1 each land cover class is associated with more habitat classes than with threshold 3.
100 Lumbierres et al. (2021b) produced a set of AOH maps for each of the three different thresholds by
101 clipping out of each species' range any cells of land cover that were not linked by the model to the
102 habitat class(es) to which the species was coded, then further clipping out parts of the range falling
103 outside the elevation range of the species.

104 In order to identify the best threshold among the three thresholds and to validate the set of AOH
105 maps with the best threshold at species level, we quantified two measures: ‘model prevalence’ and
106 ‘point prevalence’. Model prevalence is defined as the proportion of pixels inside the range that
107 were retained in the AOH. For example, if 25% of the pixels present in the original range map are
108 clipped out because they contain unsuitable habitat, fall outside the species’ elevation range or both,
109 the model prevalence is 0.75. Point prevalence is defined as the proportion of point localities (or
110 their buffers) out of all points inside the range of a species falling inside the suitable pixels. For
111 example the Red-tailed Comet (*Sappho sparganurus*) had a total of 71 point localities within its
112 range, of which 62 fell in pixels coded as suitable in the species’ AOH map, giving a point
113 prevalence of $62/71 = 0.88$.

114 Because the number of habitats associated with each land cover class decreases with increasing
115 thresholds, model prevalence is highest for threshold 1 models and lowest for threshold 3 models.
116 With increasing threshold, commission errors are expected to decrease (which is the main purpose
117 of AOH) but omission errors might increase. Our validation protocol therefore aimed to control for
118 omission errors. We did this by calculating point prevalence and model prevalence across the three
119 thresholds and identified the set of AOH maps for which the mean model prevalence was lowest
120 without compromising the mean point prevalence.

121 The point localities for bird species were downloaded from eBird (www.ebird.org), the largest
122 global repository for data on point localities of birds. eBird provides a metadata file called “eBird
123 basic data set” (Cornell Lab of Ornithology, 2020) which is a compilation of all the validated point
124 localities at species level and is updated monthly. These point localities are submitted by citizen
125 scientists as well as experts worldwide and are checked by local experts to remove obvious
126 misidentifications before they are made available for download (Sullivan et al., 2009). We first
127 downloaded the metadata file from eBird updated in January 2020 which was then queried in R (R
128 Core Team, 2018) using the *auk* package (Strimas-Mackey et al., 2018), as recommended by eBird,
129 to extract the point localities at species level. The taxonomy of Birdlife International (BirdLife
130 International and Handbook of the Birds of the World, 2020), which is that followed by the IUCN,
131 was matched with eBird’s taxonomy and point localities of only those species common to both were
132 queried and extracted from the metadata. Of the 10,813 species listed in Birdlife International’s list
133 for which AOH maps were produced, 9628 species matched by name. Of these 9628 species, 8998
134 species shared the same taxonomic concept and for 730 species the scientific names matched but
135 the taxonomic concept did not.

136 To ensure that only high-accuracy points were used for the validation, we selected the stationary
137 points from eBird’s metadata. The stationary points are those that have coordinate uncertainty of

138 less than 30 m. We then applied a temporal filter of 2019-2020 because the point localities from
139 2005-2018 were used to calibrate the habitat-land cover model in Lumbierres et al. (2021a). This
140 ensured there was no overlap between the calibration and validation data. The points were further
141 filtered by the range polygon of the species provided by the IUCN Red List website (IUCN, 2020)
142 to remove the small number of points falling outside the range (many of them likely to be
143 misidentifications). Since the AOH maps in question only include a certain combination of
144 presence, origin and seasonality of the range, we used the same combination to filter the point
145 localities. This ensured that we only included points which fell inside the boundaries of the selected
146 range maps. We also made sure that only one point locality was allowed per pixel of the AOH map
147 to avoid clustering of points. Finally, we excluded species which had fewer than 10 point localities
148 after all the filters were applied. A total of 4889 bird species had 4,836,141 point localities after
149 filtering. For mammals, point localities were downloaded from GBIF (Cold Spring Harbor
150 Laboratory, 2021) following the taxonomy of Global Mammal Assessment (which is followed by
151 IUCN) with same temporal and spatial filters as with birds except the filter of coordinate
152 uncertainty which was set to 300 m for mammals. This was done because far too many mammal
153 species would be excluded in the validation if we only considered point localities with coordinate
154 uncertainty of less than 30 m. The *rgbif* package (Chamberlain et al., 2021) in R was used to
155 download the points for mammals. A total of 107,061 point localities for 420 species were available
156 for mammals after applying all the filters.

157 A buffer of 300 m was applied around all the point localities to account for the positional
158 uncertainty of the points and for the fact that the location usually records that of the observer at the
159 time of observation and not the focal animal, following Jung et al. (2020). The buffers of point
160 localities were then overlaid on top of the AOH maps across all three thresholds at species level and
161 if at least one pixel coded to suitable habitat was found inside the buffer, the pixel was considered to
162 be validated at that point locality. The count of validated pixels was used to calculate point
163 prevalence at species level across all three thresholds.

164 We identified the threshold that produced a set of AOH maps for which the mean model prevalence
165 was lowest without detriment to the mean point prevalence.

166 We then employed a two-step approach to validate the set of AOH maps with the optimal threshold.
167 In the first step, we identified potential systematic errors in the AOH maps using a modeling
168 approach that aimed to identify species whose model prevalence was larger or smaller than
169 expected, given the characteristics of the species concerned. In the second step, we validated the
170 AOH maps using point localities following Rondinini et al. (2011).

171

172 2.1 A modeling approach to identify outliers

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174 We used logistic generalized linear models to predict model prevalence of the set of AOH maps
175 produced using the optimal threshold as a function of a number of independent variables, and
176 identified outliers whose observed model prevalence was significantly higher or lower than
177 predicted by the model. Outliers were then examined to identify systematic errors in, for example,
178 the way habitats were coded to land cover classes in the production of the AOH maps, and to
179 identify species that might be coded to the wrong habitats or elevation limits. For example, if a
180 species' range includes a high proportion of a particular land cover type not associated with the
181 suitable habitats of the species in the land cover-habitat association table (Lumbierres et al., 2021b),
182 or if errors in coding species to elevation limits mean that most of the range is outside the species'
183 stated limits, the model prevalence would be lower than predicted by the model.

184 The predictors fitted to the logistic models included: elevation range of the species (upper elevation
185 limit minus lower elevation limit), mid-point of the elevation range, number of habitats to which the
186 species is coded against in the IUCN Red List, seasonality of species (breeding and non-breeding
187 ranges in case of migratory birds) and the geographical realm of the species. In case of migratory
188 birds, Lumbierres et al. (2021b) has three different classes (resident, breeding and non-breeding
189 seasonalities) of AOH maps based on seasonality of the species. We merged resident seasonality to
190 breeding and non-breeding seasonalities to have AOH maps with only two seasonalities (breeding
191 and non-breeding). The dependent variable was the model prevalence of the AOH maps. Data from
192 a total of 10475 AOH maps for 9163 bird species (including for some species with separate
193 breeding and non-breeding ranges) and 2758 AOH maps for 2758 mammal species were used to
194 build logistic regression models for birds and mammals separately using the *lme4* (Bates et al.,
195 2015) package in R. Data on elevation were lacking for many mammal and bird species which is
196 the reason why not all species could be included in the logistic model. After testing taxonomic
197 genus, family and order as random effects in the model to control the non-independence of closely
198 related taxa, family was selected for fitting as the residual variance was lowest for the models with
199 family as the random effect for both birds and mammals. The predictive power of the model was
200 assessed by calculating marginal R^2 and conditional R^2 using the *insight* (Lüdecke et al., 2019)
201 package in R. The marginal R^2 expresses how much of the variation in data is explained by the fixed
202 effects and conditional R^2 tells how much of the variation in data is explained by both fixed and
203 random effects.

204 The Tukey fences outlier detection test (Wilcox, 2017) was used to identify outliers based on the
205 difference between the estimated and observed values of model prevalence. This test uses the

206 interquartile ranges to estimate the outliers in a data-set. The outlier test identified mild lower and
207 upper threshold values for the difference between estimated and observed values.

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209 *Mild upper threshold = (interquartile range * 1.5) + upper quartile*

210 *Mild lower threshold = lower quartile - (interquartile range * 1.5)*

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212 The AOH maps identified as mild upper outliers have an observed model prevalence much larger
213 than their predicted model prevalence, whereas maps identified as mild lower outliers have an
214 observed model prevalence much smaller than their predicted model prevalence.

215 In order to investigate the sources of errors in the outliers, we produced two more sets of AOH
216 maps for the outliers. One set included AOH maps which were produced by clipping the range of
217 the species by the altitudinal range only (AOH_{Elevation only}). Similarly, the other set included AOH
218 maps which were derived by clipping the range with only suitable habitat of the species (AOH_{Habitat}
219 only). If the model prevalence of an outlier was equal or nearly equal to the model prevalence of its
220 AOH_{Elevation only}, then we concluded that the under-representation of model prevalence could be
221 attributed to errors in elevation range of the species. If the model prevalence of an outlier was equal
222 or nearly equal to the model prevalence of AOH_{Habitat only}, then the source of error could be attributed
223 to the mapping of the habitats inside the range using the habitat-land cover crosswalk (Lumbierres
224 et al., 2021a) or to errors in the species' habitat coding. Furthermore, in some of the outliers the
225 under-representation could result from inclusion of large proportion of habitats which were
226 unsuitable for the species but were inside the range map of the species. Outliers do not necessarily
227 represent errors in AOH, as species might legitimately have very high or low model prevalence, but
228 by identifying suites of outliers sharing common characteristics we were able to identify and correct
229 a number of systematic errors in AOH production. The models also allowed us to identify species
230 whose AOH maps might be unreliable and whose habitat and elevation coding needs to be checked.

231

232 2.2 Point validation of AOH maps of terrestrial birds and mammals

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234 We validated 4889 bird and 420 mammal species' AOH maps using the filtered point localities. The
235 point validation was done by comparing the model and point prevalence at species level. If the point
236 prevalence exceeded model prevalence at species level, the AOH maps performed better than
237 random, otherwise they were no better than random. We also calculated the percentage of suitable

238 habitat pixels inside the buffers to ensure that the validation success wasn't due to one or few ~~one-off~~
239 pixels falling inside the 300 m buffer.

240 One of the major issues with citizen science data is that there is often a non-representative spread of
241 data across species. It is therefore possible that the species included in the point validation analysis
242 are not representative of the species not included. We assessed how representative the validation
243 sample size was by comparing the representation of variables such as family, order, genus, realm,
244 elevation range, mid-point of the elevation range, range size and extinction risk categories for birds
245 and mammals between species with and without point data. The point validation was done in R and
246 GRASS (GRASS Development Team, 2017).

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248 3. Results

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250 After comparing point and model prevalence of 4889 birds and 420 mammal species across all the
251 three thresholds, we selected the set of AOH maps derived by using threshold 3 in the habitat-land
252 cover model. At threshold 3, the mean model prevalence decreased as compared to thresholds 1 and
253 2 with much lower change in the mean point prevalence (Table 1 and 2) for both birds and
254 mammals.

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	Threshold 1	Threshold 2	Threshold 3
Mean model prevalence	0.81 ± 0.21 SD	0.77 ± 0.23 SD	0.65 ± 0.25 SD
Mean point prevalence	0.95 ± 0.14 SD	0.94 ± 0.14 SD	0.90 ± 0.17 SD

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257 **Table 1:** Mean model and point prevalence for AOH maps with standard deviation of 4889 bird
258 species across 3 different thresholds.

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	Threshold 1	Threshold 2	Threshold 3
Mean model prevalence	0.87 ± 0.21 SD	0.83 ± 0.22 SD	0.73 ± 0.24 SD
Mean point prevalence	0.95 ± 0.14 SD	0.95 ± 0.15 SD	0.93 ± 0.17 SD

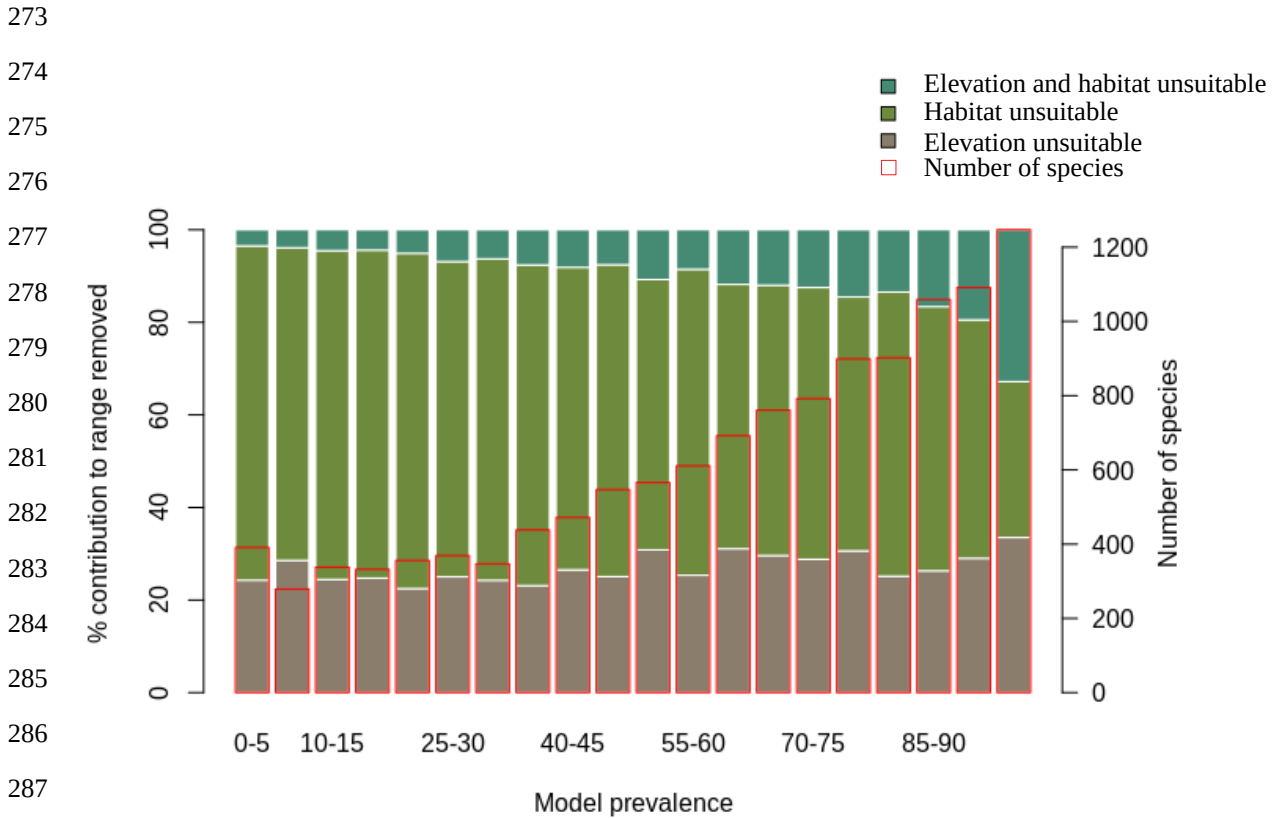
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261 **Table 2:** Mean model and point prevalence for AOH maps with standard deviation of 420 mammal
262 species across 3 different thresholds.

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264 We also assessed the relative contribution of elevation range, habitat, and both in reducing the range
265 to AOH. For both birds and mammals, most of the pixels removed from the range were because

266 either the habitat or the elevation were unsuitable, with a relatively small proportion being removed
 267 because both were unsuitable (Figs. 1,2). The proportion of the range that was clipped out on the
 268 basis of having unsuitable habitat at suitable elevations increased as model prevalence decreased,
 269 whereas there was little change across the same axis in the proportion of the range that was
 270 excluded on the basis of having suitable habitat at unsuitable elevations (Figs. 1,2). The number of
 271 both bird and mammal species peaked at model prevalence of 95-100% and gradually decreased as
 272 the model prevalence decreased.



289 **Figure 1:** Percentage contribution of elevation range, habitat and both in clipping the IUCN range
 290 to produce AOH maps for birds. Each bar represents a 5% bin of model prevalence, divided to show
 291 how much of the range was clipped out due to unsuitable habitat at suitable elevations (“Habitat
 292 unsuitable”), by suitable habitat at unsuitable elevations (“Elevation unsuitable”) and by unsuitable
 293 habitat at unsuitable elevations (“Elevation and habitat unsuitable”). The red blocks correspond to
 294 the second y-axis and show the number of species falling into each 5 % bin of model prevalence.

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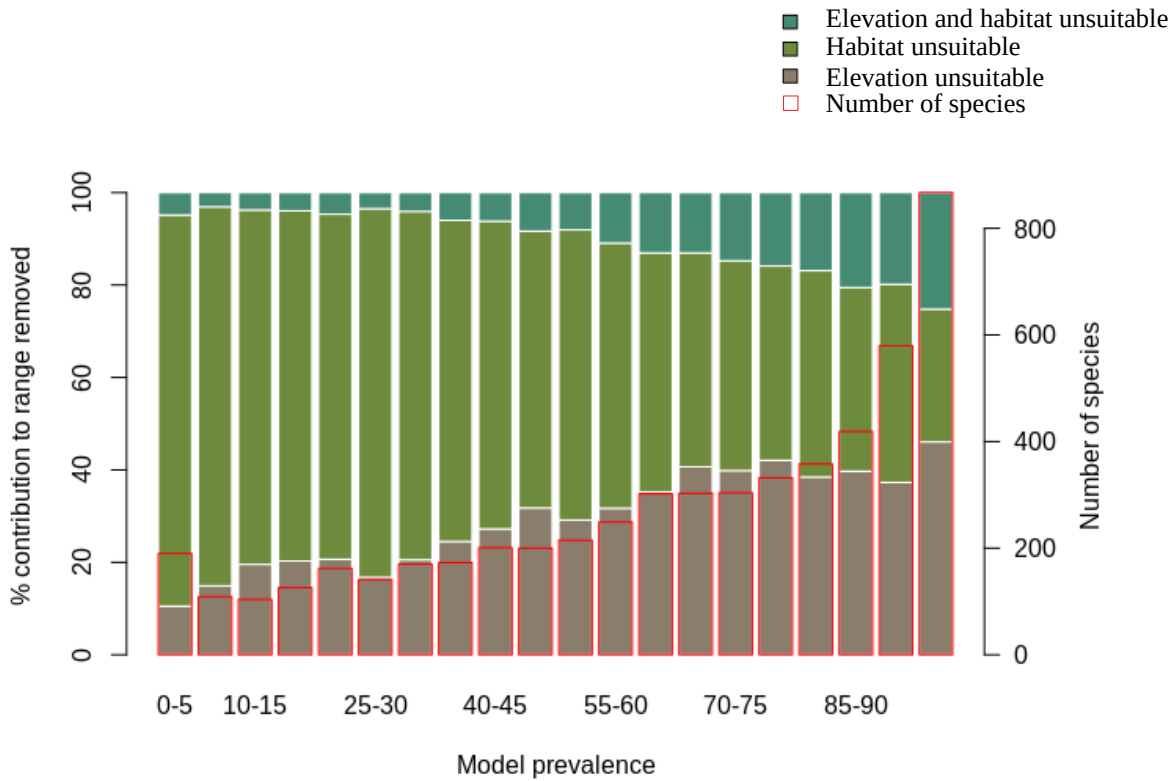


Figure 2: Percentage contribution of elevation range, habitat and both in clipping the IUCN range to AOH for mammals. See caption to Fig. 1 for interpretation.

For birds, the logistic model identified 178 AOH maps (1.7%) as lower outliers and 118 AOH maps (1.1%) as upper outliers out of 10475 AOH maps for 9163 terrestrial bird species. Similarly for mammals, the logistic model was applied to the AOH maps of 2758 species and identified 64 (2.3%) as lower outliers and 21 (0.8%) as upper outliers.

The mean of mid-point of elevation of the bird and mammal species identified as upper outliers was 2725 m and 3193 m respectively while the mid-point of elevation for species which were not identified as upper outliers was 1261 m for birds and 1289 m for mammals. This suggests that species identified as upper outliers were those found in higher elevation. These species were identified as upper outliers because the logistic models predicted low model prevalence at higher elevations. Also, the range maps for high-altitude species are drawn using contour maps, therefore most of the range is within the correct altitudinal band leading to high model prevalence for these species.

The lower outliers indicate where model prevalence was possibly underestimated due to potential errors in habitat mapping/coding and elevation range of the species. We found that the habitats “Shrubland” and “Savannah” in the habitat-land cover crosswalk were not associated with the land

334 cover class “Herbaceous cover”, leading to under-representation of these habitat types and hence
335 lower model prevalence than estimated by the logistic model (Fig. A1). We also found mismatch in
336 the elevation range and geographical range for the lower outliers (Fig. A2). There were few cases
337 where the range included large proportion of a particular land cover type which was not associated
338 with the suitable habitat of the species (Fig. A3). Moreover, we found that there was no land cover
339 information in the Copernicus land cover map for very small range polygons located on oceanic
340 islands which caused the AOH maps for these species to be empty. Furthermore, the land cover
341 class “open forest unknown” was discarded in the habitat land cover model. This led to low model
342 prevalence of AOH maps for some species whose ranges included this land cover. This was
343 corrected and a new set of AOH maps produced.

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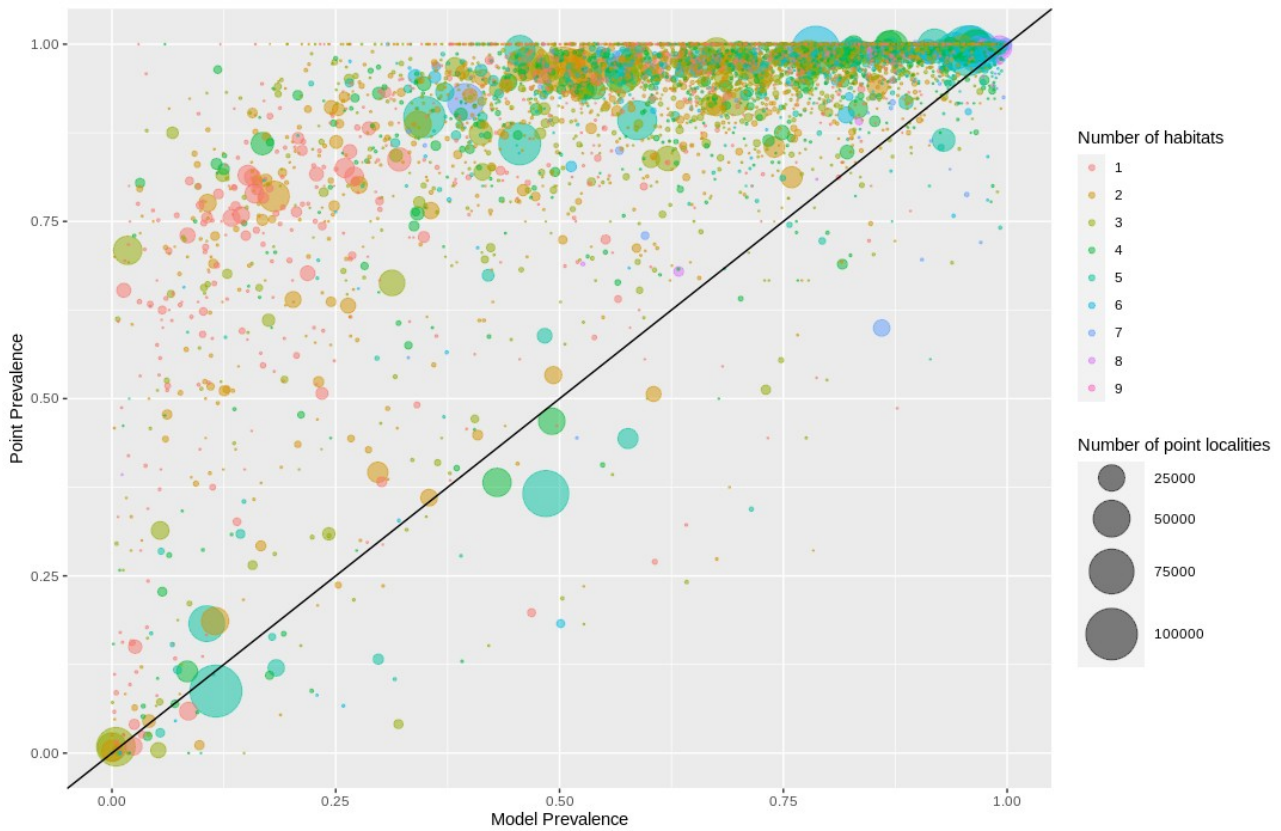
345 **Point validation**

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347 Out of 4889 bird species (45% of all bird species) for which point data were available, 4689
348 (95.9%) had higher point prevalence than model prevalence and 200 species had lower point
349 prevalence than model prevalence (Fig. 3). The mean percentage of pixels coded as suitable inside
350 the 300 m buffers of point localities of 4889 species of birds was 62% (Fig. A5).

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354 | **Figure 3:** Point prevalence vs model prevalence for terrestrial birds. Colors indicate the number of
 355 habitats each species is coded to, size of circles indicates the number of point localities.

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357 Out of 420 mammal species (8% of all mammal species) for which point data were available, 399
 358 (95.0%) had point prevalence higher than model prevalence (Fig. 4). The mean percentage of pixels
 359 coded as suitable inside the 300 m buffers of point localities of 420 species of mammals was 78%
 360 (Fig. A5).

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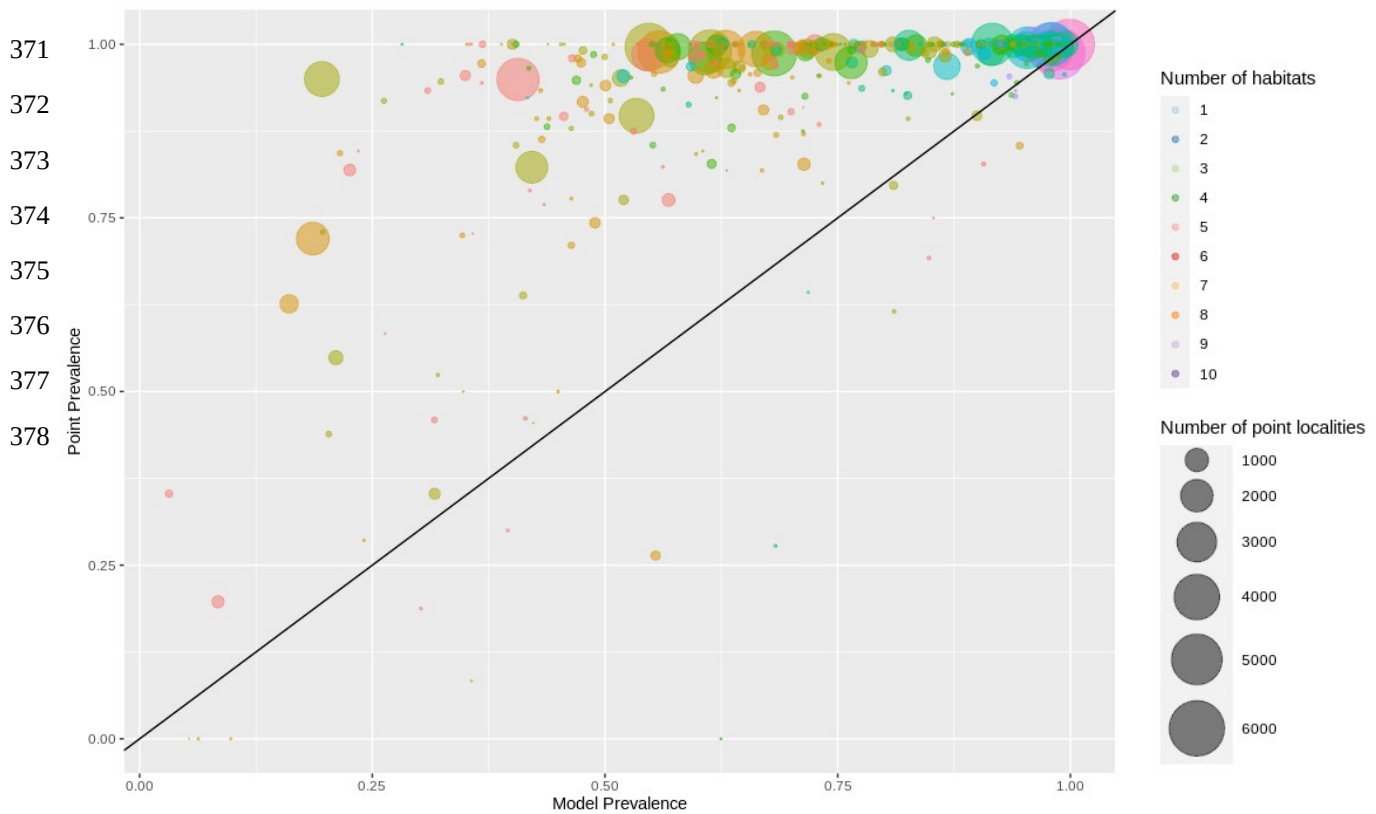
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380 | **Figure 4:** Point prevalence vs model prevalence for terrestrial mammals. Interpretation as in Fig. 3.

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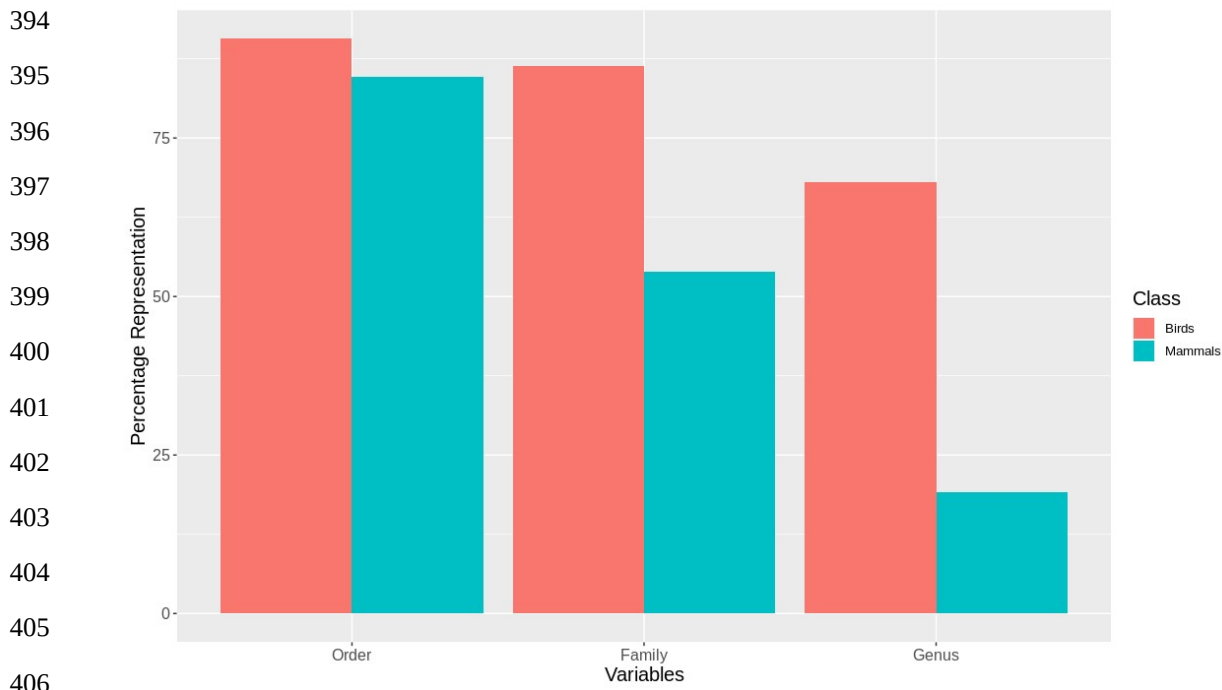
382 **Representativeness of validation sample**

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384 We found that for birds over 60% all families, genera and orders were represented in the sample
 385 included in the point validation and species from all biomes were represented but representation for
 386 mammals was lower, as expected due to the much lower proportion of mammal species for which
 387 point locality data were available (Fig. 5).

388 The validation points were spread across all of the variables and majority of their sub-classes (Fig.
 389 A6, Fig. A7). Species with validation points tended to have larger range sizes, wider elevation
 390 ranges and to be coded to more habitat classes than those without. Furthermore, validation points
 391 were not available for any critically endangered or endangered mammals as these species are rare in
 392 the wild.

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Figure 5: Taxonomic representativeness of validation sample for birds and mammals.

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408 Discussion

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411 On comparing our point validation results with previous validation analysis of AOH maps, we
412 found that validation results are similar to or better than previous exercises. For mammals,
413 Rondinini et al. (2011) evaluated AOH maps for 263 species at 300 m resolution, of which 241
414 (91.6 %) were better than random as compared to 95.0% in our analysis. However, it should be
415 noted that the mean model prevalence for AOH maps of Rondinini et al. (2011) was 54.8 ± 21.5 SD
416 as compared to 65.16 ± 25.42 for our AOH maps. The ratio of mean point prevalence to mean
417 model prevalence for Rondinini et al. (2011) was 1.4 compared to 1.38 in our case. Ficetola et al.
418 (2015) found that AOH for 94% of 115 amphibian species used in the validation analysis were
419 better than random with the mean model prevalence for species with validation points being $0.79 \pm$
420 0.21 SD. The ratio of mean point prevalence to mean model prevalence was 1.18 in this case.

421 Moreover, Catullo et al. (2008) found that 140 AOH maps out of 190 (73.7 %) South Asian
422 mammal species gave positive validation results while Rondinini et al. (2005) found the mean
423 proportion of suitable habitats correctly mapped inside the range for 181 species of African
424 vertebrates was 0.55 ± 0.01 SE using presence-absence data sets. The high validation success in our
425 analyses could be attributed to the use of novel habitat-land cover model (Lumbierres et al., 2021a),
426 the use of logistic regression models to identify systematic errors and the larger validation sample
427 as compared with previous exercises. Furthermore, the underlying land cover map used in

428 Lumbierres et al. (2021b), has the highest resolution among the global land cover maps providing
429 **with** more detailed land cover classification.

430 The point validation identified a small proportion of AOH maps which were no better than random.
431 Some of these had high model prevalence. In such cases, point prevalence must be exceptionally
432 high for the models to be better than random since even if a majority of point localities fall within
433 the AOH these maps may perform no better than random. For the AOH maps which were no better
434 than random and had low point prevalence, this was usually due to an apparent error in the coding
435 of elevation range of the species, the areas inside the range of the species where the point localities
436 fell being clipped out by what was assumed to be an erroneous elevation range. A list of species
437 with probably erroneous elevation coding will be forwarded to IUCN Red List team for future
438 corrections.

439 AOH maps aim to minimize the commission errors known to be present in species ranges without
440 increasing omission errors (Rondinini and Boitani, 2006). One of the limitations of this validation
441 analysis is the inability to quantify the commission errors of the AOH maps as we don't have the
442 true absence data of the species. Therefore, some uncertainty remains in AOH maps regarding the
443 commission errors.

444 Also, there are some intrinsic errors in the models as identified by the logistic regression analysis.
445 The species which are coded only to habitats like "Shrubland" might have under-represented model
446 prevalence as discussed above. However, the number of AOH maps identified as lower outliers by
447 the application of the logistic model was low for birds (178/10475) and for mammals (64/2758),
448 indicating that for the majority of AOH maps the observed model prevalence was fairly close to that
449 predicted by the model.

450 The AOH maps validated in this paper is the largest validation done till date in terms of number of
451 species validated for birds and mammals. These maps will be freely available after the publication
452 of Lumbierres et al. (2021b). We have also provided the metadata for all the species along with
453 validation statistics in this paper which can be used as a guideline by the users while using the AOH
454 maps.

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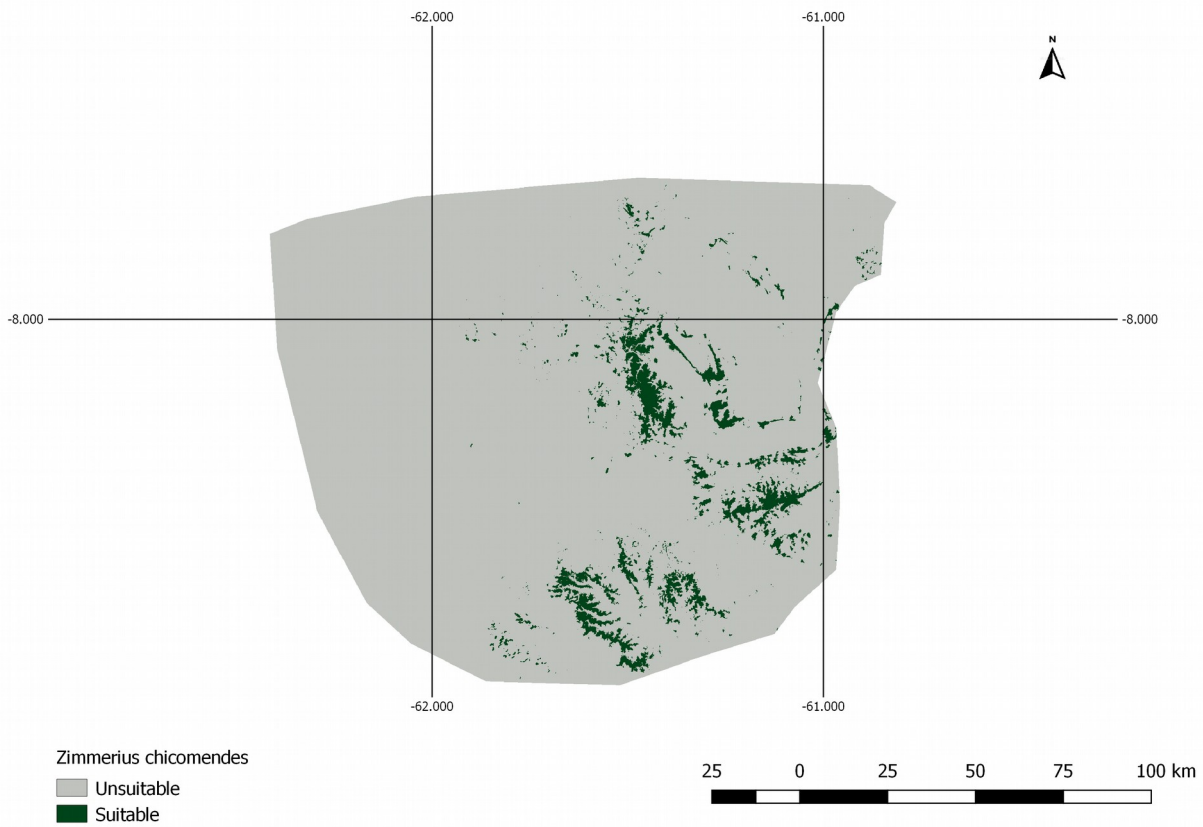
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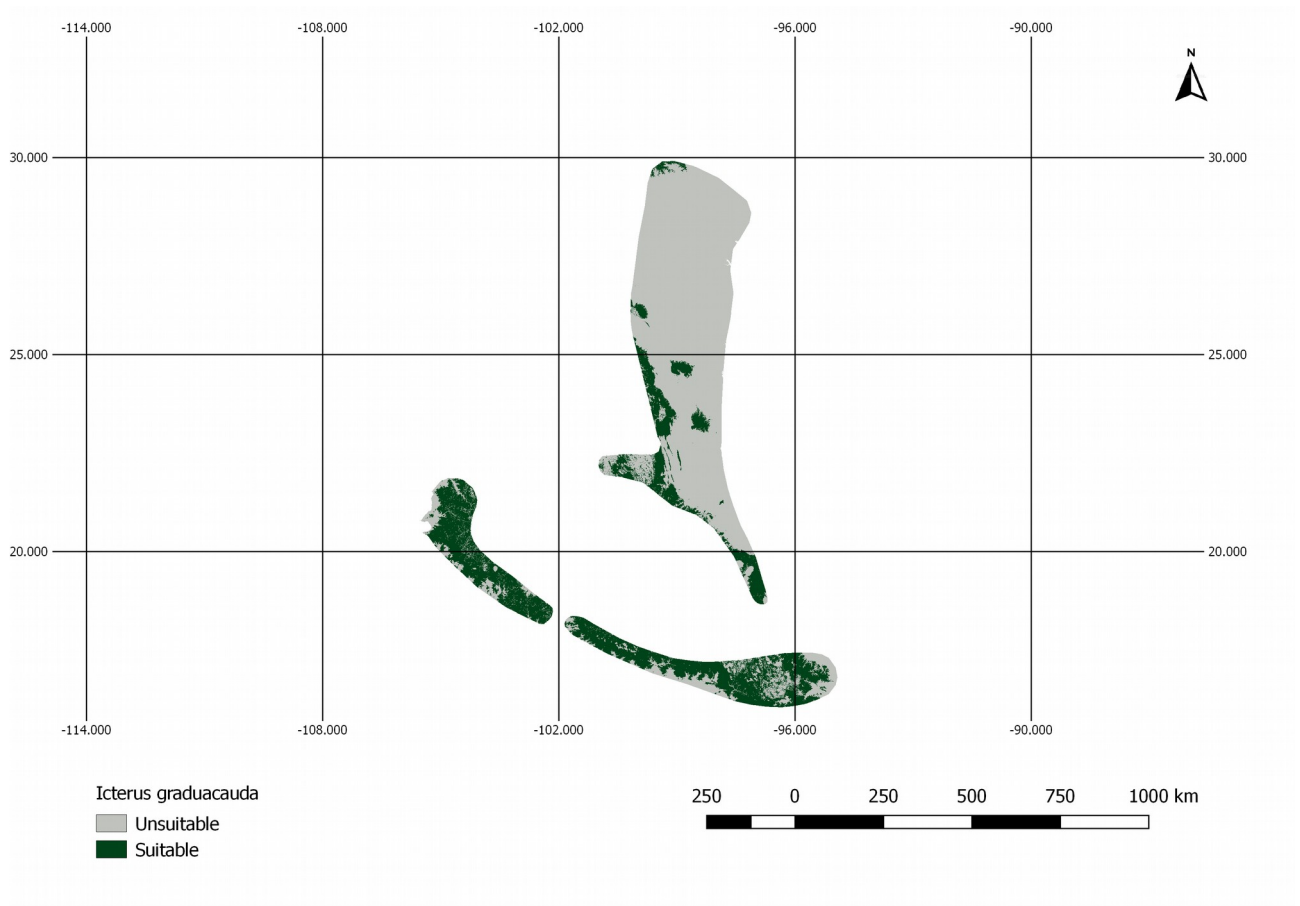
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463 **Appendix A**

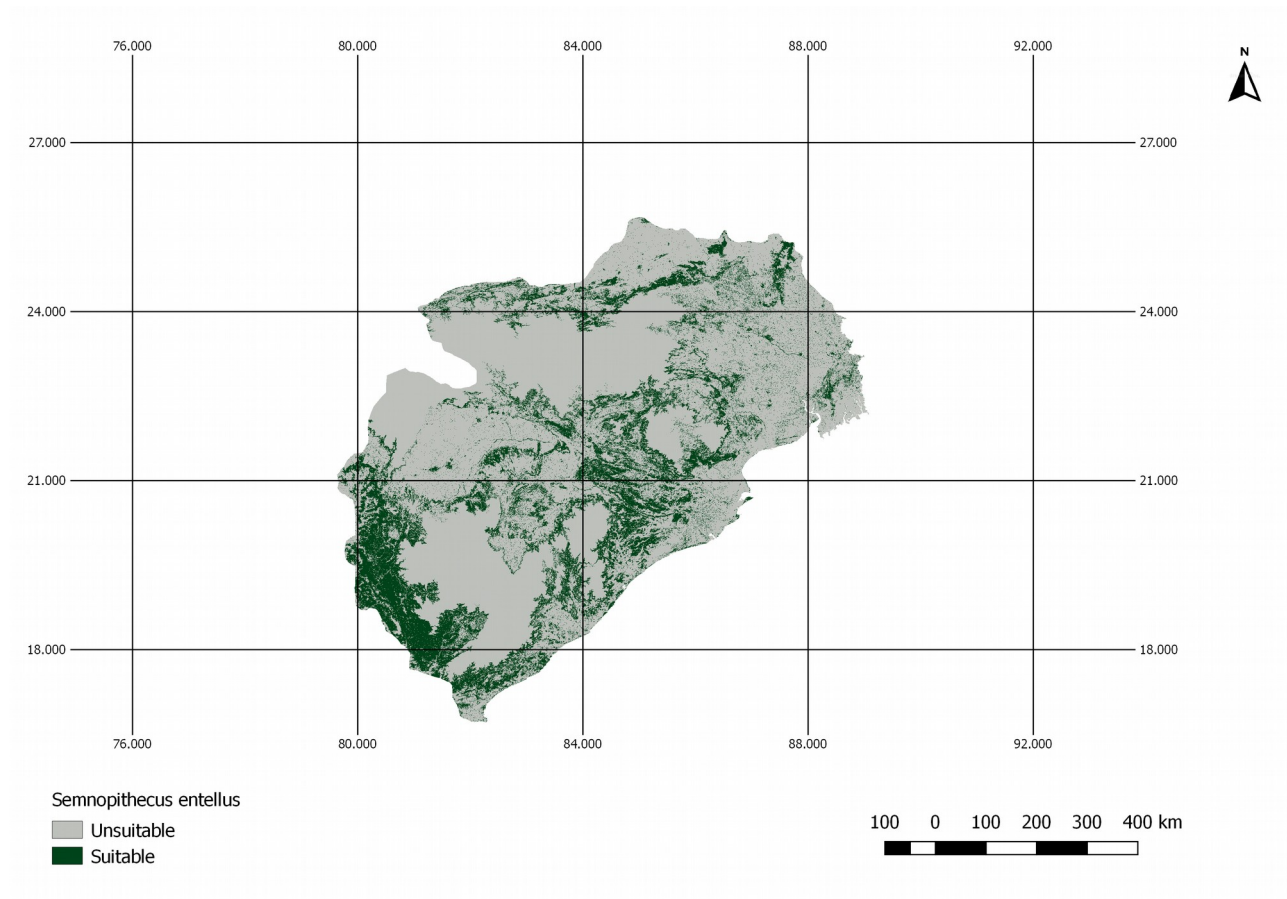
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466 **Figure A1:** AOH map for species *Zimmerius chicomendes*. The species is coded against “Forest”
467 and “Shrubland” habitats and the elevation range falls inside the IUCN range. However, the land
468 cover inside this range map includes a high proportion of “Herbaceous cover” land cover type
469 which is not associated with “Shrubland” habitat in the habitat – land cover association table.
470 Therefore, the model prevalence of this AOH is much lower than expected.
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473 **Figure A2:** AOH map for the species *Icterus graduacauda*. The IUCN range of the species doesn't
 474 cover much of the elevation range. Therefore, the model prevalence of this species is lower than
 475 estimated.
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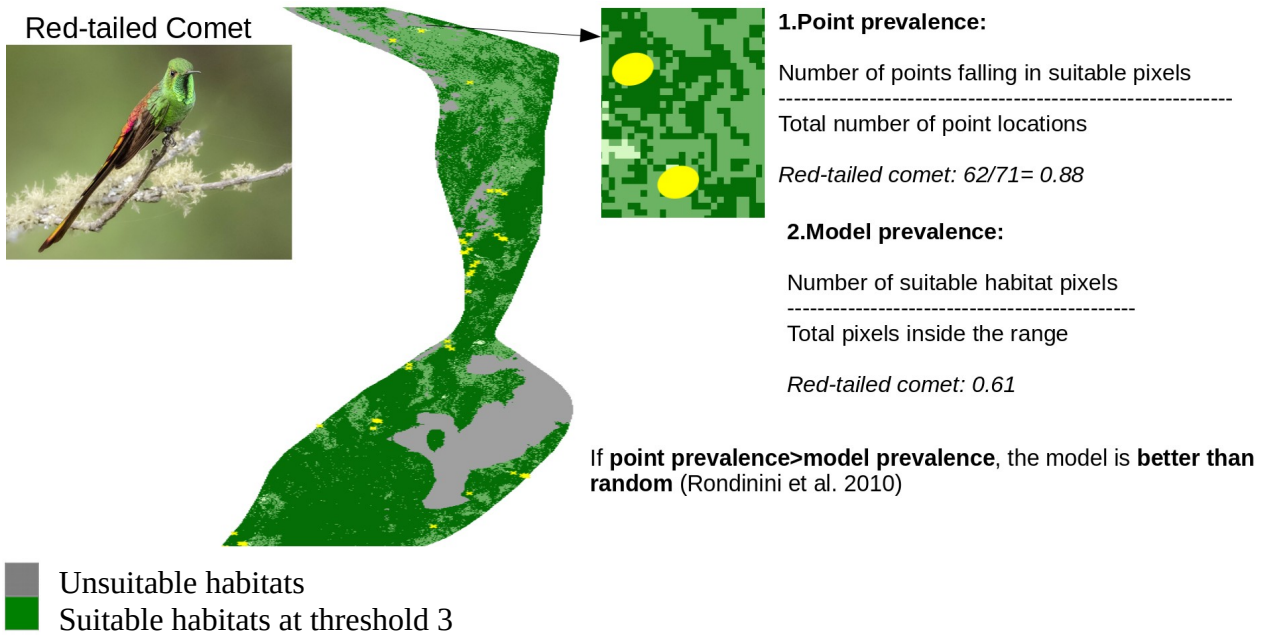


478 **Figure A3:** AOH for the species *Semnopithecus entellus*. There is a large proportion of land cover
 479 class “Cropland” inside the range map of this species. However, this species is not coded to habitats
 480 that are associated with the land cover “Cropland”. Therefore, the model prevalence is lower than
 481 estimated.

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499 | **Figure A4:** Point validation of the AOH map of the species Red-tailed Comet using model and
 500 point prevalence. The yellow circles represent the buffered point localities of Red-tailed Comet.

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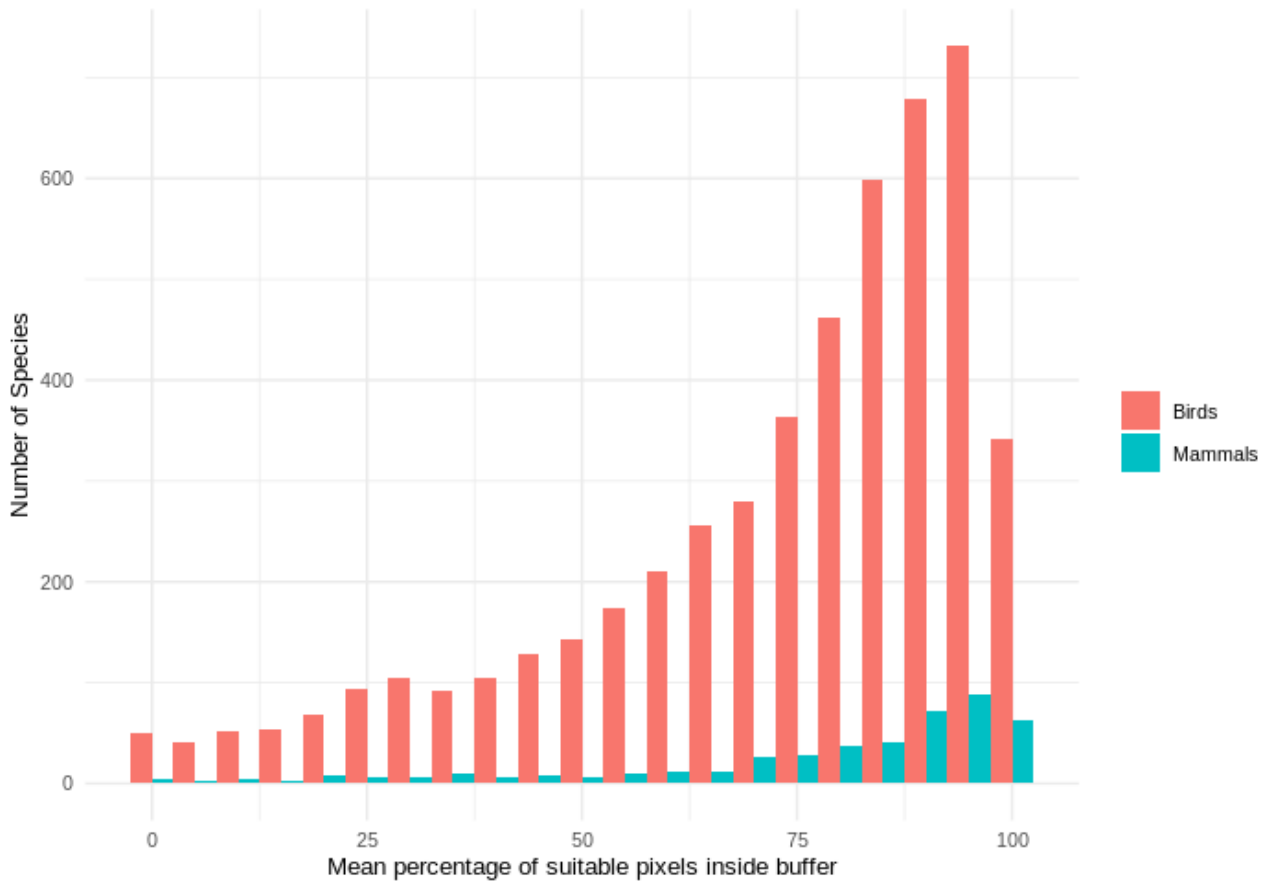
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514 **Figure A5:** Histogram of mean percentage of suitable AOH pixels inside the 300 m buffer for
 515 mammals and birds species used in point validation.

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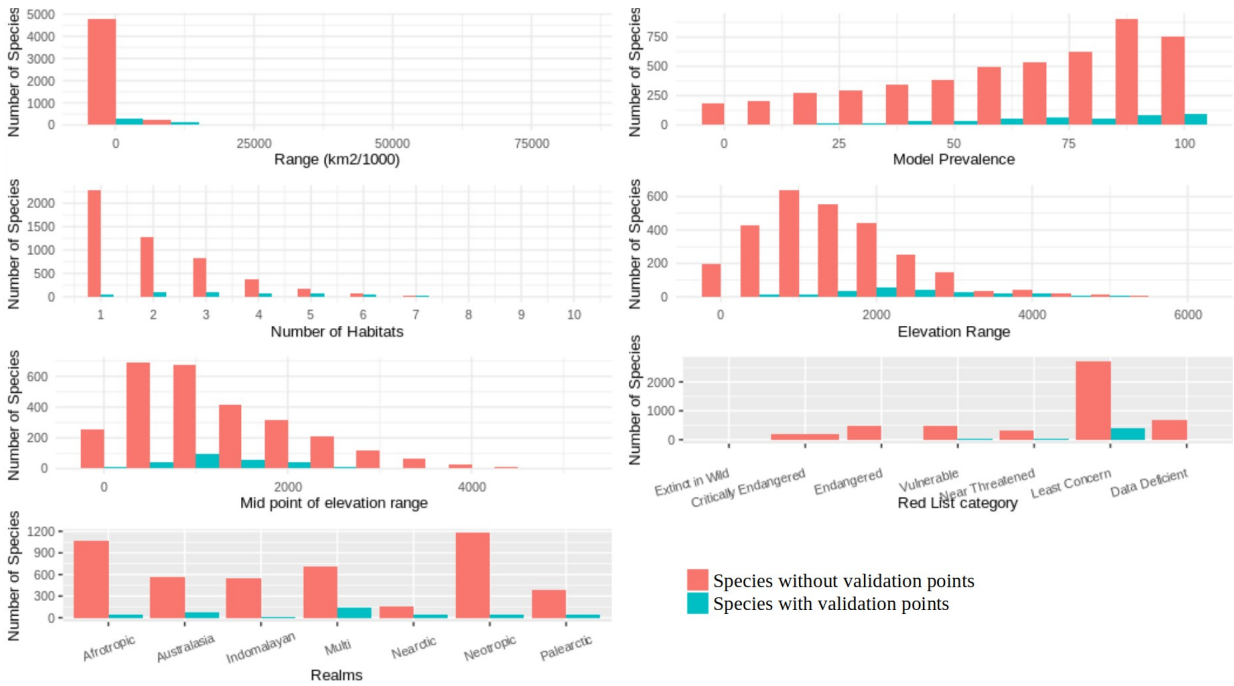
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545 | **Figure A6:** Comparison of species with and without validation points for mammals.

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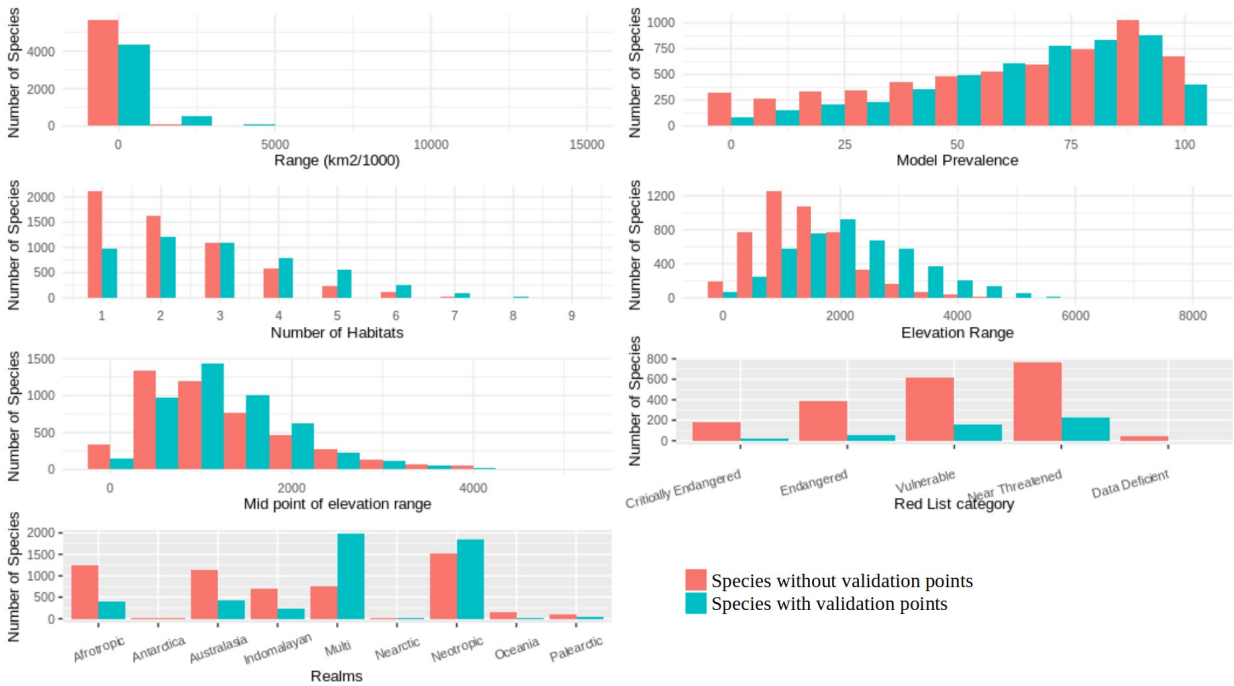
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561 | **Figure A7:** Comparison of species with and without validation points for birds.

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564 **Data and code availability**

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566 The point localities used in the validation analyses along with the metadata tables summarizing the
567 validation analyses can be found at <http://doi.org/10.5281/zenodo.5109073>. The same DOI can be
568 used to access the code used for validation and to also access some sample AOH maps which were
569 validated.

570

571 **Author contribution**

572

573 PRD PFD and CR conceptualized the idea. PRD and ML curated and did the formal data analysis.
574 PRD led the manuscript writing with contributions from all the authors. PFD CR SHMB supervised
575 the whole process.

576

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578

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