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PREFACE

The Peter Cooper Landfill Site Remedial Investigation (RI) Report was originally submitted to the United States Environmental Protection Agency (USEPA) in December 2002. This Final RI Report incorporates responses to USEPA comments dated July 18, 2003 and subsequent comments dated October 29, 2003. For completeness, USEPA comments and respondent responses are included in Appendix R.

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REMEDIAL INVESTIGATION REPORT Peter Cooper Gowanda Site Gowanda, New York

1.0 INTRODUCTION

1.1 SITE LOCATION AND DESCRIPTION

The Peter Cooper Landfill National Priority List (NPL) Site, hereinafter referred to as the "Peter Cooper Gowanda Site" or the "Site," is comprised of an inactive landfill area and former animal-glue and adhesives manufacturing plant located on approximately 26 acres of property between Palmer Street and Cattaraugus Creek (i.e., the Creek) in the Village of Gowanda, Cattaraugus County, New York (Figures 1-1). The Site is bordered to the north by Cattaraugus Creek, to the south by Palmer Street, to the west by a former hydroelectric dam and wetland area, and to the east by residential properties. The former office, laboratory, plant water reservoir and employee parking lot are located south of Palmer Street on an approximately 20 acre parcel. This parcel, hereafter referred to as the "office parcel", is currently owned by a private interest and is not part of the NPL Site.

The Inactive Landfill Area is situated on the western side of the Site (Plate 1), with the western edge of the landfill located on property owned by NYSEG. Specifically, the Inactive Landfill Area is an approximately 15.6-acre area bordered to the north and south by Cattaraugus Creek and Palmer Street, respectively; to the west by a former hydroelectric dam and the approximate western limit of contiguous wetlands; and to the east by a line running approximately perpendicular to Cattaraugus Creek from the former sluiceway to Palmer Street. The portion of the Inactive Landfill Area that contains waste fill encompasses only an approximate 5-acre sub-area (i.e., the elevated fill area) in the northwest corner of the Site.

The Former Manufacturing Plant Area is located on the eastern side of the Peter Cooper Site (Plate 1), and includes the remaining 10.4-acre portion of the Site outside the Inactive Landfill Area. The Former Manufacturing Plant Area is bounded on the north by Cattaraugus Creek, on the south by Palmer Street, on the west by the Inactive Landfill Area and the east by the residential property boundary.



1.2 SITE HISTORY

1.2.1 Historic Operations

The Peter Cooper Site was previously used to manufacture animal glue and industrial adhesives. Peter Cooper Corporation (PCCI) and/or its predecessors, Eastern Tanners Glue Company and successors (Rousselot Gelatin Corporation (PCCII)), manufactured animal glue at the site from 1904 to 1971 and adhesives from the 1950s until the plant closed in 1985. Animal glue manufacturing operations were reportedly closed by the early 1970s. The northwest portion of the Inactive Landfill Area was reportedly used to dispose of residuals from the animal glue manufacturing process, commonly referred to as cookhouse sludge. The cookhouse sludge was derived from animal hides, some of which were chrome-tanned, used as a feedstock in the process. Based on observations of the landfill sludge material made during this RI, the cook house sludge appears to be mixed with cinders, ash, and construction and demolition debris. This sludge mixture is referred to in this report as sludge fill.

Benchmark's review of historic (1924 and 1948) fire insurance (Sanborn) maps and aerial photos from 1939, 1956,1966, 1973, 1980, 1983 and 1990 indicates that the Former Manufacturing Plant Area was substantially covered by buildings and support structures throughout its operational history. The historic Site features and manufacturing process areas present on a 1948 Sanborn map are shown in Appendix A. The 1980 aerial photo for the Site indicates that animal glue manufacturing facilities were decommissioned/demolished at that time.

In June 1971, the New York State Supreme Court ordered PCCI to remove all or part of the waste pile and terminate discharges into Cattaraugus Creek. In response, PCCI reportedly removed approximately 38,600 tons of waste pile material to its Markhams, New York site in early 1972. Between 1972 and 1975, the remaining waste pile at the site was graded, covered with a 6" clay barrier layer and 18-30 inches of barrier protection soil, and vegetated with grass (O'Brien & Gere, 1991). Stone rip-rap and concrete blocks were placed along the bank of Cattaraugus Creek to protect the fill material from scouring.

In July 1976, the assets of original PCCI, including the manufacturing plant and property located in Gowanda, were purchased by Rousselot Gelatin Corporation and its parent, Rousselot, S.A., of France. Rousselot Gelatin was renamed Peter Cooper Corporation (PCCII) and this newly-formed PCCII sold the Gowanda site to the current owner, JimCar Development, Inc., in April 1988.

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1.2.2 Previous Investigations and Remedial Measures

NYSDEC performed Phase I and Phase II Site Investigations at the Peter Cooper Gowanda Site in 1981 and 1983, respectively (RCRA Research, Inc. 1983 and 1984). The Phase I included limited soil and seep sampling performed in November 1981.

The Phase II investigation was performed in May 1983 and included the investigation of soil, groundwater, surface water and sediment. Samples were analyzed for total halogenated organics and total volatile halogenated organics, as well as priority pollutant metals. Analytical results indicated the presence of arsenic, chromium and zinc in soil and sediment samples. Surface water and groundwater inorganic analyses were non-detect with the exception of low levels of chromium in groundwater. Phase I and II analytical results are presented in Appendices B-1 and B-2.

The current PCCII subsequently agreed with the New York State Department of Environmental Conservation (NYSDEC) to perform a Remedial Investigation and Feasibility Study (RI/FS) at the site. The RI was performed by O'Brien & Gere Engineers under a NYSDEC-approved work plan. Activities performed during the RI included collection of soil, surface water, sediment, waste material, seep, and groundwater samples. The 1989 RI data is presented in Appendix B-3. Most of the O'Brien & Gere investigation targeted the Inactive Landfill Area. The RI Report was issued in January 1989. The RI concluded that there were no significant health risks associated with the site.

The FS Report was issued in March 1991. In June of 1991, NYSDEC and PCCII reportedly agreed upon a remedial alternative for the site that included containment of source materials, leachate collection and access restrictions through fencing and deed restrictions.

In 1991 NYSDEC removed the site from its Registry of Inactive Hazardous Waste Sites because it did not meet the statutory definition of an inactive hazardous waste disposal site. As a consequence of this designation, NYSDEC could not use State resources to implement a remedial program. NYSDEC and the Village of Gowanda reportedly requested EPA to evaluate the site for NPL listing.

In 1996, United States Environmental Protection Agency (USEPA) Region II activated the Response Engineering and Analytical Contract (REAC) and the Superfund Technical and Assessment Response Team (START) to collect and analyze soil, groundwater, surface water,

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and sediment samples from the Peter Cooper Site. Results from the 1996 sampling event are presented in Appendix B-4 and discussed in further detail in Section 1.2.3.

In 1997, New York State Electric & Gas Corporation (NYSEG), under an order on consent with USEPA, placed an approximately 150-foot long rip-rap revetment adjacent to Cattaraugus Creek on the portion of the site owned by NYSEG (i.e., northwest portion of the site).

In 1998, USEPA Region II prepared a Hazard Ranking System Model score for the site and listed the Peter Cooper Site on the NPL. USEPA subsequently notified several potentially responsible parties (PRPs) of their possible involvement in the site investigation and remediation, and proceeded to develop a Remedial Investigation/Feasibility Study (RI/FS) Work Plan for the site. The Revised Final RI/FS Work Plan was issued by USEPA on June 15, 1999. Representatives of certain PRPs subsequently met with USEPA and volunteered to prepare a modified RI/FS Work Plan addressing the Inactive Landfill Area of the site. The final RI/FS work plan for the Inactive Landfill Area was submitted to USEPA in March 2000. In April 2000 USEPA issued a Unilateral Administrative Order to certain PRPs directing completion of the RI/FS for the Inactive Landfill Area as well as the Former Manufacturing Plant Area. Representatives of the cooperating PRPs subsequently submitted an Addendum to the March 2000 RI/FS Work Plan that extended RI/FS activities to the Former Manufacturing Plant Area. The Addendum was approved in August 2000.

1.2.3 Chemical Constituents Historically Detected in Site Media

The following subsections describe the results of historic sampling programs that we used to characterize the nature and distribution of chemical constituents in site media at the Site.

1.2.3.1 Landfill Waste

The 1989 RI Report describes the collection and analysis of six landfill waste material samples. Each sample was obtained from below the landfill cap and consists of black, cindery waste/fill material. Low concentrations of organic compounds were detected in these samples including chlorobenzene (0.06 mg/kg) and 2-butanone (0.14 mg/kg). Average concentrations for inorganic compounds in these samples were 13,000 mg/kg for total chromium, 9.8 mg/kg for arsenic, and 620 mg/kg for zinc. EP Toxicity testing was conducted on each of the samples. Analytical results indicate that the concentrations were below EP Toxicity criteria (40CFR 261.24). The 1989 RI results indicate that materials present in the landfill are consistent with that expected from the production processes for the facility. A summary of the inorganic waste source results for the RI is provided in Appendix B-3.



During the 1996 USEPA investigation, six waste samples were collected at depths of 1 to 3 feet below the ground surface (bgs). The samples were analyzed for chromium, hexavalent chromium, and arsenic. The locations of these samples are shown in Appendix B-4. Results of inorganic analyses for soil samples are also provided in Appendix B-4. Concentrations of arsenic ranged from 4.6 to 33 mg/kg. Total chromium concentrations ranged from 2,900 mg/kg for sample HA MW-03 to 37,000 mg/kg at sample location HA SB-71DUP. Hexavalent chromium was not detected in any of the waste samples.

The six waste samples collected during the 1996 USEPA investigation also were analyzed for semi-volatile organic compounds (SVOCs). A summary of the analytical results for these samples is provided in Appendix B-4 (Landfill Waste Samples). Compounds detected in the waste included low concentrations of PAHs and phenolic compounds. A number of tentatively identified compounds (TIC) in the alkane group were found at estimated concentrations up to 0.26 percent by weight. These compounds are likely associated with the anaerobic decomposition of processed animal hide material.

1.2.3.2 Chemical Constituents in Soil

Four rounds of surface soil sampling were conducted during the 1989 RI. Surface soil samples were collected near the landfill and in the area of the former factory area. Sample locations are shown in Appendix B-3.

The initial sampling, in September 1986, consisted of the collection of 20 samples for arsenic, total chromium, and hexavalent chromium analysis. The second round of sampling was conducted in April 1987, and consisted of the collection of 13 samples. During this event, each sample submitted for analysis was a composite of four grab samples collected equidistant from the sampling point (O'Brien & Gere, 1989). The third and fourth rounds of sampling were performed during July and August 1988. The third sampling event included 10 waste material samples collected below the landfill cap. The fourth round included nine sampling points located in the area adjacent to the old concrete dam.

Background soil concentrations for arsenic, zinc, total chromium and hexavalent chromium were established by O'Brien & Gere by doubling the mean background concentration for each of these compounds to provide a basis for evaluating sample concentrations. A total of 46 surface soil (depth of 0-3 inches below grade); 21 shallow subsurface soil (9 to 12 inches) and one subsurface soil (33 to 36 inches) sample were collected and analyzed. Three of the 46 surface soil samples were collected from background sample locations south of Palmer Street



approximately 2000 feet south of the landfill. Arsenic concentrations detected in the soil samples fell within the calculated background range with only one exception. The reported value of 417 mg/kg at location 30 (see Appendix B-3) was over ten times higher than any other sample result and was considered as anomalous (O'Brien & Gere, 1989).

The 1989 RI Report indicated that 33 of 65 soil samples contained total chromium concentrations above the calculated background concentration range in soils (52 mg/kg). Five soil sampling locations west of the former Cattaraugus Creek dam (in the wetland area) contained chromium at concentrations greater than 1000 mg/kg. Hexavalent chromium concentrations in the surface soils exceeded detection limits in 33 of 65 samples analyzed. The maximum concentration of hexavalent chromium detected was 24 mg/kg. However, the number of positive detections of hexavalent chromium may be overstated. The analytical method used is subject to interference from chromium and an apparent positive detection of hexavalent chromium may result. Zinc analyses were performed on 20 surface soil samples. The average zinc concentration in surface soils was 401 mg/kg.

Nine surface soil samples were collected during the 1996 USEPA investigation. Sampling included seven soil samples collected from the southern bank of the Cattaraugus Creek (bank samples) and two samples from the wetland area (wetland samples) located west of the landfill. The bank samples were collected from the face of the landfill, adjacent to Cattaraugus Creek. Surface soil samples were analyzed for chromium, hexavalent chromium, zinc, and SVOCs. Summaries of the analytical results for these samples are provided in Appendix B-4.

Arsenic concentrations in soil samples collected along Cattaraugus Creek range, from 4.9 mg/kg (Bank 100) to 15 mg/kg (Bank 200). Total chromium concentrations ranged from 27 mg/kg (Bank 100) to 750 mg/kg (Bank 0). Maximum arsenic and chromium concentrations in the wetland samples were 6.7 mg/kg and 27 mg/kg, respectively. Hexavalent chromium was not detected in any of the soil or sediment samples (Weston, 1996).

Low concentrations of several SVOCs, including low levels of PAHs and one phthalate, were detected. PAH compounds, likely associated with the historic storage of coal on the bed of the railroad spur, were detected at concentrations ranging from 52 to 950 ug/kg.

1.2.3.3 Chemical Constituents in Subsurface Soil

During the USEPA investigation, a total of five subsurface soil samples were collected from depths of 0 and 2 feet bgs in the southeastern area of the property. Refer to Appendix B-4 for a



summary of the analytical results. Arsenic concentrations ranged from 5.3 mg/kg (HA SB-75) to 19 mg/kg (HA SB-76) while total chromium concentrations ranged from 23 mg/kg (HA SB-75) to 730 mg/kg (HA SB-76). Hexavalent chromium was not detected in any of the samples. Analysis of the subsurface soil samples collected during the 1989 RI reported concentrations within the ranges of constituents reported by USEPA Region II.

Three of the five subsurface samples (HA SB-74, HA SB-75 and HA SB-76) contained relatively low concentrations of PAH compounds. These three sample locations were those closest to the waste pile access road and the abandoned rail line.

1.2.3.4 Chemical Constituents in Surface Water and Sediment

Six surface water and five sediment samples were collected from Cattaraugus Creek during the 1996 USEPA investigation. Samples were collected adjacent to and downstream of the landfill. Sample locations and analytical results are shown in Appendix B-4. Samples were analyzed for Target Analyte List (TAL) inorganics, hexavalent chromium, and SVOCs. Analytical results for surface water samples ranged from 0.0023 mg/l to 0.0046 mg/l for arsenic and from 0.006 to 0.009 mg/l for total chromium. Analytical results for sediment samples ranged from 4.8 to 6.3 mg/kg for arsenic and from not detected (10 mg/kg) to 12 mg/kg for total chromium. Hexavalent chromium was not detected in any of the surface water or sediment samples. Low concentrations of PAHs were detected in the sediment samples.

Four rounds of surface water samples were collected from three locations in Cattaraugus Creek during the 1989 RI. The samples were analyzed for arsenic, chromium, hexavalent chromium, zinc, calcium, and magnesium. Sampling results are summarized in Appendix B-3. Arsenic was not detected in any of the samples. Total chromium concentrations ranged from 0.008 to 0.019 mg/l in the downstream samples. Hexavalent chromium was detected inconsistently in both upstream and downstream surface water samples. The highest concentration of hexavalent chromium (0.016 mg/l) was detected in downstream sample 13 in September 1986. One sediment sample (sample 14) was collected from a sand bar in Cattaraugus Creek adjacent to the wetland northwest of MW-4. No other sediment samples were collected because the upstream sampling locations were devoid of sediment due to the bedrock creek bed and the high stream velocity in the vicinity of the site. Both arsenic and chromium were detected at concentrations within the range of background of 6.6 mg/kg. Hexavalent chromium was not detected in the sediment sample. Sample locations 13 and 14 are shown in Appendix B-4.



1.2.3.5 Chemical Constituents in Groundwater

Groundwater samples collected from on-site monitoring wells during the 1989 RI and 1996 USEPA investigations indicate that chemical constituents associated with waste materials at the Site also are present in groundwater in the overburden and bedrock. Results of groundwater sampling for the 1991 RI and the 1996 USEPA are provided in Appendix B-3 and B-4, respectively.

Three rounds of groundwater samples were collected during the 1989 RI: September 1986, April 1987, and, July 1988. Groundwater sample results for total metals and filterable metals are summarized in Appendix B-3. Monitoring wells that were sampled included MW- 1 SR, MW- 1 D, MW-2D, MW-4D, MW-5, and, MW-6. Monitoring wells MW-2S, MW-3, and MW-4S (wells along the creek) were not sampled (see map in Appendix B-3 for monitoring well locations).

1989 RI

During the first round of sampling, ground water samples for metals analysis were filtered in the field. As requested by NYSDEC, samples collected for the second round of sampling were not filtered, however, selected samples were filtered to assist in comparing data with the first round results (O'Brien & Gere, 1989). Groundwater samples were analyzed for arsenic, total chromium, and hexavalent chromium. Samples from monitoring wells from MW- 1S, MW-1D, MW-4D, MW-5 and MW-6 were analyzed for priority pollutants including PCB/Pesticides, purgeable organic compounds, and SVOCs.

Results of groundwater monitoring conducted during the 1989 RI indicate the presence of chromium, hexavalent chromium, and arsenic in both the shallow and bedrock wells and in both filtered and unfiltered samples. In the April 1987 sampling round (unfiltered samples), arsenic, total chromium, and hexavalent chromium were detected in shallow (MW-1S) and bedrock monitoring wells (MW-1D) located upgradient and southwest of the landfill. One or more inorganic substances including chromium, hexavalent chromium, and arsenic were detected in samples from bedrock monitoring wells MW-2D and MW-4D and in overburden monitoring well MW-6. In the filtered samples collected in September 1986, arsenic, chromium, and hexavalent chromium were detected in monitoring well samples MW-1D and MW-2D; chromium was detected in monitoring well MW-6. Other inorganic compounds detected at elevated concentrations in the overburden where waste/fill is present include: ammonia, Total Kjeldahl Nitrogen (TKN), BOD₅, and specific conductance.



No organic compounds were detected in MW-5 during the first round of sampling. In the second sampling round, six organic compounds including acetone, 2-butanone, bis (2ethylhexyl) phthalate, chlorobenzene, methylphenol, and 2,4,-methylphenol were detected in the groundwater sample from monitoring well MW-5. Round three consisted of sampling of four monitoring wells: MW-1S,MW-1D,MW-4D and MW-6. Monochlorobenzene was detected in one well (MW-6) at a concentration of 26 ug/l.

1996 USEPA Investigation

Seven downgradient monitoring wells were purged and sampled during the USEPA investigation in September 1996. Groundwater sampling results for organic and inorganic parameters are provided in Appendix B-4. Monitoring wells that were sampled included MW-02S, MW-02D, MW-03, MW-04S, MW-4D, MW-05, and MW-06.

Inorganic compounds including: chromium; hexavalent chromium; and arsenic were also detected in groundwater samples collected in the 1996 USEPA investigation. Arsenic concentrations ranged from not detected (MW-05) to 0.1 mg/l (MW-02S). Chromium concentrations ranged from 0.004 mg/l (MW04D) to 1.1 mg/l (MW-03). Hexavalent chromium was detected in all of the groundwater samples; concentrations ranged from 0.008 mg/l to 0.06 mg/l. SVOCs, including phenols and trace concentrations of PAHs, were also detected in groundwater samples collected from monitoring wells MW-2S, MW-3, and MW-6 during the 1996 USEPA investigation. Phenol (8,000 ug/l), 2-methylphenol (69 ug/l), and 4-methylphenol (42,000 ug/l) were detected in monitoring well MW-2S. Phenol (99 ug/l) and 4-methylphenol (1,200 ug/l) were detected in the groundwater sample from MW-3.

1.2.3.6 Chemical Constituents in Air

No data were collected concerning the quantity and characteristics of gas generated by the landfill.

1.3 CURRENT CONDITIONS

Benchmark Environmental Engineering & Science, PLLC (Benchmark) and Geomatrix Consultants, Inc. (Geomatrix) performed Remedial Investigation field activities at the Site on several occasions beginning in August 2000 and continuing through April 2001. A brief description of site conditions as observed during these investigations is provided below. Section 3.0 provides additional detail concerning site conditions.



1.3.1 Inactive Landfill Area

Cover on the surface of the landfill is generally well vegetated and is preventing direct contact with waste materials, but appears to be thin in a few places where stressed vegetation is visible. Odors have been detected where stressed vegetation is visible within the elevated fill area and near certain seeps which are present along the northeast side of the Inactive Landfill Area.

The remains of a concrete and cemented boulder dam are present on the western edge of the elevated fill area that separates the fill area from the adjacent wetland area. The dam was reportedly part of a hydroelectric generating station. The dam is constructed of a large concrete monolith that at one time extended into the Creek, and cemented boulders that extend toward Palmer Street. The top of the dam sits approximately 8-feet above the adjacent wetland area. Riprap revetment is present on the creek bank in the northwest corner of the site. The revetment runs from the water to the top of the bank, and extends approximately 150 feet east from the former dam.

As indicated in Plate 1, a sluiceway was present on a portion of the northern border of the Area. According to the O'Brien & Gere Remedial Investigation Report, the sluiceway served as a Creek water source for the plant, possibly for fire protection system and/or process water feed. The sluiceway no longer exists as an open channel. It appears that fill has been placed up to the outer wall of the sluiceway.

1.3.2 Former Manufacturing Plant Area

In general, all former buildings and support structures have been demolished to grade. Wood, masonry demolition debris, the remnants of former foundations and various other salvage materials reportedly brought to the Site by the current property owner, JimCar Development, substantially cover the eastern side of the area. The debris exists in mounds several feet high, and limits site access on the southeastern side of the Area. Scrub vegetation and deciduous trees of various sizes are present outside and around the debris piles and foundation slabs, and along the northwestern side of the Former Manufacturing Plant Area. No visible evidence of production waste or vegetative stress was encountered.

1.4 PURPOSE AND SCOPE

This Remedial Investigation (RI) Report has been prepared on behalf of the responding PRPs (the Respondents) to present a characterization of the nature and extent of chemical impacts in both the Inactive Landfill Area and the Former Manufacturing Plant Area of the Site.



This Report contains seven sections.

- Section 2.0 presents a discussion of the RI sampling and methodology.
- Section 3.0 presents a discussion of land use and physical conditions of the Site.
- Section 4.0 presents the nature and extent of chemical presence in Site media.
- Section 5.0 describes chemical constituent migration pathways.
- Section 6.0 presents a summary of the Baseline Risk Assessment.
- Section 7.0 presents cited references.



2.0 SAMPLING LOCATIONS AND RATIONALE

This section presents a discussion of the rationale for the data collection program of the RI. The rationale was used to select the locations and depths at which to sample a given environmental media. This section also presents the methodologies used to collect samples and make physical measurements and observations, and the methodologies used to chemically analyze the environmental samples. The Remedial Investigation scope of work was documented in the USEPA-approved RI/FS Work Plan dated October 1999, revised March 2000 and Addendum dated May 2000.

The information in this section is sorted by type of environmental media (e.g., soil, water). This organization is paralleled in the presentation of results. In some cases, the rationale or methodology was dependent on the location at the site, that is, whether the data were being collected at the Inactive Landfill Area or at the FMP Area. In such cases, the information pertaining to the Inactive Landfill Area is presented first and the information pertaining to the FMP Area is presented second.

In addition to sample collection of Site media, some data collected during the RI required the installation of monitoring wells and survey benchmarks from land surveying. Data collection methods are made available to the reader in this section.

RI field activities were conducted by Geomatrix Consultants, Inc. (Geomatrix) and Benchmark Environmental Engineering (Benchmark) in accordance with the Site Health and Safety Plan (HASP) for Remedial Investigation Activities, Peter Cooper Site, Gowanda, New York, (Benchmark, May 2000). Environmental sample collection was performed in accordance with the Field Operating Procedures (FOPs) provided in the Quality Assurance Project Plan for Remedial Investigation/Feasibility Study, Peter Cooper Site, Gowanda, New York (QAPP) prepared in May 2000, revised in August 2000, by Geomatrix and Benchmark. All field activities were conducted under oversite from a USEPA contractor, TAMS Consultants, Inc. (TAMS) (now known as EarthTech). Each sampling location was surveyed by TVGA Engineering & Surveying (TVGA) and plotted on the site base map shown on Plate 1.

2.1 SLUDGE FILL CHARACTERIZATION

The sludge fill was characterized through the following activities:



- Geophysical (electromagnetic) methods were conducted to delineate the area occupied by fill and assess the vertical extent of groundwater impacted by constituents in the sludge fill;
- Test pits and hand holes were excavated to make physical observations and measurements of the extent of the sludge fill; and,
- Soil borings were completed to evaluate physical and chemical properties of the sludge fill.

The following sections describe the methodologies that were employed.

2.1.1 Collection of Geophysical Data

Two geophysical surveys were conducted at the Inactive Landfill Area. These surveys included a downhole electromagnetic (EM) survey and a surface EM survey. The downhole EM survey was conducted to profile the vertical terrain conductivity to assist with the characterization of the vertical extent of groundwater impacts from sludge fill disposal. The surface EM survey was conducted to geophysically characterize the lateral location of conductive fill material. These surface EM data were subsequently used to focus intrusive test pit and soil boring activities.

The geophysical methods used during this survey are established, indirect techniques for nondestructive subsurface reconnaissance exploration. As these instruments utilize indirect methods, they are subject to inherent limitations and ambiguities. Metallic surface features (electrical wires, scrap metal, etc.) preclude reliable non-invasive data/results beneath, and in the immediate vicinity of, the surface features. Targets such as conductive plumes, buried wastes, etc. are detectable only if they produce recognizable anomalies or patterns against the background geophysical data collected.

2.1.1.1 Geophysical Survey Methodology

Downhole Logging: The deep bedrock monitoring well, MW-4D2, at the Inactive Landfill area was geophysically logged using induction logging methods on October 5, 2000. Landfill leachate usually promotes an increase in the total dissolved solids (TDS) concentration of ground water. The elevated values of TDS create higher than background values of electrical conductance. EM conductivity logging was performed for the purpose of mapping formation conductivity. Electromagnetic surveys map the distribution of conductivity in the subsurface. A detailed discussion of the downhole logging is presented in Appendix C.



Surface Electromagnetic Survey: The Geonics EM-31 device was used to map the apparent electrical conductivity of shallow soils from the ground surface to a depth of 15 feet across the Inactive Landfill Area during the week of September 15, 2000. A reference grid was installed over the area that was geophysically surveyed by TVGA. The grid consisted of alternating orange and yellow pin flags spaced to facilitate data acquisition along lines spaced 12.5 feet apart. Select grid coordinates were marked to assure that grid coordinates could be reoccupied if necessary. Surface features were annotated on-site to assist with geophysical data interpretation.

The terrain conductivity (quadrature) component of the EM field is a measurement of the apparent ground conductivity. All readings were taken with the instrument oriented parallel to the direction of travel, in the vertical dipole mode and with the instrument at waist height. The depth of investigation with the instrument in this configuration is approximately 15 feet. Readings were automatically stored in a solid state memory data logger during the survey. The data logger was interfaced to a portable computer and the data were electronically transferred for subsequent processing and interpretation. A detailed discussion of the geophysical survey is presented in Appendix C.

2.1.2 Physical Observations and Measurements

2.1.2.1 Waste Fill Delineation

In accordance with the RI/FS Work Plan, test pits were excavated around the apparent perimeter of the elevated fill area to establish the limits of buried waste fill material. This work was performed on October 5 through 12, 2000 using a rubber-tired backhoe. Test pit excavations were initially targeted at locations inside the apparent edge of the elevated fill mound, and were extended or relocated radially outward until the transition between sludgebearing fill and surrounding soil/fill was evident. A total of five test pits (TP-A through TP-E) were excavated to locate the limits of the waste fill.

Upon establishing a transition point between waste fill and surrounding soil/fill material, the transition point was staked for tie in to the site survey. Test pits at transition zones were generally excavated to a depth up to 12-feet below ground surface to verify that deeper zones of waste fill were not present. Test pit observations and field measurements are discussed in Section 3.5.1. Test pit locations are shown on Figure 2-1. All test pits were backfilled with excavated materials in the opposite order from which they were removed to assure cover soil replacement.



In addition to the perimeter test pits described above, five test pits labeled TP-1/G through TP-5/G (Figure 2-1) were excavated adjacent to the riprap revetment on the Creek bank and the former hydroelectric dam. The purpose of test pit excavation was to characterize sludge fill materials and determine the thickness of the sludge fill in the northwest corner of the landfill. Although additional test pits were planned along the northeastern portion of the elevated fill area, access was limited due to the steep slope of the bank and tree cover.

Test pits TP-1/G & TP-2/G were excavated to the top of rock. Test Pits TP-3/G through TP-5/G were excavated until water was encountered. Rapid infiltration of groundwater prevented further test pit advancement. Excavated material was placed on plastic sheeting adjacent to the test pit. Soils and fill were logged by a field geologist and screened for organic vapors with a photoionization detector (PID). Findings are presented in Section 3.5.1.

2.1.2.2 Existing Cover Evaluation

A total of 24 test holes were excavated on October 10 and 11, 2000 across the elevated fill portion of the Inactive Landfill Area (Figure 2-1). Test holes were excavated to determine existing landfill cover system thickness and characteristics. Although the RI/FS Work Plan specified hand excavation of the test holes, cover soils were found to be thicker than originally anticipated. Accordingly, USEPA's on-site representative, TAMS Consultants, approved excavation via a rubber-tired backhoe. Each test hole was extended into waste material to allow measurement of approximate cover thickness and description of the soil horizons. A field geologist recorded a description of the soil types encountered on field test pit log forms (Appendix D) and photographed each test hole. Excavated soil material was placed adjacent to the hole and returned in the opposite order that it was removed. Test hole locations were surveyed upon completion of the program. Findings are discussed in Section 3.5.1.

2.1.3 Collection of Geotechnical Data

Geotechnical samples were collected from within the Inactive Landfill Area to characterize the engineering properties of cover soils and sludge fill material and assess the existing cover soil's effectiveness in minimizing infiltration of precipitation. Results will be used in the evaluation of remedial alternatives for the Site. Soil samples for geotechnical analysis were sent to Third Rock, LLC, in East Aurora, New York.

2.1.3.1 Cover Soil

At each test hole, a representative soil sample was collected. Four individual samples were combined to represent one composite sample such that each composite represents



approximately one-acre of cover soil. In addition, one undisturbed soil sample (i.e. Shelby tube) per acre was collected to evaluate in place properties. A total of six composite soil samples and six undisturbed soil samples were analyzed to represent the cover system.

Bulk composite samples were analyzed for grain size distribution (ASTM D421, D422), Atterberg Limits (D4318), Modified Proctor (ASTM D1557), and Recompacted Permeability (ASTM D5084). Shelby tube samples were analyzed for moisture content (ASTM D2216), hydraulic conductivity (ASTM D5084) and shear strength (ASTM D3080). Results are presented in Section 3.5.1.

2.1.3.2 Sludge Fill

Three Shelby tube soil samplers were advanced using direct push techniques in the elevated portion of the Inactive Landfill Area to allow for sludge fill geotechnical sample collection. At each location, additional sludge fill material was collected and combined into one composite sample representative of the sludge fill material. The composite sample was analyzed for Total Organic Carbon (Walkley Black Titration Method), Atterberg Limits (ASTM D4318), grainsize distribution (ASTM D421, 422), and Shear Strength (ASTM D3080). A discrete sample (ST-2) was also collected for vertical permeability (ASTM D5084) determined in the laboratory. Results are presented in Section 3.5.1.

2.2 Soil

Surface and subsurface soil samples were collected across the Inactive Landfill Area and the Former Manufacturing Plant Area to evaluate the extent of chemical impact in soil, if any, and support human health and ecological risk assessments. The following sections describe the sampling rationale and methodology.

2.2.1 Sampling Rationale

2.2.1.1 Inactive Landfill Area

Surface and subsurface soil samples were collected across the Inactive Landfill Area to evaluate the physical characteristics of the soil and fill including the presence of odors or staining of soil related to potential chemical impact and assess the nature, magnitude and extent of chemical concentrations in soil and fill. Surface soil sampling was conducted in a grid-like pattern to provide complete characterization of the approximately 8-acre portion of the nonelevated area of the Inactive Landfill Area. Subsurface soil samples were collected at the Inactive Landfill Area in areas of anomalously elevated conductivity (fill) established by the geophysical investigation. Subsurface sampling in these areas was used to establish fill type



and collect soil samples for chemical characterization. Also, the area of the former concrete sluiceway channel was investigated to ascertain the location of a settling basin shown on historical fire insurance maps and aerial photos and assess chemical presence in soil and sediment in the basin.

2.2.1.2 Former Manufacturing Plant Area

Surface soil and subsurface soil sampling at the Former Manufacturing Plant Area was conducted to assess potential chemical presence in soil/fill located near historic operational areas of the former glue factory (see map in Appendix A). These areas include:

- a former unloading house,
- downgradient of ponds formerly in the northwest portion of the Former Manufacturing Plant Area,
- the former Fertilizer Plant,
- downgradient of the former Machine Shop and storage area,
- the former Vat House,
- historic storage tanks adjacent to the former Cook House,
- the former Cook House,
- the former Acid Room,
- the former Dry House,
- the former Finished Product Warehouse,
- downgradient of storage buildings formerly adjacent to the Finished Product Warehouse, and
- general areas the southeastern property boundary.

2.2.2 Surface Soil Sampling Methodology

Surface soil samples from the Inactive Landfill Area and the Former Manufacturing Plant Area were collected using dedicated and disposable, stainless steel sampling equipment. Samples designated for VOC analysis were collected using EnCore® samplers. Samples to be analyzed for SVOCs and metals were placed in laboratory provided, certified clean, glass sampling jars. Each sample was given a unique nine-digit sample identification code and placed on ice until a



laboratory provided courier picked up the samples under chain-of-custody procedures, as described in the QAPP.

2.2.2.1 Inactive Landfill Area

Surface soil sampling at the Inactive Landfill Area was conducted on October 11 and 12, 2000. A total of 20 soil samples (LFSS-1 through LFSS-20) were collected from 0 to 6-inches below ground surface (immediately below the soil and vegetative layer) at locations shown on Figure 2-2. A shovel was used to remove the vegetative layer and expose the sampling interval.

2.2.2.2 Former Manufacturing Plant Area

Surface soil sampling at the Former Manufacturing Plant Area was conducted on October 5, 6, and 9, 2000. A total of 12 borings were advanced and 10 soil samples (SB-1, SB-2, SB-4, SB-5, SB-7, SB-8, SB-9, SB-10, MWFP-2, and MWFP-3) were collected from 0 to 2.5 feet below ground surface as shown on Figure 2-3. Direct push techniques were used in combination with 4-1/4-inch hollow stem augers (HSA) to advance soil borings and allow for continuous sampling using dedicated and disposable acetate sleeves. Borings were logged by a field hydrogeologist in accordance with the Unified Soil Classification System (USCS) and screened with a photoionization detector (PID) for VOCs. Boring logs are provided in Appendix E.

A surface soil sample was not collected from borings SB-3 and SB-6. Concrete foundation slabs were present. Subsurface soil samples were collected by coring through the slabs and advancing the boring below the concrete corehole.

2.2.3 Subsurface Soil Sampling Methodology

Subsurface soil samples from the Inactive Landfill Area and the Former Manufacturing Plant Area were collected using dedicated and disposable, stainless steel sampling equipment. Samples designated for VOC analysis were collected using EnCore® samplers. Samples for SVOCs and metals were collected in laboratory provided, certified clean, glass sampling jars. Each sample was given a unique nine-digit sample identification code and placed on ice until a laboratory provided courier picked up the samples under chain-of-custody procedures, as described in the QAPP.

2.2.3.1 Inactive Landfill Area

Test pit excavation within the Inactive Landfill Area (outside or adjacent to the elevated fill area) and subsurface soil sample collection activities were conducted October 6 through the week of October 9, 2000. A total of 9 soil samples (TP-1/SUB through TP-9/SUB) were



collected from test pit excavations from depths ranging from 3 to 12.5-feet below ground surface (bgs) at locations shown on Figure 2-4. A rubber tired backhoe was used to excavate test pits to the water table with the exception of TP-2/SUB which was excavated to native material since groundwater was not encountered. The test pits allowed for continuous observation of soil horizons and features including zones of groundwater seepage. Soil samples were collected from the sidewalls of test pit excavations immediately above the apparent saturated zone.

Three test pits were excavated in the area of the former sluiceway and settling basin and two subsurface soil samples were collected. The test pit excavated in the former sluiceway (TP-10/SUB) was sampled immediately above the concrete sluiceway bottom at a depth of 1-foot bgs. The soil sample collected from a test pit associated with the former settling basin (TP-Settling Basin/SUB) was collected just above the concrete basin pad as was TP-5/SUB which is assumed to be the eastern limit of the former settling basin.

At the completion of each test pit, the excavated material was replaced in the opposite order that it was removed.

2.2.3.2 Former Manufacturing Plant Area

Subsurface soil sampling at the Former Manufacturing Plant Area was conducted on October 5, 6, and 9, 2000. Soil borings were advanced using direct push techniques at surface soil sampling locations (see Figure 2-3) to facilitate subsurface soil sample collection. A total of 12 subsurface soil samples (SB-1, SB-2, SB-3, SB-4, SB-5, SB-6, SB-7, SB-8, SB-9, SB-10, MWFP-2, and MWFP-3) were collected from depths ranging from 3 to 12-feet bgs. Soil samples exhibiting the highest degree of suspected chemical impact were selected for laboratory analysis. If evidence of impact was not observed, the soil sample was collected immediately above the apparent saturated zone.

2.2.4 Geotechnical Data

A total of 11 surface and 12 subsurface soil samples from the Former Manufacturing Plant Area collected on October 5, 6, and 9, 2000 were evaluated for grain size distribution analysis (ASTM D421). The geotechnical data were used to evaluate potential soil mobility and assess the engineering properties for use in the Baseline Risk Assessment and the engineering Feasibility Study.



2.2.5 Methods of Chemical Analysis

Surface and subsurface soil samples from the Inactive Plant Area and the Former Manufacturing Plant Area were couriered to Columbia Analytical Services (CAS), in Rochester, New York for chemical analysis. The laboratory employed analytical testing methods described in USEPA Test Methods for Evaluating Solid Wastes contained in SW-846, revised 1991.

Based on historic analytical data obtained from previous investigations (Appendix B) and agreed to by the USEPA, soil samples collected from the Inactive Landfill Area were analyzed for a list of target analytes referred to as chemicals of potential concern (COPCs). The COPCs included volatile aromatic hydrocarbon compounds and select metals (arsenic, chromium, hexavalent chromium, and zinc). The target VOCs were verified during a preliminary sampling event of groundwater collected and analyzed from the Inactive Landfill Area described in Section 2.4.1.7. The target VOC analyte list includes:

- benzene,
- chlorobenzene,
- ethylbenzene,
- 1,2-dichlorobenzene,
- 1,4-dichlorobenzene,
- toluene, and
- total xylenes.

VOCs were analyzed by EPA Method 8260B. Arsenic, total chromium, and zinc were analyzed by EPA Method 6010B and hexavalent chromium analyzed by EPA Method 7196. The hexavalent chromium was prepared per EPA Method 3060A to suppress oxidation of soluble trivalent chromium to hexavalent chromium. The analysis also included Total Organic Carbon (TOC), percent solids, and pH.

Soil samples collected from the Former Manufacturing Plant Area were analyzed for a full suite of compounds including: TCL VOCs (EPA Method 8260B), TCL SVOCs (EPA Method 8270C) and Target Analyte List (TAL) Metals (EPA Method 6010B, 7470) including hexavalent chromium (EPA Method 3060A/7196), TOC, percent solids, and pH.



2.3 LANDFILL GAS

Landfill gas samples were collected within the elevated fill area of the Inactive Landfill Area to evaluate VOC presence in landfill gas. Landfill gas analytical results were used to evaluate potential air-borne pathways from volatilization of chemicals present in the sludge fill.

2.3.1 Gas Sampling Methodology

Landfill gas sampling was facilitated by installing gas monitoring wells. Three borings were advanced to the top of bedrock within the elevated fill area using 4-1/4-inch hollow stem augers (HSA) and continuous split spoon sampling. At boring completion, landfill gas monitoring wells (GMW-1, GMW-2, and GMW-3) were constructed with 2-inch diameter, Schedule 40 PVC riser pipe. Well locations are shown on Figure 2-5A. The gas probes were set to screen the sludge fill material with slotted well screens ranging from 3.5 to 7.5-feet in length. Each gas monitoring well was completed with #00N sand pack and a bentonite chip seal. Boring logs are included in Appendix E.

The headspace of each gas monitoring well was field screened for landfill gases on October 12, 2000 and May 2, 2001 using hand-held field instruments. The well headspace was screened for methane, hydrogen sulfide gas, percent oxygen, carbon monoxide, and VOCs with a PID. Water was present in GMW-2 and GMW-3 and was purged using a disposable bailer prior to instrument screening.

During the October 12, 2000 sampling event, 6-liter summa canisters were used to collect landfill gas samples from each well. The initial pressure in each summa canister was confirmed at 30 millimeters of mercury (mm Hg) prior to use. Dedicated Teflon®, tubing was connected to each canister and inserted into the well headspace several feet below the top of the PVC riser. The canister vacuum was released and the canister was filled until the regulator registered a vacuum pressure of 2 to 5 mm Hg.

Each canister was given a unique nine-digit sample identification code and a laboratory provided courier picked up the samples under chain-of-custody procedures, as described in the QAPP. Analytical results are discussed in Section 4.3.

2.3.2 Methods of Chemical Analysis

The summa canisters were analyzed by Performance Analytical, a specialty analytical laboratory specializing in air analysis (subcontracted through CAS) using EPA Method TO-



14A for VOCs and landfill gases, oxygen, nitrogen, methane, and carbon dioxide (Modified EPA Method TO-3).

2.4 GROUNDWATER

A groundwater monitoring program was conducted at the Peter Cooper Site to assess chemical presence in groundwater to support human health and ecological risk assessments. Additionally, the groundwater monitoring program provides an assessment of potential groundwater migration pathways, supports the hydrogeologic and fate and transport conceptual models for the Site, and considers the effects of seasonal variation on groundwater quality and flow. The following sections describe the sampling rationale and methodology. Monitoring well locations discussed in this section are shown in Figure 2-5A and 2-5B for the Inactive Landfill Area and Former Manufacturing Plant Area, respectively. Monitoring well logs for all wells at the Peter Cooper Gowanda Site are included in Appendix E.

2.4.1 Initial Groundwater Monitoring Well Evaluation

Previous investigations of the Inactive Landfill Area (1989 RI) installed 10 groundwater monitoring wells (MW-1, MW-1SR, MW-1D, MW-2S, MW-2D, MW-3, MW-4S, MW-4D, MW-5, and MW-6). Monitoring well MW-1, originally installed by RECRA Research in 1995, was found unusable by OBG and MW-1SR was installed as a replacement well. The original well MW-¹ was not abandoned by OBG. On June 15 and 16, 2000, Geomatrix conducted a groundwater monitoring well integrity program of the existing wells at the Inactive Landfill Area to determine the usability of each well for the Remedial Investigation. All wells were located to ascertain well integrity by evaluating the condition of the protective casing, access to the well, and potential quality of groundwater samples that would be collected from the well. An attempt was made to re-develop each well for future sampling for the RI, however, wells MW-2, MW-3, MW-4S, and MW-4D were determined unusable for groundwater sample collection as a result of obstructions in the well or failure to produce adequate water.

2.4.1.1 Replacement of Groundwater Monitoring Wells

Four replacement wells (MW-2S(R), MW-3(R), MW-4S(R), and MW-4D(R)) were installed by Nothnagle Drilling Company (Nothnagle) of Rochester, New York the week of July 10, 2000 in accordance with the FOPs provided in the QAPP. The replacement wells were drilled to a similar total depth as the original groundwater monitoring well using a CME-55 trackmounted drill rig equipped with 4 ¹/₄-inch hollow stem augers (HSAs) which were decontaminated between locations in accordance with the QAPP. Continuous split spoon samples were obtained and logged by a Geomatrix hydrogeologist. The bedrock well, MW-



4D(R), was advanced into bedrock using an HQ core barrel. Drill cuttings were containerized on-site in drums and rock cores were retained in core boxes provided by the drilling contractor.

At the completion of the borings, a 2-inch diameter schedule 40-PVC monitoring well was installed in each borehole. The well screen (0.01-inch slot size) interval was placed to span a similar screen interval as the original monitoring well. A filter pack sand (#00N) material extended to the top of the well screen after which 1-foot of transition sand (#00) was placed to prevent potential seepage of bentonite or grout into the sand pack. Bentonite chips were poured to form the filter pack seal and the remainder of the borehole annulus to 3-feet below grade was filled with a cement/bentonite grout using a tremie pipe. The top of the riser pipe extended approximately 3-feet above grade and installed with a lockable, protective 4-inch diameter steel casing anchored in concrete to allow surface water to drain away form the well.

The drilling augers were decontaminated between well replacement locations using a steam cleaner. Water used for equipment decontamination was containerized in 55-gallon drums, marked with the date and contents.

Boring and monitoring well installation logs are provided in Appendix E. Well construction details are summarized in Table 2-1.

2.4.1.2 Abandonment of Groundwater Monitoring Wells

The five unusable monitoring wells (MW-1S {original well installed by RECRA}, MW-2S, MW-3, MW-4S, and MW-4D) were abandoned after well replacement.

The protective casing of non-functional monitoring wells MW-1S, MW-2S, MW-3S, and MW-4S were removed and the risers and well screens were pulled and the borehole was over drilled. The protective casing and well materials at MW-4D could not be removed until the well was reamed. Each remaining borehole was backfilled with cement/bentonite grout using a tremie pipe. The wire-wrapped well screens of the abandoned wells were highly corroded with ½-inch to 3-inch diameter openings. The observed corrosion is typical of galvanized iron well screens that do not have cathodic protection and have been in the ground for more than 15 years.

The hollow stem augers were decontaminated between well abandonment locations using a steam cleaner. Water used for decontamination was containerized in 55-gallon drums marked with the date and contents.

1: Project W05771 PRP Group Peter Cooper NPL/RI report/FINAL REPORT (November 2003 Submittal)/Text (Final)/Remedial Investigation Text - Final.doc 31



2.4.1.3 Groundwater Monitoring Well Development

The newly installed groundwater monitoring wells (MW-2S(R), MW-3(R), MW-4S(R), and MW-4D(R)) were developed on July 14, 2000 to reduce the turbidity of the water and improve the hydraulic communication between the well bore and the water-bearing zone.

A stainless steel bailer was used to surge the entire length of the well screen and remove water from the well. After each well volume of water removed, field parameters (temperature, pH, specific conductivity, and turbidity) were measured using hand held field instruments. Development was considered complete when a minimum of ten well volumes was purged and differences between measurements was less than 10% difference for three or more well volumes.

2.4.1.4 Initial Groundwater Monitoring Well Sampling

On August 14, 2000, four groundwater monitoring wells (MW-2S(R), MW-3(R), MW-4S(R), and MW-6) screened in or immediately below waste were sampled and analyzed to refine the list of Chemical Constituents of Potential Concern (COPCs) defined in the Work Plan for the Inactive Landfill Area.

Prior to sampling, the water level was recorded in each well and purged with dedicated tubing and a peristaltic pump to remove a minimum of three well volumes or until field measured parameters (temperature, pH, specific conductivity, oxidation-reduction potential, turbidity, and dissolved oxygen) stabilized. Field measured parameters were measured using hand held portable instruments which were calibrated in accordance with the FOPs of the QAPP on the day of sampling.

After stabilization was achieved, a dedicated and disposable bailer was used to collect the groundwater sample. Samples were collected in the order of volatilization sensitivity. Groundwater samples were collected in pre-preserved, laboratory provided, certified clean sample containers and labeled with a unique nine-digit code in accordance with the FOPs of the QAPP and placed on ice.

Samples were submitted under chain of custody procedures to CAS for target compound list volatile organic compounds (TCL VOCs) by EPA Method 8260B, target compound list semi-volatile organic compounds (TCL SVOCs) by EPA Method 8270C, and total target analyte list (TAL) inorganics by EPA Method 6010B/7470 plus hexavalent chromium by EPA Method 7196.

1: Project V005771 PRP Group Peter Cooper NPL/RI report/FINAL REPORT (November 2003 Submittal) \Text (Final) \Remedial Investigation Text - Final.doc 32



To satisfy quality assurance and quality control requirements, a duplicate sample was collected from groundwater monitoring well MW-3(R) and a matrix spike/matrix spike duplicate (MS/MSD) was requested for the sample collected from groundwater monitoring well MW-4S(R).

2.4.1.5 Initial Groundwater Sampling and Selection of COPCs - Inactive Landfill Area

Analytical data packages were validated by a third party certified data validator (Data Validation Services). The data validation determined that the data were usable with minor qualifications and satisfied the data quality objectives. The analytical results and data validation report is included in Appendix F. As a result of the initial groundwater sampling, aromatic hydrocarbon VOCs were added to the list of analytes defined as COPCs for the Inactive Landfill Area in a September 28, 2000 letter to the USEPA (Appendix G). The selected aromatic VOC list includes:

- benzene,
- chlorobenzene,
- ethylbenzene,
- 1,2-dichlorobenzene,
- 1,4-dichlorobenzene,
- toluene, and
- total xylenes.

2.4.2 Groundwater Monitoring Wells and Piezometer Installation Rationale

2.4.2.1 Inactive Landfill Area

Upgradient water quality data were lacking in the southwestern and southeastern portions of the Inactive Landfill Area. These data in additional to hydraulic data were needed to support the Baseline Risk Assessment and assess and support an assessment of chemical constituent migration pathways. Therefore, two additional upgradient groundwater monitoring well pairs were installed in the southeast (MW-7S/MW-7D) and southwest (MW-8S/MW-8D) portion of the Inactive Landfill Area.

Bedrock water quality data was lacking in the western portion of the Inactive Landfill Area. A bedrock well (MW-5D) was installed adjacent to the existing monitoring well MW-5.

1: Project 1005771 PRP Group Peter Cooper NPL/RI report/FINAL REPORT (November 2003 Submittal)/Text (Final)/Remedial Investigation Text - Final.doc 33



A deep bedrock groundwater monitoring well (MW-4D2) was installed adjacent to the MW-4S(R)/MW-4D(R) well pair to establish the vertical hydraulic gradient in the bedrock, to characterize the depth of the active groundwater flow, and assess potential chemical constituents from the Inactive Landfill Area in deeper bedrock.

A shallow piezometer was installed adjacent to the former hydroelectric dam (PZ-1) and a well point was driven into the wetland soil on the opposite side of the dam (DP-1). Water level data from the piezometer and drive point were used to evaluate the dam as a barrier to groundwater flow from the Inactive Landfill Area to the wetland.

2.4.2.2 Former Manufacturing Plant Area

Overburden and bedrock groundwater quality and hydrogeologic information were not available for the Former Manufacturing Plant Area prior to this remedial investigation. A total of three well pairs were proposed in the Former Manufacturing Plant Area: an upgradient well pair (MWFP-1D, an overburden well was not installed since the overburden was unsaturated) and two well pairs downgradient of process areas of the former plant (MWFP-2S/2D and MWFP-3S/3D) were installed to provide water quality data for the overburden and bedrock and support chemical fate and transport assessment of detected compounds.

The monitoring well pair MWFP-2S and -2D is downgradient of a former Finished Product Warehouse and Storage Areas. The locations of monitoring wells MWFP-3S and -3D are downgradient of former the Vat House and Machine Shop.

2.4.3 Groundwater Monitoring Wells and Piezometer Installation Methodology

2.4.3.1 Inactive Landfill Area

The overburden groundwater monitoring wells (MW-7S and MW-8S) were installed to a depth approximately 5 to 10 feet below the water table. The bedrock groundwater monitoring wells (MW-5D, MW-7D, and MW-8D) were installed in the upper bedrock zone and MW-4D2 was installed approximately 15-feet below the water level of Cattaraugus Creek.

The wells were installed in accordance with the FOPs in the QAPP and as described in Section 2.4.1.2. Well construction details are summarized in Table 2-1.

The piezometer (PZ-1) was installed using the same methods of overburden well installation to a depth of approximately 5 to 10 feet below the water table. The drive point (DP-1) was installed using a "drive point" piezometer consisting of a 2-foot well screen attached to a


carbon steel riser to a depth of approximately five feet below the ground surface of the wetland. The base of the screen was fitted with a flush-threaded drive point for installation. The drive point was advanced using a pounding block and a sledgehammer to a depth of approximately 5feet bgs. The drive point was fitted with a vented, locking J-plug.

2.4.3.2 Former Manufacturing Plant Area

Groundwater monitoring completion depths (lower overburden and shallow bedrock) in the Former Manufacturing Plant Area were consistent with those established for the Inactive Landfill Area. The overburden groundwater monitoring wells (MWFP-2S and MWFP-3S) and the bedrock groundwater monitoring wells (MWFP-1D, MWFP-2D, and MWFP-3D) of the Former Manufacturing Plant Area were installed in accordance with the FOPs in the QAPP and as described in Section 2.4.1.2. Well construction details are summarized in Table 2-1.

2.4.4 Newly Installed Groundwater Monitoring Well Development

Groundwater monitoring well development of the newly installed wells (MW-4D2, MW-5D, MW-7S, MW-7D, MW-8S, MW-8D, MWFP-1D, MWFP-2S, MWFP-2D, MWFP-3S, and MWFP-3D) was conducted using a stainless steel bailer and/or submersible pump in accordance with the FOPs in the QAPP and as described in Section 2.4.1.4.

Complete development, as described in the FOPs was achieved in all wells with the exception of MW-4D2, MW-7S, and MW-8D which exhibited extremely slow recharge rates during development. After purging these groundwater wells, recovery to static conditions required in excess of 24 hours. These wells were pumped and/or bailed dry many times (over several days) during the well development program. Due to the slow recharge of these wells, the water quality stabilization and volume requirements as stipulated in the QAPP (stabilization of water quality parameters, turbidity at 5 NTUs, and removal of 10 well volumes) were not achievable.

These variances from the Work Plan and FOP procedures did not negatively impact project objects and were discussed with the USEPA oversite contractor, TAMS Consultants.

2.4.5 Hydraulic Conductivity Estimates

The hydraulic conductivity of the overburden and bedrock was estimated for saturated site media. Hydraulic conductivity estimates were developed for the screened interval of each monitoring well using the "rising head" variable head slug test method. This was accomplished by removing a "slug" of known volume from the column of water in the well with a bailer and measuring the rate of water level recovery. Slug test data were analyzed using Aqtesolve®



aquifer test software by applying the unconfined and confined Bouwer and Rice (1976) methods for monitoring wells screened in the overburden and bedrock, respectively. Hydraulic conductivity data reduction for each slug test is presented in Appendix H. Hydraulic conductivity values estimated using these methods are presented and discussed in Section 3.5.

Hydraulic conductivity estimates were also obtained from the bedrock using packer test methods. Packer test methods were applied during bedrock drilling at monitoring well MW-4D2. The purpose of the packer testing was to evaluate hydraulic properties of distinct intervals within the bedrock downgradient from the inactive landfill area. Packer tests were conducted at four, 5-foot intervals:

- 18 to 23 feet below ground surface (bgs);
- 23 to 28 feet bgs;
- 28 to 33 feet bgs; and
- 33 to 38 feet bgs.

A double packer assemblage was used to isolate the test intervals in the HQ-size core hole. The pack test consisted of the injection of potable water at several different injection pressures for time periods ranging from 5 to 10 minutes. The flow rate from injection is calculated for each test interval. The water injection pressures, total quantity of water injected into the formation, and test specifications are included in the pressure test reports presented in Appendix H. The data from the packer injection tests were used to estimate the effective transmissivity of the isolated interval using the Thiem equation:

$$T = \frac{\mathcal{Q}\ln\left(\frac{R}{r_b}\right)}{2\pi P_i}$$

where:

- T = transmissivity (m^2/day) ;
- $Q = injection rate (m^3/day);$
- R = radius of influence (m);
- r_b = radius of borehole (m); and
- P_i = net injection pressure (m).

The net injection pressure is the combined pressure head based on the following:

$$P_{i} = P_{g+}h_{g+}h_{s} - h_{f}$$



where:

- P_i = net injection pressure (m);
- $P_g = gauge \text{ pressure (m)};$
- h_g = height of gauge above ground level (m);
- h_s = depth of pre-test water level (m); and
- h_f = friction losses (m).

Hydraulic conductivity is estimated by dividing the transmissivity by the length of the interval tested. Injection flow rates were measurable in two of the four tests performed. The total water injected into the 23 to 28 foot bgs and 33 to 38 foot bgs interval was less 0.1 gallons indicating very low transmissivity of these intervals. Nearly 7 gallons of water was injected into the 18 to 23 foot bgs interval during 19 minutes of injection and nearly 5 gallons was injected into the 28 to 33 foot bgs interval during 20 minutes of injection. Applying the Thiem equation to these test intervals, hydraulic conductivity estimates for each interval are:

| Test Interval (feet bgs) | Hydraulic Conductivity (cm/s) |
|--------------------------|--|
| 18 to 23 | 3.6 X10 ⁻⁵ |
| 23 to 28 | Not quantifiable; likely less than $1 \ge 10^{-6}$ |
| 28 to 33 | 1.6 X 10 ⁻⁵ |
| 33 to 38 | Not quantifiable; likely less than $1 \ge 10^{-6}$ |

Hydraulic conductivity data reduction for each packer test is presented in Appendix H. Hydraulic conductivity values estimated using these methods are discussed in Section 3.5.

2.4.6 Groundwater Elevation Measurements

Groundwater elevations were measured in all existing and newly installed wells/piezometers on a monthly basis beginning on August 14, 2000 or after installation. Groundwater elevation measurements were collected for the entire period between the first and second groundwater sampling events (November 2000 and April/May 2001). Groundwater elevations were



measured using an electric water level meter to the nearest 0.01 feet in accordance with the FOPs in the QAPP.

An upstream and downstream surface water monitoring station were installed in the Creek adjacent to the Site. The water level in Cattaraugus Creek was measured beginning in December, 2000 to assess the interaction between groundwater and surface water.

2.4.7 Groundwater Sampling

All existing and newly installed groundwater monitoring wells were sampled during two (2) full rounds of sampling. The first sampling event occurred during the high water table conditions (November 2000) and the second sampling event occurred during the low water table conditions (April/May 2001) to allow for seasonal variations in groundwater quality.

Upgradient groundwater monitoring wells were sampled first. Prior to sample collection, the water elevation was recorded in each well and the well was purged to ensure a representative groundwater sample. Purging was accomplished using low-flow purging/sampling techniques as described in the FOPs of the QAPP which references the USEPA Region II Low Flow Standard Operating Procedure.

The intake of the pump was placed approximately in the middle of the well screen. The pumping rate was measured and maintained between 100 and 500 milliliters (ml) per minute. During pumping, the water level measurements were recorded to avoid dropping the water' elevation more than 0.3-feet below the static water elevation.

During the first sampling event, several wells were determined not suitable for low-flow sampling techniques. As described in Section 2.4.4, MW-4D2, MW-7S, and MW-8D were extremely slow to recover. As such, each well was completely evacuated over a three day period and a sample was collected using a bailer.

During the second sampling event, wells MW-7S and MW-8D were able to maintain low-flow purging rates as described in the Work Plan and the FOPs in the QAPP. However, groundwater monitoring well MW-4D2 was purged and sampled with a dedicated bailer.

Monitoring wells MW-2S(R), MW-3(R), and MW-6 were sampled during both sampling rounds using a bailer after slow purging with a peristaltic pump based on the very low yield of the wells.



Hand held field instruments were calibrated daily in accordance with the FOPs of the QAPP to measure groundwater parameters. Field measured parameters (temperature, pH, specific conductivity, dissolved oxygen, oxidation-reduction potential and turbidity) were measured every three minutes until parameters were stabilized. Stabilization is achieved after three field parameter readings are within ± 0.1 unit for pH, $\pm 3\%$ for specific conductivity, ± 10 millivolts for oxidation-reduction potential, and $\pm 10\%$ for turbidity and dissolved oxygen.

Samples were collected from the dedicated tubing or dedicated/disposable bailers in the order of volatilization sensitivity. Where groundwater samples collected for total metals analysis had field measured turbidity values exceeding 50 NTU, a second sample was collected and filtered in the field using a 0.45 micron water filter for soluble metals analysis. Groundwater samples were collected in pre-preserved, laboratory provided, certified clean sample containers and labeled with a unique nine-digit sample identification code and placed on ice until a laboratory provided courier picked up the samples under chain-of-custody procedures, as described in the FOPs of the QAPP.

2.4.8 Methods of Chemical Analysis

2.4.8.1 First Sampling Event

Groundwater samples from the Inactive Landfill Area were analyzed for the following constituents, as described in the September 28, 2000 letter to the USEPA (Appendix G):

- Aromatic Hydrocarbon VOCs by EPA Method 8260B (defined in Section 2.4.1.6);
- SVOCs by EPA Method 8270C;
- Total COPC Metals by EPA Methods 6010B, 7470, and 7196 (arsenic, chromium, hexavalent chromium, and zinc);
- Dissolved COPC Metals by EPA Methods 6010B, 7470, and 7196 (arsenic, chromium, hexavalent chromium, and zinc) when field turbidity measurements were over 50 NTUs; and,
- Water Quality Parameters.

Groundwater samples from the Former Manufacturing Plant Area were analyzed for the following constituents, as presented in the Work Plan:

- TCL VOCs by EPA Method 8260B;
- TCL SVOCs by EPA Method 8270C;
- Total TAL Metals by EPA Methods 6010B, 7470, and 7196;



- Dissolved TAL Metals by EPA Methods 6010B, 7470, and 7196 when field turbidity measurements were over 50 NTUs; and,
- Water Quality Parameters.

2.4.8.2 Second Sampling Event

Groundwater samples from the Inactive Landfill Area were analyzed for the same constituents identified for the first sampling event.

Groundwater samples from the Former Manufacturing Plant Area were analyzed for the following constituents, as presented in the April 12, 2001 letter to the USEPA (Appendix I) based on the results of the pathways analysis assessment and reported in the Pathways Analysis Report (PAR). The Chemicals of Potential Concern (COPCs) identified in the PAR were selected as proposed Target Analytes for the second groundwater sampling event within the Former Manufacturing Plant Area. These constituents included:

- Select VOCs by EPA Method 8260B Benzene Carbon tetrachloride Chloroform Tetrachloroethene Trichloroethene;
- Select SVOCs by EPA Method 8270C Benzo(a)anthracene Benzo(a)pyrene Benzo(b)fluoranthene Indeno(1,2,3-cd)pyrene Benzo(a,h)anthracene; and
- Water Quality Parameters

2.5 SEEPS

Groundwater seeps discharging north of the Inactive Landfill Area were sampled to characterize a potential chemical constituent migration pathway from groundwater to surface water and support human health and ecological risk assessments.

2.5.1 Rationale for Sampling Locations

Based on an inspection of groundwater seepage along the northern border of the Inactive Landfill Area, three seep locations appeared to be conducive to sampling (adequate volume of



seepage to fill sample containers with visible precipitate). Sample locations are shown on Figure 2-6.

2.5.2 Seep Sampling

Seep samples (Seep L-1 through Seep L-3) were collected concurrent with the first and second groundwater sampling event (November 8, 2000 and May 2, 2001, respectively) during dry weather to avoid sample dilution. A soil pick was used to carve a channel in the weathered bedrock surface and direct seepage into a shallow depression at each location. Seep samples were collected from the shallow depressions using disposable, laboratory-provided, sample containers (i.e., unpreserved 40 ml vials). The seep water was transferred to appropriate pre-preserved sample containers. Field measured parameters included temperature, pH, oxidation-reduction potential, and specific electrical conductance.

Samples for laboratory analysis were each given a unique nine-digit sample identification code and placed on ice for a laboratory provided courier to pick up under chain-of-custody procedures, as described in the FOPs of the QAPP.

2.5.3 Methods of Chemical Analysis

Seep samples were analyzed for the identical list of parameters assigned to the Inactive Landfill Area groundwater.

2.6 SURFACE WATER

Surface water samples located in Cattaraugus Creek were sampled to characterize surface water chemistry to support human health and ecological risk assessments.

2.6.1 Rationale for Sampling Locations

A total of four surface water sample locations were selected in Cattaraugus Creek to be sampled concurrent with the first and second groundwater sampling event (Figure 2-6). One location (SW-1) was selected upstream of the Peter Cooper Site to represent the background water quality of the Creek. Sample location, SW-2, was upstream of the approximate division between the Inactive Landfill Area and the Former Manufacturing Plant Area. Sample location, SW-3, was selected immediately downstream from the sludge fill disposal area of the Inactive Landfill Area and SW-4 was selected at a location approximately 400-feet downstream of the Inactive Landfill Area. Sample locations are shown on Figure 2-6.



2.6.2 Surface Water Elevation

Surface water elevations of the Creek were collected beginning in December 2000, concurrent with the groundwater elevation measurements described above. Two surface water elevation locations were selected, SW-1 and SW-4, to assess the interaction between groundwater and surface water.

2.6.3 Surface Water Sampling

Surface water samples were collected concurrent with the first and second groundwater sampling event (November 8, 2000 and May 2, 2001, respectively). Surface water samples were collected from the downstream location first and progressively moving to upstream locations. The samples were collected by slowly dipping a non-preserved, laboratory provided, sample bottle into the creek with minimal disturbance. Field measured parameters included temperature, pH, oxidation-reduction potential, and specific electrical conductance.

Samples for laboratory analysis were each given a unique nine-digit sample identification code and placed on ice for a laboratory provided courier to pick up under chain-of-custody procedures, as described in the FOPs of the QAPP.

2.6.4 Methods of Chemical Analysis

Surface water samples for the first sampling event were analyzed for the identical list of parameters and methods as the Former Manufacturing Plant Area groundwater described , above.

Surface water samples for the second sampling event were analyzed for a constituent list inclusive of both the Inactive Landfill Area and Former Manufacturing Plant Area target analytes as described in the April 12, 2001 letter to the USEPA.

2.7 SEDIMENT

Sediment samples were collected from Cattaraugus Creek and the wetland adjacent to the Inactive Landfill Area. Sediment samples were collected to determine if chemical constituents from the Site affected sediment quality and to support human health and ecological risk assessments.

2.7.1 Rationale for Sampling Locations

Sediment samples collected from the Creek were collected at the same locations as the surface water samples for the same rationale. Sample locations are shown on Figure 2-6. Sediment



samples collected from the wetland adjacent to the Inactive Landfill Area were collected in a grid-like pattern to achieve a uniform characterization of sediment quality data. Sediment sample locations are shown with surface soil sample locations in the Inactive Landfill Area shown on Figure 2-2.

2.7.2 Sediment Sampling

Ten sediment samples (WSS-1 through WSS-10) were collected from the wetland on October 10, 2000 using surface soil sample methods employed across the Inactive Landfill Area. Creek sediments were collected concurrent with the first groundwater sampling event on November 7, 2000. Sediment samples were collected using dedicated and disposable stainless steel sampling equipment in accordance with the FOPs in the QAPP. Sediment samples were collected from sandbars downstream to upstream to avoid disturbing the quality of the sediment samples.

Samples for laboratory analysis were each given a unique nine-digit sample identification code and placed on ice for a laboratory provided courier to pick up under chain-of-custody procedures, as described in the FOPs of the QAPP.

2.7.3 Methods of Chemical Analysis

Surface water sediment samples were analyzed for TCL VOCs (EPA Method 8260B), TCL SVOCs (EPA Method 8270C), and Total TAL Metals (EPA Methods 6010B, 7470, and 7196), TOC (Walkley Black Titration Method) and pH.

Additional sediment samples were sent to the geotechnical laboratory for grain size distribution.

2.8 QUALITY ASSURANCE/QUALITY CONTROL MEASURES

Field investigation data were collected and processed using the procedures outlined in the QAPP and the Work Plan to ensure representative sample collection and to achieve the data quality objectives of the Remedial Investigation. The field activities were recorded in bound project field books consisting of field forms from the FOPs in the QAPP. Any deviation from the Work Plan or the QAPP procedures was recorded in the Variance Log shown in Appendix J.

As part of the quality assurance/quality control (QA/QC) measures, the Project Quality Control Officer conducted a QA/QC audit of sample collection activities during the first and second



groundwater sampling events. The audit did not identify any procedures or activities that deviated from the QAPP or impacted the quality of the data.

A Site sampling inspection visit was also conducted by Ms. Sherrel Henry of the USEPA on November 8, 2000 during the first sampling event. The inspection did not identify any procedures or activities that would adversely affect the quality of the data.

The entire field investigation program was conducted with USEPA contractor oversight provided by TAMS. The TAMS oversite person recorded field data (sample locations, depths of borings, soil classifications, etc.) and collected several split samples of environmental media. The samples collected by TAMS were sent to a USEPA selected laboratory for analysis of select parameters.

Geomatrix collected blind duplicates and matrix spike/matrix spike duplicates (MS/MSD) at a quantity of one in every 20 samples for each environmental media. A trip blank, analyzed for the most comprehensive VOC list accompanied each cooler of aqueous media to be analyzed for VOCs. An equipment blank was collected on non-dedicated equipment prior to collection of Site environmental media samples. The equipment blank was analyzed for the COPC list requested for the Site sample. Table 2-2 summarizes the QA/QC sample locations. The correlation between samples and duplicate samples are provided in Table 2-3. The relative percent difference (RPD) between detected compounds in landfill gas media are fairly large. This is likely due to the removal and uptake of more ambient quality air in the gas probe riser pipe during collection of the first gas sample using the Summa canister compared with more concentrated gas that entered the gas probe during duplicate sample collection (second sample collected from the same gas probe).

The laboratory provided complete data packages suitable for full data validation. Data packages were validated by a third party data validator, Ms. Judy Harry of Data Validation Services in North Creek, New York. Data validation reports are provided in Appendix K.

Data validation reported usable data with minor qualifications with the exception of several non-detect total hexavalent chromium and soluble hexavalent chromium values during the second groundwater sampling event (April/May 2001). The rejection of the data was based on low percent recovery of hexavalent chromium in the matrix spike sample. The matrix spike sample was selected from groundwater in the Inactive Landfill Area. The matrix spike sample was spiked with a known concentration of hexavalent chromium and the sample analyzed to



determine the concentration of the spike. The laboratory indicated poor to no recovery of the matrix spike (100 mg/l of hexavalent chromium). Negative matrix interference can be caused by numerous factors including certain chemical presence (i.e., sulfate compounds) and organic matter in the sample media. However, it was determined that the redox conditions of the sample (negative Eh values, low dissolved oxygen) naturally caused reducing conditions in the sample to exist thereby converting the hexavalent chromium spike to trivalent chromium.

The data validator indicated that the laboratory was operating with proper procedures and made an effort to verify negative interference through multiple analysis of project specific samples, additional project specific sample spikes, and laboratory blanks (beyond requirements). However, because of the negative matrix interference effects, the data validator qualified all laboratory-reported non-detects of hexavalent chromium as unusable data (R). Laboratory reported detected values of hexavalent chromium may have low bias and are therefore qualified as estimated (J). The data validator did not reject these values.

Although some hexavalent chromium data are considered not usable (R), where total chromium values were reported as non-detect, it can be assumed that hexavalent chromium is not present above the detection limit in the sample. Samples in which the total chromium was reported as non-detect and the hexavalent chromium was reported slightly above detection limits are likely a result of the difference in method techniques. In these cases, the hexavalent chromium result is qualified by the data validator as estimated (J). USEPA split samples were analyzed using a different analytical method and either did not detect hexavalent chromium, or, when detected, was present at concentrations substantially below groundwater standards.

Matrix interference effects were not identified during the first sampling event because the matrix spike samples were collected from Cattaraugus Creek and the Former Manufacturing Plant Area. These sample locations are not subject to the stronger reducing conditions present in groundwater at the Inactive Landfill Area.

Based on an assessment of precision, accuracy, and completeness, sample collection and laboratory analyses met data quality objectives of the remedial investigation with the exception of certain data rejections for hexavalent chromium. However, the overall characterization of metal constituent concentrations in groundwater was not compromised based on the data quality of total chromium analysis.



3.0 LAND USE AND PHYSICAL CONDITIONS OF THE SITE

As discussed in Section 1.0, the 26-acre Peter Cooper Gowanda Site is comprised of the 15.6-acre Inactive Landfill Area and the 10.4-acre Former Manufacturing Plant Area. The site is located in the Village of Gowanda in Cattaraugus County, NY approximately 30 miles south of Buffalo, New York, and 20 miles east of Lake Erie. The site surroundings include Cattaraugus Creek to the north, Palmer Street to the south, remnants of a former hydroelectric dam and wetland area to the west, and residential properties to the east (see site location map shown on Figure 1-1).

Regionally, the Village of Gowanda is located both in Erie County and Cattaraugus County and is bisected by Cattaraugus Creek. In Erie County, the Village of Gowanda is included in the Town of Collins. The Town of Collins is bordered by the Town of North Collins to the north, to the east by the Town of Concord, to the south by the Cattaraugus Creek and Cattaraugus County (Towns of Otto and Persia) and the Seneca Nation of Indians Cattaraugus Indian Reservation to the west. In Cattaraugus County, the Village of Gowanda is located in the Town of Persia. The Town of Persia is bordered by the Cattaraugus Creek and Erie County to the north, to the east by the Towns of Otto and Perrysburg, to the south by the Town of New Albion, and the Town of Dayton to the west.

3.1 **POPULATION AND LAND USE**

Historical population of the Village of Gowanda from 1940 to 2000 is presented in tabular and graphical form on Table 3-1. According to 2000 Census data, the Village of Gowanda has approximately 2,842 residents. This represents a net reduction of 59 residents from the 1990 Census count of 2,901, with the portion of the Village within Erie County accounting for a decrease of 4 persons and the portion of the Village within Cattaraugus County experiencing a population decrease of 55 persons from the 1990 census data. Census data presented on Table 3-1 also indicate an approximately 13% decrease in population from the period high of 3,352 residents in 1960. Although the Master Plan for the Town of Collins and the Village of Gowanda (Erie County Department of Environment and Planning, August 1999) predicts a population increase in the Village to 3,374 persons through the year 2020, historic census data support a steady to declining population base.

Total populations of all townships surrounding the Village of Gowanda, including the Seneca Nation Cattaraugus Indian Reservation, are presented in Table 3-2. Population figures for two of the six surrounding Cattaraugus County townships declined while the remaining townships



have increased slightly since 1990. In addition, the Cattaraugus Indian Reservation population has increased from a population of 1,789 to 2,001 over the same ten-year period. Population figures for two of the four surrounding Erie County townships indicate a decline while the remaining townships have increased since 1990. While most of the Township population changes are minor, the Town of Collins experienced a 38% increase in population over the period. However, this increase is entirely attributable to increased population at the Collins Correctional Facility, a minimum-security correctional facility located approximately 1.5 miles north of the Village of Gowanda within the Town of Collins.

Zoning surrounding the Village of Gowanda is primarily agriculture, including dairy farming, forestry and minor crop production. Agriculture is the largest industry in towns of Cattaraugus and Erie Counties adjacent to the Village of Gowanda. However, the Village is an urbanized community zoned for a mixture of residential, commercial and industrial land uses. Residential zoning is the dominant parcel designation within the Village. Industrialized zones are primarily concentrated in the southeast portion of the Village, primarily along the Cattaraugus Creek. The Site is located in an area zoned industrial. Appendix L provides a tabular breakdown of land uses and a zoning map for the Village as presented in the 1999 Master Plan for the Village of Gowanda and Town of Collins.

3.2 SITE PHYSIOGRAPHY AND CLIMATE

3.2.1 Site Physiography

The Village of Gowanda is located within a valley and is surrounded by rolling hills with steep slopes. The southeastern portion of the Town of Collins contains the largest area of steep slopes with grades greater than 15%. Cattaraugus Creek bisects the Village and provides the natural boundary between Erie and Cattaraugus Counties. Cattaraugus Creek features are discussed in detail in Section 3.3 below.

Plate 1 presents site topography and surface features for the Site and immediately surrounding property. Within the Inactive Landfill Area of the Site, topography is generally flat with a slight (<1-3%) slope from Palmer Street to the north toward Cattaraugus Creek. An approximately 5-acre elevated fill area is present in the northwestern portion of the Inactive landfill Area. The top of the Cattaraugus Creek bank and remnants of a former hydroelectric dam bound the elevated fill area on its northern and northwestern sides, respectively. The elevated fill area is mounded approximately 10 feet above its surroundings.



Former Manufacturing Plant Area topography also presents a slight (<1 - 4%) grade toward the creek in most locations, with low topographic relief across several large areas in the eastern side of the site where the foundations of former buildings remain.

Vegetative cover, including a mix of low-lying scrub, brush and mature trees characterizes most areas of the site. A cover-type map for the site developed in support of the ecological risk assessment identifies vegetative cover over the Site. Three federal wetland communities were delineated within the boundaries of the Site. An approximately 0.25-acre wetland area, characterized as a combination forested/scrub-shrub wetland, is present at the northeastern limit of the site on the eastern side of the former hydroelectric dam. A 36-inch municipal storm water outfall pipe discharges into the southern portion of this wetland area near the base the adjacent cemented-stone remnants of the former hydroelectric dam. Elsewhere, a small (less than 1,200 square foot) emergent wetland exists in a depression along the southern side of the elevated fill area, and an approximately 3,000 square foot scrub-shrub wetland is located in the center portion of the site. A 12-inch storm water outfall, which drains a small section of ditch along the south side of Palmer Street near the former Peter Cooper office entrance drive, discharges to the site at the southern end of the scrub-shrub wetland.

In addition to the former hydroelectric dam remnants, other surface features present at the site include an approximately 150-foot long riprap revetment wall along the Creek Bank adjacent to the elevated fill area of the site. The riprap was constructed in January 1997 to prevent erosion of the elevated fill area by Cattaraugus Creek. In addition, a former concrete sluiceway and retaining wall are present along the majority of the creek bank in the Inactive Landfill Area and Former Manufacturing Plant Area of the site, respectively. The inner wall and base of the former sluiceway are covered with the exception of an approximately 5-foot section of the sluiceway near the upstream limit (i.e. sluiceway entrance). As indicated above, several former building foundations and slabs are present across the Former Manufacturing Plant Area of the Site. Masonry, lumber and other construction demolition debris are also piled on the Former Manufacturing Plant Area in significant quantities.

3.2.2 Climate

The Site is located in the southeastern portion of the Village of Gowanda, which lies in both Erie and Cattaraugus counties of western New York State. This area of New York has a cold continental climate. Moisture evaporating from Lake Erie causes heavy winter snowfalls along the high ridges closest to the lake, averaging approximately 165.5 inches per year (NOAA,



1998). Annual precipitation averages approximately 49 inches per year (NOAA, 1998). Average temperatures range from 21 degrees Fahrenheit in January to 66 degrees Fahrenheit in July (NOAA, 1998). The ground surface and lakes generally remain frozen from December to March. Natural stream temperatures range from 32 degrees Fahrenheit in winter to 81 degrees Fahrenheit in summer (O' Brien & Gere, 1989). Winds are generally from the southwest (240 degrees) with a mean velocity of 11.6 miles per hour (Buffalo Airport, 1998).

3.3 SURFACE WATER

3.3.1 Storm water

The majority of storm water generated at the site and surrounding property drains through a combination of infiltration and/or overland flow toward Cattaraugus Creek. With the exception of the 12-inch outfall present at the head of the above-described scrub-shrub wetland, engineered storm water conveyance in the vicinity of the site is generally limited to the areas near the intersection of Palmer and Broadway Streets. During the 1990s, the Village of Gowanda installed a storm water collection and conveyance system that services Broadway Street. The collection and conveyance system discharges to a 36-inch PVC conveyance line that runs beneath the open lot area on the southwest side of the Site, and discharges to the forested/scrub-shrub wetland on the northwest side of the Site. Accordingly, the wetland serves as a sediment settling area for storm water discharge. The wetland drains in a northerly direction via overland flow into the Cattaraugus Creek.

3.3.2 Cattaraugus Creek

The Cattaraugus Creek is a surface water body suitable for fishing and secondary recreation (not primary contact recreation such as swimming) but not as a drinking water supply (NYSDEC designated Class C(T)). The Cattaraugus Creek watershed predominantly drains a rural environment that varies in topographic nature from hilly terrain, steep slopes and narrow valleys upstream of the Village to a generally flat slope and wide valley downstream of Gowanda (Wendell-Duchscherer, Flood and Hazard Mitigation Plan for the Village of Gowanda, April 2001). The drainage area of the Creek is approximately 436 square miles and its length is approximately 70 miles. In the vicinity of the Site, the Creek meanders through an incised bedrock valley cut by thousands of years of stream flow. The Creek channel width is 130 feet and of variable depth in the area forming the northern Site property boundary. Cattaraugus Creek flows in a westerly direction eventually discharging into Lake Erie at Irving.

A USGS Gauging Station (#04213500) is located on Cattaraugus Creek west of the Route 62 bridge after the confluence of the east and west branches of Cattaraugus Creek. Stream flow



data collected from the USGS gauging station indicates a mean annual stream flow of 1,030 cubic feet/second (USGS, 2001). No significant discharges occur to the Creek within a few miles of the Site from upstream sources. Presently, the Village of Gowanda sewage treatment plant outfall discharges to Cattaraugus Creek approximately 2 miles downstream of the Peter Cooper Gowanda site. A mean annual stream flow near the Peter Cooper Site of approximately 600 cubic feet/second is reported by O'Brien & Gere in the 1989 RI Report. However, this flow rate has not been confirmed.

The 100-year and 500-year floodplain areas for the Village are mapped in Appendix M. The flood plains are located at varying distances and elevations from the banks of Cattaraugus Creek and are positioned along the entire length of the Creek as it bisects the Village. The elevation of the 100-year flood elevation is approximately is 768 feet mean sea level. Within the Village all water is drained to three different watersheds, the Grannis Brook watershed, Thatcher Brook watershed and the Cattaraugus Creek watershed (Wendell-Duchscherer, Flood and Hazard Mitigation Plan for the Village of Gowanda, April 2001). The Village's primary municipal water supply, the Point Peter Reservoir, is located approximately 1.6 miles south of the site in the Point Peter watershed.

3.4 **REGIONAL GEOLOGY AND HYDROGEOLOGY**

The Peter Cooper Gowanda Site is situated within the uplands of the Allegheny Plateau. The bedrock geology of the area consists of Upper Devonian age shales and siltstones of the upper Canadaway Group (Hazen and Sawyer, 1969). According to the Geologic Map of New York - Niagara Sheet (1970), the Canadaway Group ranges in thickness from 700 to 1,200 feet. The Canadaway Group consists of the following bedrock units, from oldest to youngest: Northeast Shale; Shumla Siltstone; Westfield Shale; Laona Shale; Gowanda Shale; South Wales Shale; Dunkirk Shale and the Machias Formation. The Geologic Map of New York shows the site to be underlain by the Machias Formation, which consists primarily of shales and siltstones, with some sandstones. The bedrock dips uniformly south at 31 to 58 ft/mile (Hazen and Sawyer, 1969) and is cross-cut by a near perpendicular regional joint set oriented northeast-southwest and northwest-southeast. Bedrock topography varies considerably across the area resulting from Wisconsinan age glaciation (13,000-14,000 ybp). Glacial advance and recession resulted in broad river valleys, separated by extensive uplands.

Overburden deposits vary considerably across the region, and reflect the glaciation and subsequent deglaciation of the area. Basal till and glacial moraines consisting of sand, silt and gravel, mark the advance of the ice sheet across the southern tier of Western New York State.



One such glacial moraine cuts across the Village of Gowanda at Cattaraugus Creek east of Route 62 (Yager, et al.,1997). Locally, finer grained deposits (silt/clay) often predominate where glacial melt water remained. Alluvial deposits derived from re-worked glacially-derived soil occur along major streams and creeks in the area.

Groundwater flow occurs both in the unconsolidated glacial deposits across the region, and along bedding plane partings and joints in the bedrock. Confined water table conditions exist where glaciolacustrine deposits (silt/clay) overlie the coarser sand and gravel water bearing units. Buried ancestral river valleys are often floored by the coarser grained deposits associated with glacial advance, which are then overlain by lacustrine fine sand, silt and clay (Yager et al., 1997). Where these coarse-grained deposits are nearer the surface, unconfined conditions predominate.

Domestic water supply well and gas well test hole data located approximately 1-mile northeast of the Village of Gowanda identify the presence of a buried bedrock valley partially filled with glacial drift. The bedrock valley which extends to depths of more than 400 feet below ground surface is overlain by a relatively shallow unconfined deltaic aquifer (Todd, 1998). Recharge to the unconfined aquifer occurs from precipitation that infiltrates the surface soil and from upland sources along the southeastern border of the aquifer, such as runoff from hillsides and seepage from the overburden. The elevation of the upper and lower bounds of the aquifer is higher than the elevation of Cattaraugus Creek. No other widespread deposits that could be considered a high yield aquifer exist in the Gowanda area.

3.4.1 Site Geology

The Site is underlain by shale bedrock of the Canadaway Formation. Shale outcrops in and along Cattaraugus Creek, across the northern site perimeter, and the hill slope south of Palmer Street. The elevation of the bedrock surface generally slopes in a northwesterly direction, toward the Creek. The bedrock topography for the Inactive Landfill Area and Former Manufacturing Plant Area are shown on Figures 3-1A and 3-1B. The topographically flat area between the elevated areas south of Palmer Street and the Creek is a broad alluvial valley with a relatively thin layer of alluvial deposits (approximately 10 feet or less) mantling the bedrock valley floor. Anthropogenic activities have deposited fill above the alluvium. In some areas, excavations have removed alluvial soils and fill materials backfilled the excavations. Collectively, the alluvial soil and fill materials comprise the overburden at the Site. The Site stratigraphy is illustrated in cross section on Figures 3-2, 3-3 and 3-4. Cross section profile lines for each cross-section are shown on Figures 2-3 and 2-4. Based on boring data



summarized in Table 3-3, the thickness of the overburden varies from a few feet to more than 23 feet (GMW-1). The presence and thickness of fill also varies depending on location. The fill is characterized in this RI as sludge fill and cindery fill. Each fill type is usually covered with a topsoil-rich, vegetated layer. The sludge fill, which is located in the Inactive Landfill Area, is a focus of the RI as a potential source for chemical constituents that may impact environmental media. The sludge fill is characterized in more detail in Section 3.5. The cindery fill consists of silt, sand, and gravel with variable amounts of cinders, ash, and construction and demolition materials. The depth to bedrock and the thickness of cover soil and fill at soil boring, monitoring well, piezometer, and gas monitoring well locations is summarized in Table 3-3.

3.4.2 Site Hydrogeology

The overburden and upper bedrock water bearing zones were investigated. The spatial relationship between these two zones is shown in the cross sections referred to in Section 3.4. Groundwater from both zones discharges to Cattaraugus Creek. Seeps are observed at the overburden/bedrock contact and in the bedrock outcrop along the Creek.

3.4.2.1 Hydraulic Properties

The horizontal hydraulic gradient in the overburden is less than 0.01 across much of the Site and is higher near the bank of Cattaraugus Creek as discharge occurs via seeps. Groundwater is not present in the thin overburden deposits in the southeastern portion of the Former Manufacturing Plant Area near Palmer Street. Measured groundwater elevations ranged from



762.49 to 775.35 in overburden monitoring wells in measurements taken in November 2000 and April 2001. The change in groundwater elevation between the two measurements, which were taken approximately 5 months apart, was typically less than 2 feet. Groundwater elevations in monitoring well MW-6, a relatively steep hydraulic gradient between PZ-1 and DP-1 (ranging from 0.12 to 0.31), groundwater in gas monitoring wells, and groundwater seepage into test pit excavations, all suggest an area of elevated head occurring in the landfill sludge fill area of the Inactive Landfill Area. The elevated head produces a slight radial overburden groundwater flow pattern in the northwestern portion of the Inactive Landfill Area. Factors causing the elevated head include: the elevated topography of the area, the restriction of groundwater flow in an easterly direction due to the presence of the concrete foundation from the former hydroelectric dam, and the apparent higher permeability of the sludge fill compared to the surrounding soil/fill. The head variation between seasonal low and high water table conditions are substantially greater for wells screened in the sludge fill vs. wells screened in the cindery fill or native soil. This indicates a seasonal buildup and slow release of groundwater in and from the sludge fill.

The hydraulic conductivity of the overburden, based on slug tests, ranges from 2.9 x 10^{-6} (MW-7S) to 2.2 x 10^{-2} (MW-5S) cm/sec. Hydraulic conductivity testing results are summarized in Table 3-5. The table presents ranges of hydraulic conductivity values estimated for the two fill types (sludge fill and cindery fill) and the native alluvial soil. The fairly wide variation in hydraulic conductivity reflects the variable consistency of overburden material, which includes fill, silt, sand, and gravel.

The horizontal hydraulic gradient in the bedrock is higher than in the overburden, ranging from 0.02 to 0.03 across the Site. Measured groundwater elevations ranged from 753.44 to 777.49 feet above mean sea level. The fluctuation in groundwater elevation between the two measurements was less than in the overburden, and was typically less than 1.5 feet. The calculation of vertical hydraulic gradients is also shown on Table 3-4. With the exception of MW-5S and MW-5D, vertical gradients are downward between the overburden and bedrock.

Anomalously high bedrock water levels were measured in monitoring well MW-5D. Based on the hydraulic gradient established for the bedrock at other areas of the Site and the downward vertical head potential between the overburden and bedrock, a lower head level at MW-5D is expected. However, the similarity in groundwater elevations between MW-5S and MW-5D and the hydrochemistry of the wells (Piper diagrams are further discussed in Section 5.0) suggest that the shallow bedrock and the overburden are locally hydraulically connected. The



hydraulic conductivity of the bedrock at MW-5D location was the highest measured on-Site. Similarly, the hydraulic conductivity of the overburden at that location was also relatively high. A localized area of higher hydraulic conductivity in the bedrock is likely responsible for the elevated bedrock head. Therefore, the hydraulic influence from the elevated head is local to the area of MW-5D.

The vertical hydraulic gradient within the bedrock is upward based on head data for MW-4D(R) and MW-4D2. The upward vertical gradient suggests that bedrock groundwater discharges to Cattaraugus Creek from at least the upper 35 feet of bedrock.

The hydraulic conductivity of the bedrock, based on slug tests, ranges from 2.2 x 10^{-6} (MW-8D) to 3.4 x 10^{-2} (MW-5D) cm/s. The hydraulic conductivity generally decreases with increasing depth in the shale. Estimated hydraulic conductivity values in the shallow zone were in the approximate range of 1 x 10^{-4} to 1 x 10^{-2} cm/s. However, estimated hydraulic conductivity values in the deep bedrock wells, such as MW-4D2 and MW-8D are less than the range of shallow bedrock hydraulic conductivity values. Lower hydraulic conductivity with increasing depth is also supported with packer testing results. Packer test estimated hydraulic conductivity values are summarized below.

| Test Interval (feet bgs) | Hydraulic Conductivity (cm/s) |
|--------------------------|---|
| 18 to 23 | 3.6 X10-5 |
| 22 | |
| 23 to 28 | Not quantifiable; likely less than 1 X 10-6 |
| 28 to 33 | 1.6 X 10-5 |
| 33 to 38 | Not quantifiable; likely less than 1 X 10-6 |

Based on this vertical variation in hydraulic conductivity, the uppermost 10 to 15 feet of bedrock appears to be more permeable than deeper bedrock.

3.4.2.2 Description of Conceptual Groundwater Flow

The overall groundwater flow system at the site consists of two primary zones: groundwater in the overburden and bedrock groundwater. This division is based on stratigraphy, groundwater elevation and hydraulic conductivity data. Groundwater flow in the overburden is controlled



by elevation, bedding surfaces, variation in grain size, the presence of fill, and the bedrock surface. Within the overburden, there is a horizontal hydraulic potential toward the Creek and a downward hydraulic potential from the overburden into the bedrock across the Site, based on a comparison of groundwater elevations in paired overburden/bedrock monitoring wells. A localized westerly flow direction occurs in the overburden near the elevated portion of the Inactive Landfill Area. This flow component is limited in a westerly direction due to the relatively rapid ground surface elevation change that occurs between Palmer Street, the landfill area, and the wetland area adjacent to the Creek. The presence of seeps along the overburden/shale contact along the Creek and the downward vertical hydraulic gradient potential indicate that the physical characteristics of the shale pose a hindrance to groundwater flow from the overburden into the shale. Groundwater flow in the bedrock is primarily along fractures, joints, and bedding planes, which tend to be strongly horizontally oriented. The range of hydraulic conductivity in the upper bedrock is comparable to that of the overburden, where the hydraulic conductivity of the deeper bedrock is orders of magnitude lower. These results suggest that the upper bedrock is more transmissive than the deeper bedrock and the majority of bedrock groundwater receiving recharge from the on-Site overburden discharges to Cattaraugus Creek from the upper 10 to 15 feet of bedrock.

An estimate of Site groundwater contribution (groundwater flux) to surface water in Cattaraugus Creek was prepared using Darcy flux calculations for the overburden and upper bedrock (upper 35 feet) across the northern Site boundary. Darcy flux calculations are provided in Appendix N. The groundwater flux from the overburden for the combined Inactive Landfill Area and the Former Manufacturing Plant Area was estimated to be approximately 550 ft³/day. The groundwater flux from the bedrock for both the Inactive Landfill Area and the Former Manufacturing Plant Area was estimated to be approximately 2,500 ft³/day. The comparatively larger groundwater flux from the bedrock primarily results from the assumption of a substantially greater saturated thickness (35 feet in the bedrock vs. 4 to 6 feet in the overburden) since the geometric mean hydraulic conductivity for the overburden and bedrock are similar. The combined groundwater flux of approximately 3,050 ft³/day from the Site is minor compared to the mean annual stream flow of Cattaraugus Creek. The mean annual stream flow for the Creek near the Site is over 600 ft³/s where as the flux from the Site to the Creek is approximately 0.035 ft³/s.

3.5 CHARACTERIZATION OF SOIL/FILL/SEDIMENT

The Remedial Investigation characterized the physical properties of the various fill types and soil, Creek and wetland sediments. Characterization of the sludge fill and soil cover in the



Inactive Landfill Area, cindery fill and soil in the Former Manufacturing Plant Area, wetland sediment, and Creek sediment is described below.

3.5.1 Characterization and Delineation of the Sludge Fill

Sludge fill was disposed of in the topographically elevated area of the Inactive Landfill Area. The fill appears to extend down to the weathered bedrock surface near the Creek side of the Site. Farther from the creek, the sludge fill is under lain by native alluvial soil. The lateral limits of the buried sludge fill are illustrated on Figure 3-9. The limits of the sludge fill were based on observations made during the excavation of test pits near the perimeter of the elevated fill area. Table 3-6 summarizes test pit soil descriptions for test pits TP-A through TP-E. The results of surface geophysical survey (EM) show anomalously elevated soil conductivity values associated with fill materials buried in the Inactive Landfill Area. However, the conductivity of the sludge fill did not significantly contrast from the conductivity of other fill material observed at the Site and its usefulness in delineating the sludge fill at the Inactive Landfill Area is limited. EM survey results are presented in Appendix C. The thickness of the sludge fill is shown on Figure 3-10. Approximately 100,000 cubic yards of sludge fill is present at the Inactive Landfill Area.

The sludge fill consists primarily of a black, silt and fine sand matrix with various mixtures of animal hair, ash and cinders, gravel, and construction and demolition debris (bricks, glass, concrete, wood) and is associated with a strong ammonia and sulfurous-type odor. Geotechnical testing information for the sludge fill is presented in Appendix O and is summarized in Table 3-7. Grain size analysis of the sludge fill indicates the material is well graded consisting of nearly equal parts of gravel, sand, and silt size particles. The high liquid limit of the sludge is unlike natural soil and indicates a high water-bearing capacity of the material. The composite sample tested has a relatively high water content (approximately 41%) and a vertical conductivity (Shelby tube) of 1.7×10^{-5} cm/s. Based on the rapid infiltration of water into several test pits, however, the bulk hydraulic conductivity of the sludge fill is likely higher. Test pits excavated into the sludge fill, TP-1/G through TP-5/G indicate water rapidly infiltrated the excavations at depths of four feet and deeper. Test pit side walls often slumped into the excavation. Strong odors accompanied the excavations.

Section 1.2.1 reports the installation of a soil cover during the early 1970s. The soil cover over the sludge fill area was investigated and ranged in thickness from approximately 10 inches to over 45 inches. However, a localized area near GMW-2 shows vegetative stress and cover

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soils are very thin to absent. The thickness of the soil cover is illustrated on Figure 3-11. Several test pits encountered a geotextile fabric below the soil cover. Descriptions of top soil and cover soil thickness at each of the 24 test hole locations are summarized in Table 3-8. Geotechnical testing results of the cover soils are presented in Appendix O. Table 3-9 summarizes the physical parameters of the composite cover soil samples and undisturbed (Shelby tube) cover soil samples. The results indicate that the existing cover consists primarily of silt and clay with some fine sand and minor gravel with areas having relatively low hydraulic conductivity. Vertical hydraulic conductivity of cover soil (laboratory analysis from Shelby tubes) ranged from 3.6×10^{-5} to 9.0×10^{-8} cm/s. Clay content of cover soil ranged from 10.0 to 22.1 percent.

3.5.2 Characterization of Cindery Fill and Soil

A layer of a cindery fill soil mixture is present across most areas of the Peter Cooper Site. The thickness of the fill generally increases in a northerly direction across the Site. The fill extends down to the top of bedrock in several areas of the Inactive Landfill Area and the Former Manufacturing Plant Area. The fill unit is shown on cross-sections presented earlier in this section and its thickness is summarized in Table 3-3. Descriptions of the fill are included in logs for soil borings SB-1 through SB-10 and monitoring wells completed in the Former Manufacturing Plant Area and test pit excavations TP-1 through TP-9. The distribution of the cindery fill soil in the Inactive Landfill Area, containing higher quantities of slag is clearly shown on the results of the EM survey as elevated soil conductivity values (Appendix C).

The cindery fill soil consists primarily of a dark gray to black, silty sand matrix with various mixtures of gravel, cinders and ash, slag, and construction and demolition debris (bricks, glass, concrete, wood). Geotechnical testing information for the fill is presented in Appendix O and is summarized in Table 3-10. Grain size analysis of the fill indicates the material is highly variable in composition. Generally, coarser grain-size materials are present at shallow depths with many samples containing nearly 50% gravel size material. At depth, grain size decreases and grain size consists primarily of fine sand and fines (silt and clay).

3.5.3 Wetland Sediments

The wetland area resulting from storm water drainage from Palmer and Broadway Streets is underlain by a layer of organic-rich alluvial soil (referred to as wetland sediments). The thickness of the wetland sediments is greater than five feet based on the capability of advancing the drive point piezometer (DP-1) into the sediment without refusal. The sediment is brown to dark brown and contains substantial vegetative matter. Grain size analyses of three samples



indicates that the sediments consist primarily of fine sand and silt. Grain size testing results are presented in Appendix O and are summarized in Table 3-11.

3.5.4 Cattaraugus Creek Sediments

Sediment in Cattaraugus Creek is sparse because of the relatively high velocity stream flow and the shale bedrock that forms the side walls and streambed of the Creek. Where present, the sediment occurs as occasional small sandbars located near shore. Sediment was identified and sampled at an upstream (SED-1) and downstream (SED-4) location and two locations across from the Site (SED-2 and SED-3). The sediment consists primarily of sand with little to no gravel. Grain size analysis for the four sediment samples are presented in Appendix O and are summarized in Table 3-11.



4.0 CHEMICAL PRESENCE IN SITE MEDIA

The sampling programs presented in Section 2.0 describe laboratory analysis of Site media to assess chemical presence at the Site. The following subsections describe the chemical analytical results in the following media:

- Sludge fill;
- Surface and subsurface soil/fill;
- Landfill gas;
- Groundwater;
- Seeps;
- Cattaraugus Creek surface water; and
- Sediments in the wetland and Cattaraugus Creek.

4.1 SLUDGE FILL

Chemical analytical results of the sludge fill present in the Inactive Landfill Area are based on three samples (GMW-1 through GMW-3) that were analyzed for VOCs and one composite sample that was analyzed for SVOCs and metals. The chemical data for the sludge fill are presented in Table 4-1 for the VOCs and Table 4-2 for the SVOCs and metals. A summary discussion of analytical results follows.

Samples of the sludge fill contained concentrations of some VOCs. The VOCs detected at the highest concentrations are as follows:

Acetone15 mg/kg;2-Butanone3.2 mg/kg, andToluene1.7 mg/kg.

The following twelve VOCs were also detected, but at only trace (i.e., less than 1 mg/kg) concentrations:

1,1-Dichloroethane;
1,2-Dichlorobenzene;
2-Hexanone;
4-Methyl-2-pentanone;
Benzene;
Carbon disulfide;
Chlorobenzene;
Ethylbenzene;



Xylenes; Methylcyclohexane; Styrene; and Tetrachloroethene.

The composite sample of the sludge fill contained detectable concentrations of SVOCs. The SVOCs and the concentration at which they were detected in the sample are as follows:

| 4-Methylphenol | 150 mg/kg; |
|-------------------|----------------|
| Naphthalene | 22 mg/kg; |
| Phenol | 15 mg/kg; |
| Pentachlorophenol | 6.8 mg/kg; and |
| Phenanthrene | 1 mg/kg. |

The composite sample of the sludge fill contained concentrations of metals that are COPCs. These metals and the concentrations at which they were detected in the sample are as follows:

| Arsenic | 34.8 mg/kg; |
|----------|--|
| Chromium | 9,280 mg/kg (hexavalent chromium was not detected at a detection limit of 6.75 mg/kg); and |
| Zinc | 6,060 mg/kg. |

The locations at which the composite sample was collected are illustrated on Figure 2-5A along with the results for the metals COPCs.

The sludge fill sample contained 10.0 percent total organic carbon.

For comparison, maximum concentrations of chemical constituents detected by the USEPA during the 1996 Weston Study are as follows:

| 4-Methylphenol | 6.4 mg/kg; |
|---------------------|---------------------------|
| Phenol | 0.79 mg/kg; |
| Arsenic | 33.0 mg/kg; |
| Hexavalent Chromium | Not detected (10U mg/kg); |
| Chromium | 37,000 mg/kg; and |
| Zinc | 5,200 mg/kg. |

Similar compounds were detected during the 1989 investigation completed by OBG (see Appendix B-3).

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4.2 SOIL

Chemical data for soil samples collected during the RI are presented in the following sections. Previously-collected chemical data for soil samples, that is, data collected prior to the RI, are provided in Appendix B. The RI analytical data are presented for the Inactive Landfill Area first, followed by the FMP area. USEPA Region 9 Preliminary Remediation Goals (PRGs) for industrial soil from October 2002 (herein referred to as soil criteria) are presented for comparison. PRG soil screening levels (SSLs) for a dilution attenuation factor of 20 are also provided to assess a potential for chemical migration via leaching from soil. Values are shaded on data summary tables if detected organic compound concentrations are above PRG soil criteria. Metals are shaded if detected concentrations are above Eastern U.S. soil ranges referred to by various regulatory agencies and presented in NYSDEC Technical and Administrative Guidance Memorandum (TAGM) #4046: Determination of Soil Cleanup Objectives and Cleanup Levels (referred to as background values). The Eastern U.S. soil ranges were selected since background metals data specific to the Village of Gowanda were not available.

4.2.1 Inactive Landfill Area

Both surface soil and subsurface soil were sampled in the inactive landfill area.

4.2.1.1 Surface Soil

Chemical data for 20 surface soil samples, including the results of analysis for VOCs and \cdot metals, are summarized in Table 4-3. VOCs were not detected at concentrations at or above the guidance values. Hexavalent chromium was not detected in any of the samples.

The predefined COPCs arsenic, chromium and zinc were detected above typical concentrations detected in background values in some samples. Inorganic compound concentrations detected above background values are shaded on Table 4-3. Detected concentrations of these chemicals were as follows (ranges included duplicate sample results):

| Arsenic | 1,190 mg/kg to 4 mg/kg (background value is 12 mg/kg); |
|----------|---|
| Chromium | 772 mg/kg to 10.6 mg/kg (background value is 40 mg/kg); and |
| Zinc | 213 mg/kg to 46.9 mg/kg (background value is 50 mg/kg). |

The elevated concentration of arsenic detected in sample LFSS-6 (1,190 mg/kg) could be a result of cindery ash fill present in the area, which also is depicted in the electromagnetic survey. This single sample is not reflective of the arsenic detected elsewhere in the

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environmental media and, therefore, probably is anomalous. Although surface soil was not collected east of the sludge fill area during the 1996 USEPA Study conducted by Weston, the constituent concentrations detected during this RI are generally consistent with the data from the 1989 RI conducted by OBG (see Appendix B-3).

4.2.1.2 Subsurface Soil

Chemical data for 11 subsurface soil samples, including the results of analysis for VOCs and metals, are summarized in Table 4-4. As was observed in the samples of the surface soil, VOCs were not detected at concentrations at or above the guidance values, and hexavalent chromium was not detected in any of the samples.

The COPCs arsenic, chromium and zinc were detected above guidance values in some samples. Detected concentrations of these chemicals were as follows:

| Arsenic | 60.5 mg/kg to 4.3 mg/kg (background value is 12 mg/kg); |
|----------|---|
| Chromium | 623 mg/kg to 7.9 mg/kg (background value is 40 mg/kg); and |
| Zinc | 1,390 mg/kg to 57.3 mg/kg (background value is 50 mg/kg). |

These data are consistent with subsurface soil analytical obtained from the 1996 USEPA study conducted by Weston. Maximum concentrations reported by Weston were:

Arsenic25 mg/kg;Chromium750 mg/kg; andZinc520 mg/kg.

4.2.2 Former Manufacturing Plant Area

The objective of sampling and analysis in the Former Manufacturing Plant Area included determining COPCs as well as characterizing their magnitude and extent, unlike the Inactive Landfill Area, where COPCs had already been determined. As a result, the soil samples collected in the Former Manufacturing Plant Area were analyzed for a more comprehensive list of chemicals than were the soil samples collected in the Inactive Landfill Area.

Both surface and subsurface soil samples were collected in the Former Manufacturing Plant Area.



4.2.2.1 Surface Soil

Chemical data for 10 surface soil samples, including the results of analysis for VOCs, SVOCs and metals, are summarized in Table 4-5. Concentrations detected above guidance values are discussed below. The guidance values shown below are USEPA Region 9 Industrial PRGs (October 2002).

VOCs were detected above guidance values at only one location, MWFP-3, with the exception of acetone, which was detected at SB-9 and SB-10, at concentrations of 1.4 and 0.21 mg/kg, respectively. The soil sample from MWFP-3, collected at a depth of between 0.5 to 2.5 feet below ground surface, contained the following VOCs at concentrations above Region 9 PRG values:

| Carbon tetrachloride | 10 mg/kg (guidance value is 0.55 mg/kg); and | |
|--|--|--|
| Tetrachloroethene | 54 mg/kg (guidance value is 3.4 mg/kg). | |
| The presence of these VOCs in soil a | t sample location MWFP-3 was further investigated to | |
| better ascertain the extent of VOC in | pact in this area. Appendix P summarizes the | |
| investigation methodology and results of an investigation that characterized the lateral and | | |
| vertical extent of VOC impacts in surface soil. The investigation results indicate an area | | |
| approximately 20 feet by 40 feet that contains VOC concentrations. Concentrations were | | |
| below Region 9 Industrial PRG values. | | |

SVOCs were detected above guidance values in several samples. Detected concentrations of these chemicals were as follows:

| Benzo (a) anthracene | 24 mg/kg to 0.11 mg/kg (guidance value is 2.1 mg/kg); |
|--------------------------|--|
| Benzo (a) pyrene | 20 mg/kg to 0.087 mg/kg (guidance value is 0.21 mg/kg); |
| Benzo (b) fluoranthene | 15 mg/kg to 0.079 mg/kg (guidance value is 2.1 mg/kg); |
| Dibenzo (a,h) anthracene | 5.2 mg/kg to 0.076 mg/kg (guidance value is 0.21 mg/kg); and |
| | |

Indeno (1,2,3-cd) pyrene 13 mg/kg to 0.043 mg/kg (guidance value is 2.1 mg/kg).

The upper range of SVOC concentrations are above SSLs indicating a potential for migration to groundwater. However, these compounds were not detected in groundwater samples collected and analyzed from the are.

Metals were detected above guidance values in several samples. Detected concentrations of these chemicals were as follows:

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| Arsenic | 168 mg/kg to 6.6 mg/kg (background value is 12 mg/kg); |
|-----------|--|
| Calcium | 44,200 mg/kg to 1,050 mg/kg (background value is 35,000 mg/kg, exceeded in just one sample); |
| Chromium | 198 mg/kg to 9 mg/kg (background value is 40 mg/kg); |
| Copper | 171 mg/kg to 20.9 mg/kg (background value is 50 mg/kg); |
| Lead | 269 mg/kg to 8.2 mg/kg (background value is 200-500 mg/kg); |
| Magnesium | 12,600 mg/kg to 225 mg/kg (background value is 5,000 mg/kg); |
| Mercury | 3.1 mg/kg to < 0.05 mg/kg (background value is 0.2 mg/kg, exceeded in just one sample); and |
| Zinc | 728 mg/kg to 45.6 mg/kg (background value is 50 mg/kg). |

Hexavalent chromium was not detected in any of the samples.

4.2.2.2 Subsurface Soil

Chemical data for 12 subsurface soil samples, including the results of analysis for VOCs, SVOCs and metals, are summarized in Table 4-6. Concentrations detected above guidance values are discussed below. The guidance values shown below are USEPA Region 9 Industrial PRG values.

VOC results were below guidance values, except for the detection of 0.49 mg/kg acetone in a soil sample collected at SB-5. The acetone concentration at SB-5 is an estimated value, as it was below the detection limit. The USEPA Region 9 Industrial PRG value for acetone is 620 mg/kg.

SVOCs were detected above guidance values in several samples. Detected concentrations of these chemicals were as follows:

| Benzo (a) pyrene | 2.3 mg/kg to 0.058 mg/kg (guidance value is 0.21 mg/kg); and |
|---|--|
| Dibenzo (a,h) anthracene | 0.64 mg/kg to 0.11 mg/kg (guidance value is 0.21 mg/kg). |
| Metals were detected above background values in several samples. Detected concentrations of | |
| these chemicals were as follows: | |

Arsenic 23.6 mg/kg to 3.7 mg/kg (background value is 12 mg/kg);



| Cadmium | 1.3 mg/kg (exceeded in only one sample, the background value is 1 mg/kg); |
|-----------|--|
| Calcium | 67,000 mg/kg to 1,270 mg/kg (exceeded in only one sample, the background value is 35,000 mg/kg); |
| Chromium | 155 mg/kg to 6.2 mg/kg (background value is 40 mg/kg); |
| Copper | 187 mg/kg to 11.3 mg/kg (background value is 50 mg/kg); |
| Lead | 1,950 mg/kg to 7.2 mg/kg (background value ranges from 200 mg/kg to 500 mg/kg); |
| Magnesium | 5,620 mg/kg to 851 mg/kg (exceeded in only one sample, the background value is 5,000 mg/kg); |
| Mercury | 3.1 mg/kg to < 0.05 mg/kg (exceeded in only one sample, the background value is 0.2 mg/kg); and |
| Zinc | 605 mg/kg to 37.8 mg/kg (background value is 50 mg/kg). |

4.3 LANDFILL GAS

The composition and VOC content in landfill gas samples were evaluated. The evaluation used the results of chemical analysis of gas samples collected from three gas monitoring wells, GMW-1, GMW-2 and GMW-3, placed in the inactive landfill area. The construction of the gas monitoring wells is discussed in Section 2.3. The chemical data are presented in Table 4-7.

The chemical data are summarized as follows:

- 1. The LEL was exceeded in two of the samples;
- 2. Hydrogen sulfide was detected at greater than 1,000 ppm in two of the samples and at 710 ppm in the third;
- 3. Oxygen content was just 0.5 ppm or 0% in one sample, and was depressed to 17.5 ppm in a second sample;
- 4. Carbon monoxide was detected in two of the samples, at up to 6 ppm;
- 5. Carbon dioxide was detected at relatively high concentrations in two of the samples, up to 11.2 percent; and
- 6. Methane was detected in two of the samples, up to 31.1 percent.

The following VOCs were detected in the gas samples:

Acetone $1,200 \,\mu g/m^3$ to $150 \,\mu g/m^3$;



| Trichlorofluoromethane | 1.7 μ g/m ³ (detected in only one sample); |
|------------------------|--|
| Carbon disulfide | $3,200 \ \mu g/m^3$ to $93 \ \mu g/m^3$; |
| 2-Butanone | $1,100 \ \mu g/m^3$ to 43 $\mu g/m^3$; |
| Benzene | 180 μ g/m ³ to < 2 μ g/m ³ ; |
| 4-Methyl-2-pentanone | 370 μ g/m ³ to 3.4 μ g/m ³ ; |
| Toluene | 2,600 μ g/m ³ to 41 μ g/m ³ ; |
| 2-Hexanone | 7 μ g/m ³ (detected in only one sample); |
| Ethyl benzene | 84 μ g/m ³ to 3.5 μ g/m ³ ; |
| Xylenes | 130 μ g/m ³ to 1.4 μ g/m ³ ; and |
| Styrene | 20 μ g/m ³ (detected in only one sample). |

The hand held instruments used to measure hydrogen sulfide, methane, and total VOCs (by PID) did not detect these compounds in ambient air.

4.4 GROUNDWATER

Groundwater chemical data were collected during the RI for the overburden and the bedrock groundwater in both the inactive landfill area and the former manufacturing plant area. These data are presented in the following sections. Chemical data for groundwater samples collected prior to the RI are compiled in Appendix B.

The groundwater chemical data are used to compare groundwater chemistry between the bedrock and overburden groundwater, evaluate Site-derived chemical constituents in groundwater, and include parameters that assist in evaluating the fate and transport of chemical constituents in groundwater.

4.4.1 Inactive Landfill Area

Groundwater chemical conditions in the inactive landfill area are presented as follows.

4.4.1.1 Overburden

Chemical data for 16 overburden groundwater samples (i.e., samples from 8 wells in 2 separate sampling events), including the results of analysis for VOCs, SVOCs and metals, and other geochemical data are summarized in Table 4-8. Concentrations detected above guidance values are discussed for VOCs, SVOCs and metals as follows.



VOCs were detected above NYS Division of Water Technical and Operational Series Ambient Water Quality Standards and Guidance Values (groundwater criteria) at four wells, MW-2S, MW-3S, MW-4S and MW-6S. The groundwater samples contained the following VOCs at concentrations above groundwater criteria:

| Benzene | 1.6 μ g/L to not detectable, (groundwater criteria is 1 μ g/L); |
|---------------------|---|
| Chlorobenzene | 190 μ g/L to not detectable (groundwater criteria is 5 μ g/L); |
| 1,2-dichlorobenzene | 5 μ g/L (detected in just one sample, groundwater criteria is 3 μ g/L); and |
| Toluene | 17 μ g/L to not detectable: (groundwater criteria is 5 μ g/L). |

Among the SVOCs, only phenol was detected at a concentration above the groundwater criteria. Phenol concentrations ranged from 480 μ g/L to not detectable. Phenol concentrations were above the groundwater criteria of 1 μ g/L in samples collected from wells MW-2S and MW-3S in both rounds of sampling. The concentrations of phenol and phenolic compounds detected in the overburden during the RI are substantially lower than part per million level concentrations detected during the 1996 USEPA investigation conducted by Weston.

Metals were detected above groundwater criteria in several samples. Detected concentrations of these chemicals were as follows:

| Arsenic | 0.196 mg/L to < 0.01 mg/L, (is $0.025 mg/L$); |
|-----------|--|
| Chromium | 0.436 mg/L to < $0.01 mg/L$, (groundwater criteria is $0.05 mg/L$); |
| Iron | 41 mg/L to < 0.1 mg/L, (groundwater criteria is 0.3 mg/L); |
| Magnesium | 167 mg/L to 16.8 mg/L, (groundwater criteria is 35 mg/L) |
| Sodium | 1,670 mg/L to < 5 mg/L, (groundwater criteria is 20 mg/L). |

Hexavalent chromium was not detected in any of the groundwater samples. As discussed in the discussion of sample QC in Section 2.8, there was matrix interference and consequently some hexavalent chromium data are flagged as unusable (R qualifier). The "matrix interference" is the geochemical condition of the groundwater, in which hexavalent chromium is unstable and rapidly reduces to a lower valence state (i.e., +6 to +3). Therefore, while the sample results are technically unusable based on the low recovery of a matrix spike of hexavalent chromium, this condition indicates the likely real absence of hexavalent chromium in these waters.

Geochemical parameters were used to evaluate the potential presence of leachate from the inactive landfill and to evaluate chemical fate. A pattern is observed in the data with

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comparable (generally elevated) concentrations of several geochemical parameters consistently observed at MW-2S(R), MW-3(R), MW-4S(R), MW-6 and MW-7S. The geochemical data are presented in Table 4-8, and summarized briefly as follows:

Ammonia ranged from 837 mg/L to 1.05 mg/L, with greatest concentration at MW-2S(R), MW-3(R), MW-4S(R), MW-6, and MW-7S;

Nitrate concentrations were either not detected or at trace concentrations (less than 2 mg/L, except at MW-7S, where nitrate was detected at up to 22.7 mg/L in one sampling event and not detected in the second sampling event;

Total Kjeldahl Nitrogen ranged from 839 mg/L to 1.51 mg/L, with greatest concentrations at MW-2S(R), MW-3(R), MW-4S(R), MW-6, and MW-7S;

Alkalinity (bicarbonate) ranged from 3,850 mg/L to 321 mg/L, with greatest concentration at MW-2S(R), MW-3(R), MW-4S(R), MW-6, and MW-7S;

Chloride concentrations were in the range of 61.5 mg/L to 3.82 mg/L, with the exception of MW-7S, where chloride was detected at up to 2,310 mg/L;

Total and soluble organic carbons were at their greatest concentration at MW-2S(R), MW-3(R), MW-4S(R), MW-6 and MW-7S, ranging up to 187.5 mg/L and 112.75 mg/L, respectively;

Sulfate concentration ranged from 960 mg/L to 2.64 mg/L, with lowest concentration (indicating potential sulfate reduction) at MW-2S(R), MW-3(R), MW-4S(R), MW-6 and MW-7S;

Sulfide concentrations ranged from 55 mg/L to <1 mg/L, and were greatest at MW-2S(R), MW-3(R), and MW-4S(R);

Dissolved oxygen concentrations ranged from 9.34 mg/L to 0.2 mg/L, and was less than 1 mg/L at MW-3(R), MW-6, and MW-8S; and

Oxidation Reduction Potential (ORP) ranged from positive 291.1 mV to negative 371.6 mV. ORP was a negative value (indicating the predominance of reducing conditions) at all sampling locations except MW-1S. Even at MW-1(R), one result showed a relatively low ORP of 11.9 mV.

The downhole inductance survey performed in the core hole of monitoring well MW-4D2 did not assist in characterizing the vertical extent of chemical impacts in bedrock groundwater. The downhole survey results are presented in Appendix C. Similar to the surface electromagnetic survey, chemical constituents in the sludge fill are not sufficiently conductive in order to differentiate the conductivity signature of bedrock groundwater impacted with sludge fill chemicals.



4.4.1.2 Bedrock

Chemical data for 14 bedrock groundwater samples (i.e., samples from 7 wells in 2 separate sampling events), including the results of analysis for VOCs, SVOCs and metals, and other geochemical data are summarized in Table 4-9. Concentrations detected above groundwater criteria are discussed for VOCs, SVOCs and metals as follows.

Among VOCs and SVOCs, only one result exceeded groundwater criteria, that for chlorobenzene in one of the two samples collected at MW-4D. The result was 6.8 μ g/L (an estimated value), slightly above the groundwater criteria of 5 μ g/L.

Metals, as total, were detected above groundwater criteria in several samples. Detected concentrations of these chemicals were as follows:

| Arsenic | 0.0483 mg/L to < 0.01 mg/L (groundwater criteria of 0.025 mg/L, exceeded at MW-4D2); |
|------------------------|--|
| Chromium | 0.133 mg/L to < 0.01 mg/L (groundwater criteria is 0.05 mg/L, exceeded at MW-2D(R) and MW-4D(R)); |
| Hexavalent Chromium | 0.0592 mg/L to <0.01 mg/L (groundwater criteria is 0.05 mg/L, exceeded at MW-2D(R); |
| Iron | 71.4 mg/L to 0.115 mg/L (groundwater criteria is 0.3 mg/L, exceeded everywhere with the exception of $MW-2D(R)$); |
| Magnesium | 107 mg/L to 2.6 mg/L (groundwater criteria is 35 mg/L); |
| Sodium | 1,030 mg/L to 19.7 mg/L (groundwater criteria is 20 mg/L). |

Soluble metals concentrations were detected above groundwater criteria at just two locations, MW-1D and MW-4D(R). Iron and sodium levels were elevated compared to groundwater criteria in samples collected from MW-1D and chromium, iron, magnesium and sodium concentrations were elevated in MW-4D(R).

Geochemical conditions were evaluated for bedrock as in the manner described for the overburden. Parameters were used to evaluate the potential presence of chemical constituents from the inactive landfill and to evaluate chemical fate. The geochemical data are presented in Table 4-9. Inorganic compounds detected at concentrations above groundwater criteria in the overburden in the downgradient wells were also present in the bedrock, but at generally lower concentrations. Compounds detected are summarized briefly as follows:

Ammonia ranged from 353 mg/L to 0.716 mg/L, with greatest concentration at MW-2D(R), and somewhat lower but still elevated levels at MW-4D2, MW-4D(R), and MW-5D;

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Nitrate concentrations were either not detected or at trace concentrations (less than 1 mg/L);

Total Kjeldahl Nitrogen ranged from 359 mg/L to 1.29 mg/L, with greatest concentration at MW-2D, and somewhat lower but still elevated levels at MW-4D2, MW-4D(R), and MW-5D;

Alkalinity (bicarbonate) ranged from 2,010 mg/L to 4.67 mg/L, with greatest concentration at MW-2D, MW-4D2, and MW-4D;

Chloride concentrations were in the range of 914 mg/L to 11 mg/L, relatively high compared to the overburden, with the greatest concentrations detected at MW-4D2 and MW-7D;

Total and soluble organic carbons were at their greatest concentration at MW-2D(R) and MW-4D(R), relatively low compared to the overburden, ranging up to 41.7 mg/L and 42.1 mg/L, respectively;

Sulfate concentration ranged from 1,620 mg/L to 2.07 mg/L, with greatest concentrations at MW-5D, and lower but still elevated concentrations at MW-2D(R) and MW-4D(R);

Sulfide concentrations ranged from 9.7 mg/L to <1 mg/L, and were greatest at MW-2D(R) and MW-4D(R);

Dissolved oxygen concentrations ranged from 8.31 mg/L to 0.45 mg/L, and was less than 1 mg/L on average at MW-7D; and

ORP ranged from 202.5 mV to -330.5 mV. ORP was a negative value (indicating the predominance of reducing conditions) at all sampling locations on at least one of the sampling events and was at its lowest at MW-4D(R) and at its highest at MW-8D.

4.4.2 Former Manufacturing Plant Area

Groundwater chemical conditions in the former manufacturing plant area are presented as follows.

4.4.2.1 Overburden

Chemical data for four overburden groundwater samples collected from down gradient wells (i.e., samples from two wells during two separate sampling events) are discussed below. Groundwater is not present in the overburden at the upgradient location MWFP-1D. Analytical results for VOCs, SVOCs and metals, and other geochemical data are summarized in Table 4-10. Concentrations detected above groundwater criteria are discussed for VOCs, SVOCs and metals as follows.


No VOCs were detected above groundwater criteria except for 5.5 μ g/L tetrachloroethene (groundwater criteria of 5 μ g/L) in the first round of sampling from MWFP-3S. Cis-1,2-dichloroethene was detected at the groundwater criteria of 5 μ g/L in the same sample. Chlorinated aliphatic hydrocarbon compounds were also detected in surface soil collected from the boring.

No SVOCs were detected above groundwater criteria.

Metals were detected above groundwater criteria in several samples. Detected concentrations of these chemicals were as follows:

| Iron | 16 mg/L to 0.535 mg/L (groundwater criteria is 0.3 mg/L); |
|-----------|---|
| Manganese | 2.08 mg/L to 0.43 mg/L, (groundwater criteria is 0.3 mg/L); and |
| Sodium | 122 mg/L to 9.98 mg/L, (groundwater criteria is 20 mg/L). |

Hexavalent chromium was not detected in any of the groundwater samples. As discussed in the discussion of sample QC in Section 2.8, there was matrix interference and consequently some hexavalent chromium data are flagged as unusable (R qualifier). The "matrix interference" is the geochemical condition of the groundwater, in which hexavalent chromium is unstable and rapidly reduces to a lower valence state (i.e., +6 to +3). Therefore, while the sample results are technically unusable based on the low recovery of a matrix spike of hexavalent chromium, this condition indicates the likely real absence of hexavalent chromium in these waters.

Geochemical parameters were used to help evaluate the chemical fate of COPCs and for comparison to geochemical conditions in the inactive landfill area. The geochemical data are presented in Table 4-10, and summarized briefly as follows:

Alkalinity was bicarbonate type, and ranged from 700 to 435;

Chloride concentrations were in the range of 63.5 to 10;

Total and soluble organic carbons were not present at detectable concentrations;

Sulfate concentration ranged from 651 to 301, higher than the low sulfate areas observed in the inactive landfill area;

Sulfide was not present at detectable concentrations;

Dissolved oxygen concentrations ranged from 4.81 to 0.42, and was less than 1 at MW-FP-3S; and



ORP ranged from 82 to -31.6 mV, and was a negative value at MWFP-3S.

4.4.2.2 Bedrock

Chemical data for six bedrock groundwater samples (i.e., samples from three wells during two separate sampling events), including the results of analysis for VOCs, SVOCs and metals, and other geochemical data are summarized in Table 4-11. Concentrations detected above groundwater criteria are discussed for VOCs, SVOCs and metals as follows.

VOCs were detected at concentrations slightly above groundwater criteria at MWFP-2D and – 3D. The groundwater samples contained the following VOCs at concentrations above groundwater criteria (results are in $\mu g/L$):

| Acetone | 80 to not detectable, where the groundwater criteria is 50; |
|------------------------|--|
| Benzene | 3.6 (estimated) to not detectable, where the groundwater criteria is 1; |
| cis-1,2-dichloroethene | 8.2 (estimated, detected in just one sample) to not detectable, where the groundwater criteria is 5; |
| m/p-Xylene | 6.4 (estimated, detected in just one sample), where the groundwater criteria is 5; and |
| Toluene | 6.8 (estimated, detected in just one sample) to not detectable; where the groundwater criteria is 5. |

SVOCs were not detected at concentrations above the groundwater criteria in any samples.

Metals were detected above groundwater criteria in several samples. Detected concentrations of these chemicals were as follows (in mg/L):

| Iron | 21.5 to 0.211, where the groundwater criteria is 0.3 ; |
|-----------|--|
| Manganese | 2.06 to 0.0446 , where the groundwater criteria is 0.3 ; and |
| Sodium | 352 to 25, where the groundwater criteria is 20. |

Geochemical conditions were evaluated for bedrock as in the manner described for the overburden. The geochemical data are presented in Table 4-11, and summarized briefly as follows (concentrations are expressed in units of mg/L):

Alkalinity was bicarbonate type, and ranged from 575 to 187;

Chloride concentrations were in the range of 166 to 22.5;



Total and soluble organic carbons were not present at detectable concentrations, except for 4.92 mg/L soluble organic carbon in one sample from MWFP-3D;

Sulfate concentration ranged from 695 to 45.5, generally lower than the values observed in the bedrock in the inactive landfill area;

Sulfide was not present at detectable concentrations;

Dissolved oxygen concentrations ranged from 2.07 to 0.29, and was less than 1 at MWFP-2D; and

ORP ranged from -3.2 to -223.5 mV.

4.5 SEEPS

Chemical data for six samples of seeps from the inactive landfill area, including the results of analysis of two sets of data from three sampling locations, are summarized in Table 4-12. The chemical conditions of the seeps are presented to support an evaluation of the presence of chemical constituents in the seep water and geochemical conditions relevant to the fate and transport of COPCs. The seeps are frequently associated with white, calcium-rich precipitates visible at the contact between the overburden and bedrock and along bedrock outcrops in the Creek immediately downgradient from the sludge fill disposal area. Ammonia and sulfurous-type odors are frequently noted near the seeps.

No VOCs or SVOCs were detected above surface water criteria in any of the samples from the seeps. Some metals were detected above surface criteria and the results for these metals are summarized as follows (results are in mg/L):

Chromium 0.423 to 0.0949 (all but one sample exceeded the hardness-based surface water criteria of 0.120); and

Iron 4.78 to < 0.1, where the surface water criteria is 0.3.

Among other geochemical parameters, ammonia and sulfide were present at elevated concentrations. Ammonia concentrations ranged from 891 to 381 mg/L, where the surface water criteria was 1.1 to 1.3 mg/L (the guidance value varies between sampling event depending on pH and temperature of the sample). Sulfide concentrations ranged between 9 and < 1 mg/L, where the guidance value was 2 mg/L.



4.6 SURFACE WATER

Chemical data for eight samples of surface water from Cattaraugus Creek near the Site, including the results of analysis of two sets of data from four sampling locations, are summarized in Table 4-13.

No VOCs or SVOCs were detected above guidance values in any of the samples collected from Cattaraugus Creek. The only metal detected above surface water criteria was iron, which was detected at concentrations ranging from 0.47 to 0.126 mg/L, where the guidance value is 0.3 mg/L. Concentrations were detected above guidance values during the second round of sampling only. These results are for total iron. Ferrous iron was not detected in the either field or laboratory analyses. Sulfide, which was detected in seeps from the inactive landfill area at concentrations above guidance values, was not detected above guidance values in Cattaraugus Creek. The ammonia concentration in the Creek Water #4 sample (0.442 mg/L) was slightly over the calculated surface water criteria of 0.440 mg/L during the second sampling event.

4.7 SEDIMENT

Sediment samples were collected from the wetland area north of the Site and from Cattaraugus Creek at locations adjacent to the Site.

4.7.1 Wetland Area

Chemical data for 10 samples of sediment/surface soil from the wetland area, including the results of analysis for VOCs and metals, are summarized in Table 4-14. Low concentrations of benzene, toluene, ethylbenzene, and xylenes (BTEX) were detected in all of the samples. The low concentration of BTEX in the samples is likely the result of urban runoff since a Village storm sewer discharges to the wetland. None of the VOCs were detected at concentrations above soil criteria.

The results of the chemical analysis for metals COPCs are summarized as follows (results in mg/kg):

| Arsenic | 16.3 to 5.2, where the background value is 12; |
|----------|--|
| Chromium | 55.3 to 6.5, where the background value is 40; and |
| Zinc | 290 to 45.7, where the background value is 50. |

Hexavalent chromium was not detected in any of the samples.



4.7.2 Cattaraugus Creek

Chemical data for four samples from the sediment in Cattaraugus Creek are summarized in Table 4-15. Trace concentrations (i.e., less than 1 mg/kg) of several VOCs were detected. No SVOCs were detected. Arsenic concentrations (ranging from 6.7J to 9.6J mg/kg) slightly exceeded sediment screening criteria (6 mg/kg) in the four sediment samples. The nickel concentration (18.2 mg/kg) in creek sediment #4 slightly exceeded the sediment criteria of 16 mg/kg. Hexavalent chromium was not detected and total chromium concentrations were not elevated compared to sediment criteria.

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5.0 CHEMICAL MIGRATION ASSESSMENT

The results of the chemical analyses were incorporated with the characterization of the physical setting of the Site to evaluate the fate and transport of chemical constituents in Site media. There are a number of mechanisms by which the chemicals can migrate to other areas or media. These mechanisms are briefly outlined below.

Fugitive Dust Generation. Non-volatile chemicals present in soil can be released to ambient air as a result of fugitive dust generation. Although the majority of the facility is covered by vegetation that would prevent the suspension of surface soil particles, there has been some erosion of surface cover.

Volatilization. Volatile chemicals present in soil and groundwater in certain locations may be released to ambient air through volatilization either from or through the soil or fill. Elevated concentrations of volatile organic compounds are present in an isolated areas of soil/fill at the former manufacturing plant area and in landfill gas. Therefore, the release of these chemicals is relevant to the elevated fill area of the Inactive Landfill Area and a small area of the Former Manufacturing Plant Area. VOCs were also detected in groundwater at the Site. Therefore, the groundwater-to-air pathway may be relevant.

Surface Water Runoff. Chemicals present in on-site soil could be released to Cattaraugus Creek and the adjacent wetland area as a result of surface water runoff. However, the thick grasses and abundant plant growth across the site combined with the site's low topographic relief minimize off-site transport via storm water runoff.

Leaching (percolation). Chemicals present in soil may migrate downward to groundwater as a result of infiltration of precipitation. Chemicals from the site have entered the groundwater system on-site. This potential migration pathway is potentially relevant for the Site.

Groundwater Transport. Groundwater underlying the site discharges to Cattaraugus Creek. Seeps have also been observed along the Creek. Chemicals present in groundwater may be transported to surface water and sediment via this pathway.

5.1 AIRBORNE PATHWAYS

Potential migration pathways involving airborne transport include:



- Wind erosion and transport of soil particles and sorbed chemical constituents in fugitive dust emissions.
- Volatilization of chemical constituents from soils localized in the area of MWFP-3S/D in the Former Manufacturing Plant Area and from the sludge fill in the Inactive Landfill Area and subsequent atmospheric dispersion.

5.1.1 Fugitive Dust

Although the Site is well vegetated and a layer of top soil generally covers the Site, a small amount of fugitive dust emission could occur. The potential significance of fugitive dust emission is evaluated in the Baseline Risk Assessment (BRA) (Geomatrix/Benchmark, November 2002).

5.1.2 Volatilization

Volatile chemical constituents present in Site media could volatilize to the atmosphere and be transported off-site. For surface soils, volatilization of chemicals (if present) would be more or less direct into the atmosphere. For subsurface soils, volatilized constituents would have to diffuse through the overlying soil prior to reaching the atmosphere where off-site transport could occur. Volatilization from surface and subsurface soil in the Inactive Landfill Area and Former Manufacturing Plant Area could result in some chemical migration off-site in air. As described in Section 5.3, the landfill gas monitoring well assessment showed that chemicals are present in the landfill gas generated from decomposition of the waste material. Methane, hydrogen sulfide, and several volatile organic compounds were detected in landfill gas samples collected from the gas monitoring well headspace. In fact, the steel protective well casing covers and locks for wells screened in sludge fill appear to be yellow stained and exhibit corrosive effects believed to be the results of hydrogen sulfide gas reacting with condensed moisture. Hand held air monitoring equipment did not measure detectable concentrations of landfill gases in ambient air. However, landfill gases may slowly diffuse through the cover soils that exist over the sludge fill and may be present at very low concentrations. These pathways are evaluated in the BRA (Geomatrix/Benchmark, November 2002).

Volatilization of chemicals from groundwater is not a significant contributor to volatilization and off-site transport since volatile organic compound concentrations in the groundwater are very low. Volatilization of chemicals from groundwater is therefore not a significant pathway for off-site migration.



5.2 WATERBORNE PATHWAYS

Chemicals in surface soils could be potentially transported off-site via storm water runoff. Chemicals in Site soil could also leach and migrate via groundwater to groundwater discharge areas.

5.2.1 Surface Water Runoff

Erosion and transport of surface soils and associated sorbed chemicals in surface water runoff is a potential migration pathway for the Site. The site's low topographic relief, vegetated nature of the Site, and lack of visible evidence of significant erosion minimize off-site transport via storm water runoff across a majority of the Site. A greater potential for off-site transport via storm water runoff exists along the northern and western perimeter of the sludge fill area where the ground surface slopes rapidly toward the creek and the wetland. However, the generally low chemical concentrations in Site surface soils (see Section 5.0) would not result in significant concentrations in storm water and would not substantially affect off-site surface soil or Cattaraugus Creek. Off-site transport in surface water is therefore not considered to be a significant migration pathway.

5.2.2 Groundwater Migration

Groundwater in overburden and bedrock ultimately discharges to Cattaraugus Creek. In Section 3.5, the total groundwater flow rate from the Site (overburden and bedrock) to Cattaraugus Creek is estimated to be approximately 3,050 cubic feet/day. This rate is less than 0.006 percent of the mean annual stream flow in Cattaraugus Creek indicating that chemical concentrations in discharging groundwater would have to be quite high to result in significant degradation of water quality in Cattaraugus Creek.

Major cation-anion hydrochemistry in Site groundwater was evaluated using trilinear diagrams (Piper plots) to evaluate potential hydrochemical facies changes that may occur as groundwater flows across the site and becomes influenced by different geologic media or groundwater having a different hydrochemical signature. Piper plots are presented in Appendix Q.

The Piper plots show a hydrochemical facies shift in overburden groundwater between upgradient and downgradient wells at the Inactive Landfill Area. The hydrochemistry of overburden groundwater shifts from the no dominant cation-anion facies into the calciumbicarbonate dominant facies. A facies shift in bedrock groundwater hydrochemistry from the sodium/potassium-chloride dominant facies to the calcium-bicarbonate facies was observed for all bedrock wells downgradient of the sludge fill disposal area except MW-4D2. The shift to



the calcium-bicarbonate facies indicates overburden groundwater chemistry in the Inactive Landfill Area affects downgradient shallow bedrock groundwater quality. The lack of a facies shift in MW-4D2 groundwater suggests that deeper bedrock groundwater chemistry is minimally affected by chemistry in overburden and shallow bedrock groundwater. The reason for the lack of hydrochemical impact in the deeper bedrock is likely caused by upward vertical hydraulic gradients in the bedrock.

The hydrochemical facies shift observed in groundwater chemistry between samples collected from monitoring wells (both overburden and bedrock) upgradient of the sludge fill disposal area and downgradient wells is caused by migration of chemical constituents from the sludge fill in groundwater. As a result, chemical compounds such as sulfate, ammonia, and dissolved solids are elevated in overburden and shallow bedrock groundwater downgradient from the sludge fill compared to upgradient.

No significant interpretations can be made from the overburden and bedrock hydrochemistry in the Former Manufacturing Plant Area.

Once chemicals enter the groundwater flow system, the chemical environment of the groundwater influences their fate. The anaerobic and reducing conditions in Site groundwater are amenable to reductive dechlorination and degradation of chlorinated aliphatic compounds such as tetrachloroethene and trichloroethene. However, since reductive chemical transformation compounds were not detected, either complete dechlorination occurs or only attenuation processes of dispersion and dilution are important to their chemical fate. In addition, hexavalent chromium is unstable in a reducing environment. In Site groundwater, the hexavalent species are reduced to trivalent species which are not only less toxic but generally less mobile as well (LaGrega et al., 1994).

As described in Section 4.0, organic chemicals were generally not detected in Site overburden and bedrock groundwater. In the few instances where organic chemicals were detected, concentrations were relatively low. Besides phenol and chlorobenzene, no other organic chemical exceeded the guidance value for groundwater by a factor of more than two and none exceeded guidance values by any amount in more than three monitoring wells. Based on the limited distribution and low concentrations present, organic chemicals in groundwater have limited potential for impacting water quality in Cattaraugus Creek. This is evidenced by the results of water samples obtained from Cattaraugus Creek in which no organic chemicals were measured definitively above detection limits. However, the pathway was considered



potentially complete for organic chemicals and was addressed in the Baseline Risk Assessment (see Section 6.0).

Several metals exceeded guidance values in overburden and bedrock groundwater. As with organic chemicals, metals concentrations were generally low and of limited distribution. In Site overburden and bedrock groundwater, concentrations of metals in excess of 1 mg/L were limited to iron, calcium, magnesium and sodium. Based on the limited distribution and low concentrations present, metals in groundwater have limited potential for impacting water quality in Cattaraugus Creek. The only metal measured in Cattaraugus Creek above its surface water guidance value was iron. Iron is naturally occurring and was present in the water sample collected upstream of the Site at a concentration of 0.39 mg/L. Although the groundwater to surface water pathway for metals is not likely significant, it was nonetheless evaluated in the Baseline Risk Assessment (see Section 6.0).

The only Site-related chemical for which the groundwater discharge to Cattaraugus Creek has apparently had a measurable impact on water quality is ammonia. Ammonia was detected in Site groundwater more frequently and at higher concentrations than other Site-related chemicals. The reducing condition of Site groundwater allows ammonia to remain relatively stable and mobile. Ammonia conversion to nitrite and nitrate requires oxidizing groundwater conditions. The absence of nitrate and nitrite, combined with the negative ORP values, ferrous iron presence, and low dissolved oxygen concentrations, indicates a strongly reducing environment. Ammonia concentrations measured were as high as approximately 800 mg/L in overburden monitoring wells downgradient of the landfilled sludge fill. Ammonia was not detected in Cattaraugus Creek water samples upgradient of the Inactive Landfill Area and was detected at a maximum concentration of 0.442 mg/L downstream of the Site. The maximum detected level of 0.442 mg/L is approximately equal to the calculated surface water guidance value for ammonia of 0.44 mg/L. Although ammonia is relatively stable in anaerobic groundwater environments, once in the surface water Cattaraugus Creek, nitrification processes likely occur and ammonia will be rapidly assimilated by microorganisms and other aquatic life. Consequently, the attenuation of ammonia in surface water does not present a substantial concern for exceeding surface water guidance levels at locations farther downstream of the Site.

Total sulfide concentrations were also elevated in groundwater downgradient from the Inactive Landfill Area. Sulfide requires anaerobic conditions for chemical stability and is frequently

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found in the sodium sulfide form in non-acidic, reducing environments. Sulfide was not detected in surface water samples collected downstream from the Inactive Landfill Area.

Because ammonia and sulfide are relatively stable and mobile in Site groundwater, they behave as non-reactive tracers that can be used to assess groundwater flowpaths in the bedrock. Understanding groundwater flowpaths provides insight into the anticipated depth that Sitederived chemicals present in bedrock groundwater downgradient from the Site would be expected to migrate. Figure 5-1 shows sulfide and ammonia concentrations detected in the monitoring well cluster MW-4S(R), MW-4D(R), and MW-4D2. Each well screen is positioned progressively deeper into the groundwater flow system downgradient of the Inactive Landfill Area. As shown in the figure, the concentrations of both stable constituents decrease to barely detectable levels with depth. The rapid reduction in concentration of these stable chemical constituents indicates that groundwater flowpaths are not downward near the Creek. The upward vertical hydraulic gradient observed within the bedrock between wells MW-4D(R) and MW-4D2 and the lack of hydrochemical facies shift in deeper bedrock groundwater support this conclusion. Since groundwater flow paths are upward near the Creek, chemical constituents would not migrate beyond the Creek because the impacted portion of the bedrock groundwater flow system discharges to the Creek.

5.3 COMPLETE EXPOSURE PATHWAYS

Complete exposure pathways are discussed in the Pathways Analysis Report (PAR) and Baseline Risk Assessment (BRA). Based on the analysis of chemical fate and transport provided above, pathways through which Site COPCs could reach receptors at significant exposure point concentrations include:

- 1. Fugitive Dust Emissions from Site soils
- 2. Volatilization from Site soils
- 3. Direct soil contact (for burrowing animals) from Site soils

These exposure pathways, along with direct contact scenarios for visitors, trespassers and future workers were evaluated in the BRA summarized in Section 6.0. Exposure to chemicals in groundwater, although a highly improbable scenario based on current and anticipated Site use, was evaluated in an addendum to the Baseline Risk Assessment.

The table below summarizes the chemical constituents of potential concern (COPCs) for the Inactive Landfill Area and the Former Manufacturing Plant Area established from



investigations conducted at the Peter Cooper Gowanda Site. These results were presented in the PAR and evaluated in the BRA. The selection of the chemical constituents was based on potential human chemical exposure from migration mechanisms described above.

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Summary of HHRA COPCs

| Madia | Inactive Landfill A rea | Former Manufacturing |
|-----------------|------------------------------|--|
| | mactive Lanumi Arca | Plant Area |
| Surface Soil | Metals: | VOCs: |
| | Arsenic | Carbon tetrachloride, Chloroform, Tetrachloroethene, |
| | | SVOC- |
| | } | SVULS: Renze(a)anthrosona Renze(a)aurona |
| | | Benzo(a)anuracene, Benzo(a)pyrene, |
| | | Benzo(a b)anthracene |
| Subcurface Soil | Motols | VOCs |
| Subsurface Son | Arsenic Chromium III | Carbon tetrachloride Chloroform Tetrachloroethene |
| | Augeme, enconnum m | Carbon teraemonae, emotoroni, redaemoroemene |
| | | SVOCs: |
| | | Benzo(a)anthracene, Benzo(a)pyrene, |
| | | Benzo(b)fluoranthene, Indeno(1,2,3-cd)pyrene, |
| | | Benzo(a,h)anthracene |
| | | Matala |
| | | Metals: |
| Ourshunder | VOCa | |
| Overburgen | Paragana Chlorobanzana | VUCS: Tetrachloroothene Trichloroothene |
| Groundwater | Benzene, Chiorobenzene | retrachioroethene, inchioroethene |
| | SVOCs: | Metals: |
| | 4-Methyl phenol | Iron, Manganese |
| | | |
| | Metals: | • |
| | Arsenic | |
| Dedecel | Matalas | NOC |
| Bedrock | Metals: | VULS: |
| Oroundwater | Arsenic, iron | Benzene |
| | | Metals: |
| | | Manganese |
| Landfill Seeps | Metals: | |
| , i | Iron, Arsenic | |
| Landfill Gas | VOCs: | |
| | Acetone, Carbon disulfide, 2 | -Butanone, Benzene, 4-Methyl-2-pentanone, Toluene, |
| | 1,4-Dichlorobenzene, hydrog | gen sulfide |
| Cattaraugus | None | · · |
| Creek | | |
| Surface Water | | |
| Cattaraugus | Metals: | |
| Creek | Arsenic | |
| Sediments | | |
| Wetland | Metals: | |
| Sediments | Arsenic | |

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6.0 SUMMARY OF BASELINE RISK ASSESSMENT

Remedial investigation data were used to prepare a baseline risk assessment (BRA) for the Site. The BRA evaluated the potential human health and ecological risks as a result of potential exposure to chemicals in soil, groundwater, and landfill gas at the Peter Cooper Site and in sediment and seep/surface water at Cattaraugus Creek. The BRA was submitted to the USEPA in November 2002 and included a human health risk assessment (HHRA) prepared by Geomatrix and an Ecological Risk Assessment (ERA) prepared by Vanasse Hanglin Brustlin, Inc. The risk assessments provide a conservative estimate of the nature and extent of the potential cancer and noncancer human health risks and potential ecological risks from chemicals in Site media.

The results of the HHRA indicate the following:

- For adult and adolescent trespassers at the landfill, the HIs (0.1 and 0.2, respectively) and carcinogenic risk estimates (2x10⁻⁵ and 1x10⁻⁵, respectively) are below and within the acceptable risk levels.
- For adult and adolescent trespassers at the FMPA, the HIs (0.06 and 0.2, respectively) and carcinogenic risk estimates (2x10⁻⁵ and 1x10⁻⁵, respectively) are below and within the acceptable risk levels.
- For the outdoor park worker at the landfill, the HI (4) and carcinogenic risk estimate (4x10⁻⁴) exceed the acceptable risk levels. However, the risk is primarily attributed to the unlikely pathway associated with ingestion of groundwater underlying the Site, with arsenic in groundwater accounting for the majority of the risk. In the very likely event that ingestion of groundwater is not a complete pathway, the HI (1) and carcinogenic risk estimate of 8x10⁻⁵ are at or within the acceptable risk levels.
- For the outdoor industrial worker at the FMPA, the HI (4) and carcinogenic risk estimate $(4x10^{-4})$ exceed the acceptable risk levels. The primary chemical contributing the most to the risk is arsenic in groundwater. Again, however, the risk is primarily attributed to the unlikely pathway associated with ingestion of groundwater underlying the Site, with arsenic in groundwater accounting for the majority of the risk. In the event that ingestion of groundwater is not a complete pathway, the HI (1) and carcinogenic risk estimate of $9x10^{-5}$ are at or within the acceptable risk levels.
- For the indoor commercial worker at the FMPA, the HI (0.6) and carcinogenic risk estimate $(5x10^{-6})$ are below and within the acceptable risk levels.
- For the construction worker at the landfill, the HI (3) exceeds the acceptable level while the carcinogenic risk estimate $(6x10^{-6})$ is within the acceptable risk range.

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Arsenic in soil is the primary chemical contributing to the HI. At the FMPA, the HI (4) exceeds the acceptable level while the carcinogenic risk estimate $(5x10^{-6})$ is within the acceptable risk range. Chloroform in soil is the primary chemical contributing to the HI. The analytical results indicate that elevated concentrations of chloroform were limited to one location within the FMPA. Potential exposures likely are overestimated. Exposure for the construction worker was related to specific conditions during potential construction over a continuous one-year period. No construction is currently occurring. Appropriate health and safety precautions can be taken to protect workers during future construction, thereby mitigating any potential exposures and health risk. Under the CT scenario, the total HIs at the inactive landfill and FMPA are both 1.

- For the recreational users at the landfill, adult, adolescent, and child, HIs (0.3, 0.6, and 1, respectively) are at or below the acceptable level. The theoretical excess cancer risks to adult, adolescent, and child recreational users $(4x10^{-5}, 3x10^{-5}, and 3x10^{-5}, respectively)$ are within the acceptable risk range.
- For the recreational user at the FMPA, adult, adolescent, and child, HIs (0.2, 0.6, and 1, respectively) are at or below the acceptable level. The theoretical excess cancer risks to adult, adolescent, and child recreational users $(4x10^{-5}, 4x10^{-5}, and 3x10^{-5}, respectively)$ are within the acceptable risk range.
- The estimated theoretical lifetime excess cancer risks and potential noncancer hazard quotients and HIs associated with exposure to the COPCs in soil and groundwater by a future hypothetical resident (adults and children) exceeds acceptable risk levels established by the USEPA. The assessment of human health risk under this scenario was evaluated after the RI Work Plan was reviewed and approved by the USEPA and is provided for informational purposes. Based on historic and current property uses, existing conditions, surrounding land uses and zoning, no residential use of the Site is anticipated in the future.

The results of the ERA indicate the following:

- The estimated theoretical lifetime excess cancer risks the results of the ecological risk assessment for the Peter Cooper Landfill Site indicate no potential ecological risks from organic chemicals of potential ecological concern (COPECs) to fish, terrestrial plants, wetland plants, benthic invertebrates, terrestrial invertebrates, birds, and mink. With limited exception, benthic organisms and fish in Cattaraugus Creek also show no potential ecological risks from inorganic COPECs in creek sediment and surface water, and where potential risks were modeled the associated chemical was present in upstream samples at similar concentration as downstream samples.
- The toxicological food web model used in this assessment suggests that potential ecological risks may result from exposure to organic chemicals (particularly polynuclear aromatic hydrocarbons, or PAHs) for terrestrial mammalian species.

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The model similarly suggests potential risks to several measurement endpoint terrestrial biota from one or more of the inorganic chemicals.

• Site re-development plans expressed by the Village of Gowanda, while not final or fully established, will have a substantial negative impact on the ecology of the site. Wildlife and plant species will be displaced as a result of construction equipment use, disruption of site topography and vegetative cover during clearing and regrading, and ongoing human activities. Buildings and parking facilities will prevent re-establishment of vegetative cover for foraging, nesting and burrow. Continued human use of the site following redevelopment will further limit repopulation by terrestrial biota. As such, redevelopment can reasonably be expected to cause substantially more harm to the site wildlife community than would individual exposure to chemical constituents detected on the property.

In summary, under the assumptions and conditions presented in this HHRA, the estimated HI and theoretical excess cancer risk are generally below or within the acceptable levels of concern. In those limited instances where the estimated HI and/or theoretical excess cancer risk are outside acceptable levels, the exceedance is attributable to the hypothetical assumption that future groundwater consumption is a complete pathway. Groundwater in the State of New York is classified as "GA", potential potable water supply, unless it has been designated as saline. Groundwater at the Site is not used as a potable water supply and is not likely to be used as such in the future. A municipal potable water supply is available and used by all existing residences and businesses on Palmer Street. Future use of an on-site groundwater pumping well as a potable water source would be unlikely due to inherent hydrogeologic limitations. If the assumptions and/or conditions change, the results of this HHRA may need to be re-evaluated.



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TABLES





TABLE 2-1

GROUNDWATER MONITORING WELL, PIEZOMETER, AND GAS MONITORING WELL CONSTRUCTION DETAILS

| | Gowanda, New York | | | | | | | | | | | |
|---------------------|-----------------------|---------------|---------|-------------|-----------|-------------|-----------|-----------------|-----------------------|--|--|--|
| | | Surface | Top of | Total | | Screened | Interval | | | | | |
| | Installed | Elevation (1) | Riser | Depth of | Depth of | Elevation | Depth | Depth to | Formation | | | |
| Well I.D. | By: | (famsl) | (famsl) | Boring (ft) | Well (ft) | (famsi) | (fbgs) | Bedrock (fbgs) | Screened | | | |
| Alluvial Deposits | | | | | | | | | | | | |
| MW-ISR | O'Brien & Gere Dec-87 | 778.1 | 779.62 | 10.5 | 10.5 | 772.6-767.6 | 5.5-10.5 | 10.0 | sand, gravel | | | |
| MW-3(R) | Geomatrix Jul-00 | 768.1 | 770.70 | 9.2 | 9.0 | 763.6-759.1 | 4.5-9.0 | 7.5 | silt, sand and gravel | | | |
| MW-7S | Geomatrix Sep-00 | 786.1 | 787,77 | 16.6 | 16.5 | 782.1-769.6 | 4.0-16.5 | not encountered | silt, sand and gravel | | | |
| MW-8S | Geomatrix Sep-00 | 778.1 | 777.44 | 16.0 | 16.0 | 772.1-762.1 | 6.0-16.0 | not encountered | silt, sand and gravel | | | |
| MWFP-3S | Geomatrix Oct-00 | 778.5 | 780.69 | 11.5 | 11.5 | 773.5-767.0 | 5.0-11.5 | 11.5 | silt, sand and gravel | | | |
| Fill | | | | | | | | | | | | |
| Cindery Fill | | | | | | | | | | | | |
| MW-5(S) | OBrien & Gere Aug-86 | 779.1 | 781.16 | 17.0 | 15.0 | 766.1-764.1 | 13.0-15.0 | approx. 12 | sandfill/bedrock | | | |
| MWFP-2S | Geomatrix Oct-00 | 784.3 | 785.17 | 12.0 | 12.0 | 779.3-772.3 | 5.0-12.0 | 11.0 | cindery fill | | | |
| -Sludge Fill | | | | | | | | | | | | |
| MW-2S(R) | Geomatrix Jul-00 | 768.2 | 770.93 | 8.7 | 8.5 | 763.7-759.7 | 4.5-8.5 | 7.5 | sludge/ fill | | | |
| MW-4S(R) | Geomatrix Jul-00 | 765.2 | 766.97 | 9.0 | 9.0 | 760.7-756.2 | 4.5-9.0 | 9.0 | sludge/ bedrock | | | |
| MW-6 | O'Brien & Gere Aug-86 | 781.5 | 783.58 | 18.0 | 18.0 | 768.5-763.5 | 13.0-18.0 | 18.4 | silt and sand | | | |
| Shallow Bedrock | | | | | | | | | | | | |
| MW-ID | O'Brien & Gere Aug-86 | 777.6 | 779.49 | 36.9 | 36.9 | 745.7-740.7 | 31.9-36.9 | 10.0 | shale bedrock | | | |
| MW-2D | O'Brien & Gere Aug-86 | 781.3 | 782.82 | 38.3 | 38.3 | 748.0-743.0 | 33.3-38.3 | 18.1 | shale bedrock | | | |
| MW-4D(R) | Geomatrix Jul-00 | 765 | 766.36 | 23.0 | 23.0 | 747.0-742.0 | 18.0-23.0 | 5.5 | shale bedrock | | | |
| MW-5D | Geomatrix Sep-00 | 779.3 | 781.04 | 28.9 | 28.5 | 760.8-750.8 | 18.5-28.5 | 15.0 | shale bedrock | | | |
| MW-7D | Geomatrix Sep-00 | 785.8 | 787.38 | 35.5 | 35.5 | 760.3-750.3 | 25.5-35.5 | 19.5 | shale bedrock | | | |
| MW-8D | Geomatrix Sep-00 | 778.0 | 777.64 | 45.0 | 45.0 | 743.0-733.0 | 35.0-45.0 | 21.0 | shale bedrock | | | |
| MWFP-1D | Geomatrix Oct-00 | 785.2 | 787.30 | 22.5 | 22.5 | 772.7-762.7 | 12.5-22.5 | 4.5 | shale bedrock | | | |
| MWFP-2D | Geomatrix Oct-00 | 784.1 | 786.00 | 28.0 | 28.0 | 766.1-756.1 | 18.0-28.0 | 12.5 | shale bedrock | | | |
| MWFP-3D | Geomatrix Oct-00 | 778.7 | 780.51 | 26.5 | 26.0 | 762.7-752.7 | 16.0-26.0 | 11.5 | shale bedrock | | | |
| Deep Bedrock | | | | | | | | | | | | |
| MW-4D2 | Geomatrix Sep-00 | 765.1 | 766.36 | 40.5 | 40.0 | 735.1-725.1 | 30.0-40.0 | 12.5 | shale bedrock | | | |
| Piezometers/Drive I | oints | | | | | | | | | | | |
| PZ-1 | Geomatrix Oct-00 | 770.0 | 772.31 | 14.0 | 14.0 | 766.0-756.0 | 4.0-14.0 | not encountered | silt, sand and gravel | | | |
| DP-1 | Geomatrix Oct-00 | 759.2 | 761.38 | 5.0 | 5.0 | 756.7-754.7 | 2.5-4.5 | not encountered | (2) | | | |
| Gas Probes | | | | - | | | | | | | | |
| GMW-1 | Geomatrix Oct-00 | 787.1 | 787.76 | 25.0 | 10.0 | 784.6-777.1 | 2.5-10.0 | 25.4 | sludge/ fill | | | |
| GMW-2 | Geomatrix Oct-00 | 787.1 | 789.51 | 24.0 | 6.0 | 784.6-781.1 | 2.5-6.0 | 23.9 | sludge/ fill | | | |

Peter Cooper Site Gowanda, New York

GMW-3 Notes:

Survey completed by TVGA Engineering, Surveying P.C., August 28, 2000.
 One-inch diameter drive point piezometer. Native silt, sand and gravel assumed.

788.4

790.31

.

21.2

8.0

785.4-780.4

3.0-8.0

21.2

sludge/ fill

Geomatrix Oct-00

famsl: = feet above sea level fbgs: = feet below ground surface

P:Project005771.001 Peter Cooper RIPS/tables/Table 2-1 Well Construction

TABLE 2-2



SUMMARY OF QA/QC SAMPLES

Peter Cooper Site Gowanda, New York

| Sample | Sample | Sample |
|------------------------------|-----------------|---------------------------------------|
| ID | Media | Location |
| Matrix Spike/Matrix Spike Du | plicates | |
| 081400001 | groundwater | MW-4S(R) |
| 110600086 | groundwater | MWFP-1D |
| 043001121 | groundwater | MW-8S |
| 050101128 | groundwater | MWFP-3S |
| 100500010 | soil | SB-4 |
| 10060020 | soil | TP-9 |
| 100900038 | soil | MWFP-3 |
| 101000047 | soil | WSS-1 |
| 101100068 | soil | LFSS-11 |
| 110700092 | creek sediments | SED-4 |
| Duplicates | | |
| 081400002 | groundwater | MW-3 |
| 081400003 | groundwater | Duplicate for MW-3 |
| 110700088 | groundwater | MWFP-3S |
| 110700089 | groundwater | Duplicate of MWFP-3S |
| 050101126 | groundwater | MWFP-3D |
| 050101127 | groundwater | Duplicate of MWFP-3D |
| 110700098 | surface water | SW-3 |
| 110700099 | surface water | Duplicate of SW-3 |
| 050201134 | surface water | SW-1 |
| 050201133 | surface water | Duplicate of SW-1 |
| 100600015 | soil | MWFP-2 |
| 100600016 | soil | Duplicate for MWFP-2 |
| 100900040 | soil | SB-6 |
| 100900041 | soil | Duplicate for SB-6 |
| 100900026 | soil | TP-4 |
| 100900027 | soil | Duplicate for TP-4 |
| 101000052 | soil | WSS-6 |
| 101000053 | soil | Duplicate for WSS-6 |
| 101100069 | soil | LFSS-6 |
| 101100063 | l soil | Duplicate for LFSS-6 |
| 110700093 | creek sediments | SED-3 |
| 110700094 | creek sediments | Duplicate for SED-3 |
| 101200080 | landfill gas | GMW-2 |
| 101200079 | landfill gas | Duplicate for GMW-2 |
| Equipment Blanks | L | · · · · · · · · · · · · · · · · · · · |
| 100900044 | water | gas well installation |
| 110900113 | water | prior to MW-8D gw |
| 050101129 | water | prior to MWFP-2D gw |
| Trip Blanks | | |
| TB110700 | water | |
| TB110800 | Water | |
| TB110900 | Water | |
| TB111000 | Water | |
| TB043001 | Water | |
| TB050101 | Water | |
| TR050201 | water | |
| TR050201 | water | |
| | Malci | |



Page 1 of 9

TABLE 2-3

COMPARISON OF QUALITY CONTROL/QUALITY ASSURANCE SAMPLES - GAS MEDIA

Peter Cooper Site Gowanda, New York

| | Sample Location | Identification and | T |
|---------------------------------|-----------------|--------------------|------------|
| | Date Ce | liected ' | ļ |
| | GMW-2 | GMW-2 Dun | Relative |
| | 101200080 | 101200079 | Percent |
| Constituent | 10/12/00 | 10/12/00 | Difference |
| Field Measured Parameters | | + | 1 |
| Lower Explosive Limit & | 45 | NA | NA |
| | 0 | NA | NA NA |
| Hydrogen Sulfide Gas, pom | >1000 | NA | NA |
| Oxygen %, v/v | 21.3 | NA | NA |
| PID Measurements, ppm | 325 | NA | NA |
| Laboratory Parameters | | | |
| Fixed Gases, percent volume per | | 1 | |
| volume | | | |
| Carbon Dioxide | 0.136 | 9.9 | 194.6 |
| Methane | 0.145 | 17.5 | 196.7 |
| Nitrogen | 77.6 | 57.8 | 29.2 |
| Oxygen + Argon | 22.1 | 14.8 | 39.6 |
| Volatile Organic Compounds, | , · | 1 | I – |
| micrograms per cubic meter | | <u> </u> | |
| 1,1,1-Trichloroethane | 20 | 25 U | NA |
| 1,1,2,2-Tetrachloroethane | 20 | 25 U | NA |
| 1,1,2-Trichloroethane | 20 | 25 U | <u>NA</u> |
| 1,1-Dichloroethane | 20 | 25 U | |
| 1,1-Dichloroethene | 20 | 25 U | NA |
| 1.2 Disblanck engage | 20 | 25 0 | |
| 1.2 Dichloropenzene | 20 | 25 U | |
| 1.2-Dichloropropage | 20 | 25 U | NA NA |
| 1.3-Dichlorobenzene | 20 | 25 11 | NA NA |
| 1.4-Dichlorobenzeue | 20 | 25 U | NA NA |
| 2-Butanone | 43 | 25 U | NA |
| 2-Hexanone | .7 | 25 U | NA |
| 4-Methyl-2-pentanone | 3.4 | 25 U | NA |
| Acetone | 150 | 25 U | NA |
| Benzene | 2 U | 73 | NA |
| Bromodichioromethane | 2 U | 25 U | NA |
| Bromoform | 2 U | 25 U | NA |
| Bromomethane | 2 U | 25 U | NA |
| Carbon Disulfide | 93 | 3.000 | 188.0 |
| Carbon Tetrachloride | 2 U | 25 U | NA |
| Chlorobenzene | 2 U | 25 U | NA |
| Chloroethane | 2 U | 25 U | NA |
| Chloroform | 2 U | 25 U | NA |
| Chloromethane | 20 | 25 U | NA |
| cis-1,2-Dichloroethene | 20 | 25 U | NA |
| cis-1.3-Dichloropropene | 2 U | 25 U | NA |
| Dibromochloromethane | 20 | 25 U | NA |
| Einyidenzene | 3.5 | 100 | 186.5 |
| m- & p-Ayenes | 3.5 | 35 | 103.0 |
| Methylana ablanda | 20 | 25 U | NA |
| Nemylene enionde | 20 | 25 U | NA NA |
| C-Ayiciic | 1.4 IK | 25 U | NA NA |
| Terrachlornethene | 20 | 25 U | NA NA |
| Toluene | <u></u> | 820 | 181.2 |
| trans. 1.2. Dichloreethene | 211 | 25 11 | NA |
| trans_1.2-Dichloropropene | 2.0 | 25 0 | NA |
| Trichloroethene | 2.0 | 25 0 | NA NA |
| Trichlorofluoromethane | 17 TP | 25 U | NA - |
| Trichlorotrifluoroethane | 211 | 25 11 | NA NA |
| Vinyl Acetate | 14 | 25 0 | NA NA |
| Vinyl Chloride | 2 U | 25 U | NA |

Notes:

Sample locations shown on Plate 1.
 Qualifications reflect the 100% data validation performed by Data Validation Services.

NA = Not applicable; RPD cannot be calculated when analyte is qualified with a U or UJ.

- ppm = parts per million. %, v/v = percent volume per volume
- TR = trace value

U = none detected at or above the listed detection limit.

TABLE 2-3

COMPARISON OF QUALITY CONTROL/QUALITY ASSURANCE SAMPLES - SOIL MEDIA

Peter Cooper Site Gowanda, New York

| | | | | Sample Location, | Identification, and Dat | e | <u> </u> | | |
|---------------------------------|-----------|-----------|------------|------------------|-------------------------|------------|------------|------------|------------|
| | TP-4 | TP-4 DUP | | LFSS-6 | LFSS-6 DUP | | | | |
| | 10090026 | 100900027 | Relative | 101100069 | 101100063 | Relative | WSS-6 | WSS-6 DUP | Relative |
| | 7' | 7' | Percent | 0-6 in. bgs | 0.5-2.5' | Percent | 101000052 | 101000053 | Percent |
| Constituent ² | 10/9/2000 | 10/9/2000 | Difference | 10/11/2000 | 10/06/00 | Difference | 10/10/2000 | 10/10/2000 | Difference |
| Volatile Organic Compounds, | | | 1 | | | I | N | | |
| milligrams per kilogram | · | | | | | | | | |
| 1,2-Dichlorobenzene | 0.011 UJ | 0.01 UJ | NA | (0.015) U J R | 0.017 UJ | NA | 0.018 UJ | 0.0045 J | NA |
| 1,4-Dichlorobenzene | 0.011 UJ | 0.01 UJ | NA | (0.015) U J R | 0.017 UJ | NA | 0.018 UJ | 0.16 UJ | NA |
| Benzene | 0.0025 J | 0.0014 UJ | NA | 0.0029 J | 0.017 UJ | NA | 0.004 J | 0.16 UJ | NA |
| Chlorobenzene | 0.011 UJ | 0.01 UJ | NA | 0.015 U J | 0.017 UJ | NA | 0.018 UJ | 0.16 UJ | NA |
| Ethylbenzene | 0.011 UJ | 0.01 UJ | NA | 0.015 U J | 0.017 UJ | NA | 0.018 UJ | 0.16 UJ | NA |
| m-/p-Xylene | 0.0036 J | 0.0017 UJ | NA | 0.015 U J | 0.017 UJ | NA | 0.0053 J | 0.0096 J | 57.7 |
| o-Xylene | 0.011 UJ | 0.01 UJ | NA | 0.015 U J | 0.017 UJ | NA | 0.018 U | 0.0062 J | NA |
| Toluene | 0.0054 J | 0.0032 UJ | NA | 0.015 U J | 0.017 UJ | NA | 0.0082 J | 0.002 J | 121.6 |
| Metals, milligrams per kilogram | | | | | | | | | |
| Arsenic | 4.3 | 5.9 | 31.4 | 919 | 1140 | 21.5 | 15.7 | 9.8 | 46.3 |
| Chromium | 10.3 J | 11.4 J | 10.1 | 341 | 368 | 7.6 | 45.7 | 37.9 | 18.7 |
| Hexavalent Chromium | 4.78 U | 4.74 U | NA | 5.17 U | 5.08 U | NA | 5.87 U | 5.8 U | NA |
| Zinc | 57.3 | 63.1 | 9.6 | 165 | 230 | 32.9 | 136 | 82.6 | 48.9 |
| Others | | | | | | | | | |
| Percent Solids | 83.7 | 84.3 | 0.7 | 77.3 | 78.7 UJ | NA | 68.2 | 69 | 1.2 |
| pН | 8.35 | 8.44 | 1.1 | 5 | 6.82 UJ | NA | 7.74 | 7.74 | 0.0 |
| Total Organic Carbon | 0.1 UJ | 0.62 | NA | 6.61 | 4.9 J | 2.0 | 2.70 | 2.8 | 3.6 |

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

NA = Not applicable: RPD cannot be calculated when analyte is qualified with a U or UJ.

UJ = indicates compound was not detected above the listed detection limit. However, the reported quantitation limit is approximate and may or may

not represent the actual limit of quantitiation necessary to accurately

and precisely measure the compound in the sample.

J = indicates an estimated value.

U = indicates compound was not detected at or above the listed detection limit.

D = indicates spike diluted out.

E = indicates compound concentrations exceed calibration range.







TABLE 2-3

COMPARISON OF QUALITY CONTROL/QUALITY ASSURANCE SAMPLES - SOIL MEDIA

Peter Cooper Site Gowanda, New York

| | | | | Sample Location. | Identification, and Date | 7 | | | |
|--|--|--|-----------------------------------|---|---|-----------------------------------|---|---|-----------------------------------|
| Constituent ² | SB-6 100900040 4-6' 10/9/2000 | SB-6 DUP 100900041 4-6' 10/9/2000 | Relative Percent Difference | MWFP-2 100600015 0.5-2.5' 10/06/00 | MWFP-2 DUP 100600016 0.5-2.5' 10/06/00 | Relative Percent Difference | Creek Sed. #3 110700093 0-3 inches 11/7/2000 | Creek Sed. #3 DUP 110700094 0-3 inches 11/7/2000 | Relative Percent Difference |
| Volatile Organic Compounds, | | | | | | | | | |
| milligrams per kilogram | | | | | | | | | |
| 1,1,1 Trichloroethane | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| 1,1,2,2-Tetrachloroethane | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| 1,1,2-Trichloro-1,2,2-Tricfluoroethane | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| 1.1.2-Trichloroethane | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| 1.1-Dichloroethane | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 U) | NA | 0.011 U | 0.013 U | NA |
| 1,1-Dichloroethene | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| 1,2,4-Trichlorobenzene | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| 1,2-Dibromo-3-Chloropropane | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UI | NA | 0.011 U | 0.013 U | NA |
| 1,2-Dibromoethane | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| 1,2-Dichlorobenzene | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| 1,2-Dichloroethane | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| 1.2-Dichloropropane | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA NA | 0.011 U | 0.013 U | NA |
| 1.3-Dichlorobenzene | 0.013 00 | 0.013 03 | <u>NA</u> | 0.022 01 | 0.016 UJ | NA | 0.011 U | 0.013 0 | NA |
| 1.4-Dichlorobenzene | 0.013 UJ | 0.013 01 | NA NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| 2-Butanone (MEK) | 0.026 J | 0.019 1 | | 0.022 UJ | 0.094 J | NA | 0.011 0 | 0.013 0 | <u>NA</u> |
| 2-Hexanone | 0.013 01 | 0.013 03 | NA NA | 0.022 01 | 0.016 UJ | NA | 0.011 U | 0.013 0 | NA |
| 4-Methyl-2-Pentanone | 0.013 01 | 0.013 01 | NA | 0.022 01 | 0.016 01 | NA | 0.011 0 | 0.013 0 | |
| Acetone | 0.14 J | 0.093 J | 40.3 | 0.030 J | 0.004 J | (13.3 | 0.019 | 0.020 | 31.1 |
| Denzene | 0.013 03 | 0.013 07 | NA NA | 0.0076 J | 0.0039 J | 04.3 NA | 0.0015 J | 0.0018 J | 18.2 NA |
| Bromodichioromemane | 0.013 UI | 0.013 03 | NA NA | 0.022 01 | 0.016 UI | | 0.011 U | 0.013 U | NA |
| Bromomethume | 0.013 10 | 0.013 01 | NA NA | 0.022 01 | 0.016 UI | NA | 0.011 U | 0.013 U | NA |
| Cashen Disulfide | 0.013 03 | 0.013 07 | 11.8 | 0.022 03 | 0.016 01 | 177.3 | 0.011 0 | 0.013 0 | 111 |
| Carbon Tetrachloride | 0.013 111 | 0.027 5 | NA | 0.077 1/1 | 0.0431 | NA | 0.019 | 0.017 | NA NA |
| Chlorobeuzene | 0.013 111 | 0.013 UI | NA NA | 0.022 01 | 0.016 UI | NA NA | 0.011 U | 0.013.0 | NA |
| Chloroethaue | 0.013 10 | 0.013 UJ | NA | 0.022 UI | 0.016 []] | NA | 0.011 U | 0.013 U | NA |
| Chloroform | 0.013 UI | 0.013 UJ | NA | 0.022 UI | 0.016 ()) | NA | 0.011 U | 0.013 U | NA |
| Chloromethane | 0.013 UJ | 0.013 UJ | NA | 0.022 UI | 0.016 111 | NA | 0.011 U | 0.013 11 | NA |
| cis-1.2-Dichloroethene | 0.013 UJ | 0.013 UJ | NA | 0.022 UI | 0.016 UI | NA | 0.011.11 | 0.013 U | NA |
| cis-1.3-Dichloropropene | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 10 | NA | 0.011 U | 0.013 U | NA |
| Cyclohexane | 0.013 UJ | 0.013 UJ | NA | 0.0095 J | 0.0098 J | 3.1 | 0.0022 J | 0.0026 J | 16.7 |
| Dibromochloromethane | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UI | NA | 0.011 1/ | 0.013 U | NA |
| Dichlorodifluoromethane | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| Ethylbenzene | 0.013 UJ | 0.013 UI | NA | 0.022 UI | 0.016 UI | NA | 0.011 U | 0.013 U | NA |
| Isopropylbenzene | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 J | NA | 0.011 U | 0.013 U | NA |
| in-/p-Xylene | 0.013 UJ | 0.013 UJ | NA | 0.0071 J | 0.0042 J | 51.3 | 0.0015 J | 0.0018 J | 18.2 |
| Methyl Acetate | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| Methyl tert-Butyl Ether | 0.013 UJ | 0.013 UJ | ŇA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| Methylcyclohexane | 0.0082 J | 0.0056 J | 37.7 | 0.015 J | 0.011 J | 30.8 | 0.0033 J | 0.004 J | 19.2 |

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TABLE 2-3

COMPARISON OF QUALITY CONTROL/QUALITY ASSURANCE SAMPLES - SOIL MEDIA

Peter Cooper Site Gowanda, New York

| | Sample Location, Identification, and Date' | | | | | | | | |
|----------------------------------|--|-----------|------------|-----------|------------|------------|----------------|-------------------|------------|
| | SB-6 | SB-6 DUP | | MWFP-2 | MWFP-2 DUP | | Creek Sed. #3 | Creek Sed. #3 DUP | |
| | 100900040 | 100900041 | Relative | 100600015 | 100600016 | Relative | 110700093 | 110700094 | Relative |
| | 4-6' | 4-6' | Percent | 0.5-2.5' | 0.5-2.5' | Percent | 0-3 inches | 0-3 inches | Percent |
| Constituent ² | 10/9/2000 | 10/9/2000 | Difference | 10/06/00 | 10/06/00 | Difference | 11/7/2000 | 11/7/2000 | Difference |
| Methylene Chloride | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| o-Xylene | 0.013 UJ | 0.013 UJ | NA | 0.0039 J | 0.0026 J | 40.0 | 0.011 U | 0.013 U | NA |
| Styrene | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | <u>0.011 U</u> | 0.013 U | NA |
| Tetrachloroethene | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| Toluene | 0.09 J | 0.047 J | 62.8 | 0.015 J | 0.0061 J | 84.4 | 0.0045 J | 0.0047 J | 4.3 |
| trans-1,2-Dichloroethene | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| trans-1,3-Dichloropropene | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 U | NA _ | 0.011 U | 0.013 U | NA |
| Trichloroethene | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.0036 J | ŇA | 0.011 U | 0.013 U | NA |
| Trichlorofluoromethane | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| Vinyl Chloride | 0.013 UJ | 0.013 UJ | NA | 0.022 UJ | 0.016 UJ | NA | 0.011 U | 0.013 U | NA |
| Semi-Volatile Organic Compounds, | | | 1 1 | | | | | | |
| miligrams per kilogram | | | | | | | | | |
| Acenaphthene | 0.4 U | 0.41 U | NA NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 1,1-Biphenyl | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 2,2-oxybis(1-chloropropane) | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 2.4.5-Trichlorophenol | 10 | 10 | NA | 1.1 U | 0.99 U | NA | IU | <u> </u> | NA |
| 2.4.6-Trichlorophenol | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 2.4-Dichlorophenol | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 2,4-Dimethylphenol | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA _ | 0.4 U | 0.4 U | NA |
| 2,4-Dinitrophenol | 10 | 10 | NA | 1.1 U | 0.99 U | NA | 10 | IU | NA |
| 2,4-Dinitrotoluene | 0.4 U | 0.41 Ü | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 2.6-Dinitrotoluene | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 2-Chloronaphthalene | 0.4 U | 0.41 1 | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 2-Chlorophenol | 0:4 U | 0.41 Ū | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 2-Methylnaphthalene | 0.043 J | 0.042 J | 2.4 | 0.083 J | 0.082 J | 1.2 | 0.4 U | 0.4 U | NA |
| 2-Methylphenol | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 2-Nitroaniline | <u> </u> | 10 | NA | 1.1 U | 0.99 U | NA | 10 | 10 | NA |
| 2-Nitrophenol | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 3.3-Dichlorobenzidine | R | R | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 3-Nitroaniline | R | R | NA | R | R | NA | 10 | Î U | NA |
| 4.6-Dinitro-2-Methylphenol | IU | 10 | NA | 1.I U | 0.99 U | NA | 10 | 10 | NA |
| 4-Bromophenyl-Phenylether | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 4-Chloro-3-Methylphenol | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 4-Chloroaniline | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 4-Chlorophenyl-Phenylether | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 4-Methylphenol | 0.47 | 0.42 | 11.2 | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| 4-Nitroaniline | 10 | 10 | NA | 1.1 U | 0.99 U | NA | 10 | 1 U | NA |
| 4-Nitrophenol | 10 | 10 | NA | 1.1 U | 0.99 U | NA | IU | 1 U | NA |
| Acenaphthylene | 0.4 U | 0.41 U | NA | 0.29 J | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Acetophenone | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Anthracene | 0.4 U | 0.41 U | NA | 0.24 J | 0.083 J | 97.2 | 0.4 U | 0.4 U | NA |
| Atrazine | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |

1 Projuct005771 PRP (In up Pour Cooper NPL RI report FINAL REPORT (Normaliser 2003 Submitted)/Tables (Final//Tables 2-3 QA QC comparisons FINAL

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TABLE 2-3

COMPARISON OF QUALITY CONTROL/QUALITY ASSURANCE SAMPLES - SOIL MEDIA

Peter Cooper Site Gowanda, New York

| | Sample Location, Identification, and Date ' | | | | | | | | |
|---------------------------------|---|-----------|-------------|-----------|---------------|------------|---------------|-------------------|------------|
| | SB-6 | SB-6 LUP | | MWFP-2 | MWFP-2 DUP | | Creek Sed. #3 | Creek Sed. #3 DUP | |
| | 100900040 | 100900041 | Relative | 100600015 | 100600016 | Relative | 110700093 | 110700094 | Relative |
| | 4-6' | 4-6' | Percent | 0.5-2.5' | 0.5-2.5' | Percent | 0-3 inches | 0-3 inches | Percent |
| Constituent ² | 10/9/2000 | 10/9/2000 | Difference | 10/06/00 | 10/06/00 | Difference | 11/7/2000 | 11/7/2000 | Difference |
| Benzaldehyde | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 UJ | 0.4 UJ | NA |
| Benzo(a)anthracene | 0.4 U | 0.41 U | NA | 0.47 | 0.23 J | 68.6 | 0.4 U | 0.4 UJ | NA |
| Benzo(a)pyrene | 0.4 U | 0.41 U | NA | 0.46 | 0.23 J | 66.7 | 0.4 U | 0.4 U | NA |
| Benzo(b)fluoranthene | 0.4 U | 0.41 U | NA | 0.3 J | 0.18 J | 50.0 | 0.4 U | 0.4 UJ | NA |
| Benzo(g,h,i)perylene | 0.4 U | 0.41 U | NA | 0.33 J | 0.17 J | 64.0 | 0.4 U | 0.4 UJ | NA |
| Benzo(k)fluoranthene | 0.4 U | 0.41 U | NA | 0.38 J | <u>0.14 J</u> | 92.3 | 0.4 U | 0.4 UJ | NA |
| bis(2-chloroethoxy)methane | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| bis(2-chloroethyl)ether | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| bis(2-Ethylhexyl)phthalate | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Butyl Benzyl Phthalate | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Caprolactam | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Carbazole | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | <u>0.4 U</u> | NA |
| Chrysene | 0.058 J | 0.41 U | NA | 0.6 | 0.26 J | 79.1 | 0.4 U | 0.4 U | NA |
| Dibenzo(a,h)anthracene | 0.4 U | 0.41 U | NA | 0.13 J | .0.051 J | 87.3 | 0.4 U | 0.4 UJ | NA |
| Dibenzofuran | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Diethylphthalate | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Dimethyl Phthalate | 0.4 U | 0.41 J | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| di-N-Butylphthalate | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| di-n-Octyl Phthalate | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 UJ | NA |
| Fluoranthene | 0.073 J | 0.072 J | 1.4 | 0.62 | 0.31 J | 66.7 | 0.4 U | 0.4 U | NA 👝 |
| Fluorene | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Hexachlorobenzene | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Hexachlorobutadiene | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Hexachlorocyclopentadiene | 2 U | 2 U | NA | 0.42 U | 0.39 U | NA | 0.4 UJ | 0.4 UJ | NA |
| Hexachloroethane | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Indeno(1.2,3-cd)pyrene | 0.4 U | 0.41 U | NA | 0.27 J | 0.12 J | 76.9 | 0.4 U | 0.4 UJ | NA |
| Isophorone | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Naphthalene | 0.37 J | 0.32 J | 14.5 | 0.044 J | 0.063 J | 35.5 | 0.4 U | 0.4 U | NA |
| Nitrobenzene | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| n-Nitroso-di-n-Propylamine | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NĂ |
| n-Nitrosodiphenylamine | 0.4 U | 0.41 U | NA | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Pentachlorophenol | 10 | 10 | NA | 1.1 0 | 0.99 U | ŇĂ | 10 | 10 | NA |
| Phenanthrene | 0.088 J | 0.073 J | 18.6 | 0.3 J | 0.14 J | 72.7 | 0.4 U | 0.4 U | NA |
| Phenol | 0.36 J | 0.27 J | 28.6 | 0.42 U | 0.39 U | NA | 0.4 U | 0.4 U | NA |
| Pyrene | 0.072 J | 0.058 J | 21.5 | 0.86 | 0.41 | 70.9 | 0.4 U | 0.4 U | NA |
| Metals, milligrams per kilogram | | | | | | | | | |
| Aluminum | 6310 | 7230 | 13.6 | 6490 | 6420 | 1.1 | 5730 | 5150 | 10.7 |
| Antimony | 7.1 UJ | 7.2 UJ | NA | 7.6 UJ | 7.1 UJ | NA | 7 UJ | 7.01 UJ | NA |
| Arsenic | 6.1 | 6.2 | 1.6 | 29.9 | 57.9 | 63.8 | 7.1 J | 6.3 J | 11.9 |
| Bariun | 54.8 | 61.3 | <u>Í1.2</u> | 64.2 | 99.3 | 42.9 | 38.6 | 34.8 | 10.4 |
| Beryllium | 0.59 UJ | 0.6 U | NA . | 0.87 | 1.2 | 31.9 | 0.58 U | 0.58 U | NA |
| Cadmium | 0.59 UJ | 0.6 U | NA | 0.64 U | 0.59 U | NA | 0.58 U | 0.58 U | NA |

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TABLE 2-3

COMPARISON OF QUALITY CONTROL/QUALITY ASSURANCE SAMPLES - SOIL MEDIA

Peter Cooper Site Gowanda, New York

| | Sample Location, Identification, and Date' | | | | | | | | | | |
|--------------------------|--|-----------------------|------------|---------------------|-------------------------|------------|----------------------------|--------------------------------|------------|--|--|
| | SB-6 100900040 | SB-6 DUP 100900041 | Relative | MWFP-2 100600015 | MWFP-2 DUP 100600016 | Relative | Creek Sed. #3 110700093 | Creek Sed. #3 DUP 110700094 | Relative | | |
| | 4-6' | 4-6' | Percent | 0.5-2.5' | 0.5-2.5' | Percent | 0-3 inches | 0-3 inches | Percent | | |
| Constituent ¹ | 10/9/2000 | 10/9/2000 | Difference | 10/06/00 | 10/06/00 | Difference | 11/7/2000 | 11/7/2000 | Difference | | |
| Calcium | 14200 | 24100 | 51.7 | 2490 J | 25100 J | 163.9 | 11700 | 11000 | 6.2 | | |
| Chromium | 9 | 11.5 | 24.4 | 198 J | 57.4 J | 110.1 | 7.1 | 6.3 | 11.9 | | |
| Cobult | 7.5 | 9.4 | 22.5 | 7.1 | 6.1 | 15.2 | 6.7 | 6.1 | 9.4 | | |
| Copper | 19 | 19.6 | 3.1 | 29.3 | 20.2 | 36.8 | 13.9 | 15.3 | 9.6 | | |
| Hexavalent Chromium | 4.81 U | 4.97 U | NA | 5.08 UJ | 4.77 UJ | NA | 4.85 U | 4.86 U | NA | | |
| Iron | 17600 J | 20600 J | 15.7 | 18900 | 14800 | 24.3 | 16900 | 15100 | 11.3 | | |
| Lead | 8.8 | 8.7 | 1.1 | 41 J | 28.5 J | 36.0 | 8.8 | 7.3 | 18.6 | | |
| Magnesium | 3070 | 3500 | 13.1 | 1730 | 4330 | 85.8 | 3160 | 4410 | 33.0 | | |
| Manganese | 351 | 554 | 44.9 | 160 | 306 | 62.7 | 401 | 306 | 26.9 | | |
| Mercury | 0.17 | 0.13 | 26.7 | 0.16 J | 0.06 UJ | NA | 0.06 U | 0.06 U | NA | | |
| Nickel | 16 | 21.1 | 27.5 | 17.9 | 14.8 | 19.0 | 15.5 | 13.7 | 12.3 | | |
| Potassium | 516 | 486 | 6.0 | 542 | 728 | 29.3 | 617 | 545 | 12.4 | | |
| Selenium | 2 J | 1.9 J | 5.1 | 2.7 | 2.5 | 7.7 | 0.58 U | 0.74 | NA | | |
| Silver | 1.2 U | 1.2 U | NA | 1.3 U | 1.2 U | NA | L.17 UJ | 1.2 UJ | NA | | |
| Sodium | 757 | 829 | 9.1 | 411 | 459 | 11.0 | 240 | 296 | 20.9 | | |
| Thallium | <u>1.2 U</u> | 1.2 U | NA | 1.3 U | 1.2 U | NA | 1.2 U | 1.2 U | NA | | |
| Vanadium | 12.4 | 14.7 | 17.0 | 15.3 | 14.3 | 6.8 | 12.2 | 11.6 | 5.0 | | |
| Zinc | 69.6 | 77.1 | 10.2 | 84.6 J | 64.2 J | 27.4 | 47.1 | 41.8 | 11.9 | | |
| Others | | | | | | | | | | | |
| Percent Solids, % | 83.1 | 80.5 | 3.2 | 78.7 | 83.9 | 6.4 | 82.5 | 82.3 | 0.2 | | |
| pH | 10.1 | 10 | 1.0 | 7.7 | 8.13 | 5.4 | 8.21 | 8.08 | 1.6 | | |
| Total Organic Carbon, % | 1.3 | 2 | 42.4 | 1.7 | 1.3 | 26.7 | 0.1 U | 0.1 U | NA | | |

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

NA = Not applicable; RPD cannot be calculated when analyte is qualified with a U or UJ.

Us # indicates compound was not detected above the listed detection limit.

However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitiation necessary to accurately

and precisely measure the compound in the sample.

J = indicates an estimated value.

U = indicates compound was not detected at or above the listed detection limit.

D = indicates spike diluted out.

E = indicates compound concentrations exceed calibration range.

TABLE 2-3

COMPARISON OF QUALTLY CONTROL/QUALITY ASSURANCE SAMPLES - WATER MEDIA

Peter Conper Site Gowanda, New York

| | r | | | | 1.1 | - | 6 H | | | | | | | | <u> </u> |
|--------------------------------------|----------------|-------------------|-------------|-----------------|---------------------------|---------------|-------------|-------------|------------|-------------|-------------|------------------|----------|-----------|-----------|
| | C. A.W. du | Coul Water Al DUB | Relative | Sec. A Water #1 | ample Location, Identifit | ation and Dat | e Collected | HUER IS DUR | h later | Lawrence 10 | | | | | |
| | Creek Water #/ | Creek Walt #1 DUP | Reidire | 110700000 | Lintonopo | Relative | MHFP-33 | MWFP-35 DUP | Relative | MWFP-3D | MWFP-3D DUP | Relative | MW-SD | MW-SD DUP | Relative |
| Constituent | 630201134 | 50000 | Difference | 11/7/2000 | 11/7/2000 | Difference | 11/1/0000 | 11/7/2000 | Difference | 5/10001 | 0.010112/ | Percent | 50301141 | 050301142 | Percent |
| Volatile Organic Compounds | 3///001 | 301001 | Pyperence | 11//14/0 | 11//2010 | Dyjerenar | 11///2000 | 1 | Dyjerence | 3/1/2001 | 5/1/2001 | Dyjerence | 3/3/2001 | 5,97001 | Dyjerence |
| micrograms per liter | | 4 | 1 | | 1 | 1 | | | | | | | | 1 ' | 1 |
| 1.1.1-1 richloroethane | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| F.1.2.2-Tetrachlomethane | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| 1.1.2 Trichlom-1.2.2 Trifluoroethane | NA | NA | NA | 10 U | 10 U | NA | 10 11 | 10 U | NA | NA | NA | NA | NA | NA | NA |
| 1,1,2-Tricblomethane | NA | NA | NA | 10.0 | 10 U | NA | 10 17 | 10 U | NA | NA | NA | NA | NA | NA | NA |
| 1,1-Dichloroethane | <u>NA</u> | <u>NA</u> | NA | 10 U | 10 U | NA | 20 J | 2.1 J | 4.9 | NA | NA | NA | NA | NA | NA |
| 1.1-Dichlomethene | NA | NA | NA | 10 U | 10 13 | NA NA | 10 11 | 10 U | NA | NA | <u>NA</u> | NA NA | NA | NA | NA |
| 1.2.4-Trichlorobenzene | <u>NA</u> | NA | 1 <u>NA</u> | 10 U | 10 U | NA | 10 (1 | 10 U | NA | NA | NA | NA | NA | NA | NA |
| 1.2-Dihmmo-3-chloropropane | NA | <u>NA</u> | NA NA | 10 0 | 10 0 | NA | 10 0 | 10 0 | NA | <u>NA</u> | <u>NA</u> | <u>NA</u> | NA | NA I | NA |
| 1.2-Dibromoethane | | NA IO II | - NA | 10 11 | 10 0 | 1 NA | 10 10 | 10 0 | NA | <u>NA</u> | NA NA | NA | NA | NA IN | NA |
| 1.2 Dichlonotchiene | 10 0 NA | NA | NA | 10 0 | 10 11 | | 1011 | 10.0 | NA | | <u></u> | NA NA | <u> </u> | | NA NA |
| 1.2.Dichtonomoane | NA | NA | NA NA | 10 1 | 10.11 | NA | 10 11 | 10 0 | NA | | | NA NA | NA NA | | NA |
| 1.3-Dichlombenzene | NA | NA | NA | 10 U | 10 1 | NA NA | 10 1 | 10 1 | NA | NA | NA | NA | NA | NA NA | NA |
| 1,4-Dichlorobenzene | 10 U | 10 1/ | NA | 10 U | 10 Ü | NA | 10 tJ | 10 U | NA | NA | NA | NA | 10 U | 10 U | NA |
| 2-Butanone (MEK) | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| 2-Hexanone | NA | NA | NA | 10 U | 10 Ū | NA | 10 U | 10 U | NA | ŇĂ | NA | NA | NA | NA | NA |
| 4-Methyl-2-Pentanone | NA | NA | NA | 10 U | 10 U | NA | <u> </u> | 10 U | NA | NA | NA | NA | NA | NA | NA |
| Acetone | NA | NA | NA | 10 U | 4.0 J | NA | 10 U | 10 1/ | NA | NA | NA | NA | NA | NA | NA |
| Benzene | 10 0 | 10 U | NA | 10 11 | 10 U | NA | 10 U | 10 U | NA NA | 1.2 1 | 1.5 J | 22.2 | 10 1/ | 10 U | NA |
| Bromodichloromethane | NA | NA NA | <u>NA</u> | 10 0 | 10 0 | <u> </u> | 10 0 | 10 U | NA | NA | NA | NA | NA | NA NA | <u>NA</u> |
| Bromolorm | NA | <u>NA</u> | | 10 0 | 10 0 | <u> </u> | 10 10 | 10 0 | NA NA | <u>NA</u> | NA NA | NA | NA | NA Y | <u> </u> |
| General Disulfide | <u>NA</u> | <u>NA</u> | | 10 0 | 10 0 | | 10 0 | 10 0 | <u>NA</u> | NA NA | <u>NA</u> | <u>NA</u> | NA NA | | <u> </u> |
| Carbon Tetrachloride | 1011 | 10.11 | NA | 10 10 | 1010 | | 10 0 | 10 0 | NA | 10.11 | | - <u>NA</u> | NA | | NA NA |
| (hterobenzene | 10 U | 10 U | NA | 10 (1 | 10 1 | NA | 10 11 | 10.1 | NA | NA | NA | NA | 10 11 | 10 11 | NA |
| Chloroethane | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 () | NA | NA | NA | NA | NA | NA | NA |
| Chloroform | 10 U | 10 U | NA | 10 U | 10 U | NA | 10 1/ | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA |
| Chloromethane | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| cis-1.2-Dichloroethene | NA | NA | NA | 10 U | 10 U | NA | 5.0 3 | 5.11 | 2.0 | NA | NA | NA | NA | NA | NA |
| cis-1.3-Dichloropropene | NA | NA | NA | 10 U | <u> </u> | NA | 10 U | 10 U | NA | NA | NA | NA . | NA | NA | NA |
| Cyclohexane | <u>NA</u> | <u>NA</u> | NA | 10 U | 10 U | NA | 10 U | 10.0 | NA | NA | NA | NA | NA | NA NA | <u>NA</u> |
| Dibromochloromethane | <u>NA</u> | <u> </u> | <u>NA</u> | 10 0 | 10 0 | NA | 10 0 | 10 0 | NA | <u>NA</u> | NA | NA NA | NA | NA NA | - NA |
| Dichubanaga | - 1011 | | | 10.0 | 10 11 | <u>NA</u> | 10 11 | 10 0 | NA | <u> </u> | NA NA | <u><u>NA</u></u> | 10.11 | NA | |
| itomoticatene | NA NA | NA | NA | 10 11 | 10.0 | | 101 | 10 11 | NA NA | NA NA | | NA | | NA | NA NA |
| m-to-X viene | 10 U | 10 11 | NA | 10 11 | 10 1 | NA | 10 11 | 10 01 | NA | NA NA | NA NA | NA | 10 11 | 10 U | NA |
| Methyl Acetate | NA | NA | NA | 10 U | 10 Ü | NA | 10 1/ | 10 U | NA | NA | NA | NA | NA | NA | NA |
| Methyl tert-Butyl Ether | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| Methylcyclohexane | NA | NA | NA | 10 U | 10 Ū | NA | 10 U | 10 U | NA | NA | NA | NA | ŇĂ | NA | NA |
| Methylene Chloride | NA | NA | NA | 10 U | 10 Ū | NA | 10 U | 10.0 | NA | NA | NA NA | NA | NA | NA | NA. |
| o-Xylene | 10 U | 10 0 | NA | 10 0 | 10 Ū | NA | 10 U | 10 U | NA | NA | NA | NA | 10 17 | 10 U | NA. |
| Styrene | NA | NA | NA | 10 U | 10 U | NA | 10 U | <u>10 U</u> | NA | NA | NA | NA NA | NA | NA | NA |
| Tetrachloroethene | 10 () | 10 U | NA NA | 10 U | 10 0 | NA | 5.51 | 5.6 J | 8.1 | 10 U | 10 () | NA | NA | NA | NA |
| Toluene | 10 0 | 10 0 | <u> </u> | 10 11 | 10 10 | NA | 10 10 | 10.0 | NA | <u>NA</u> | NA | NA NA | 10 0 | | <u>NA</u> |
| trans-1.2-Dichlomorphene | NA NA | <u>NA</u> | | 10.0 | 10 11 | NA | 10 0 | 10.0 | <u></u> | <u>NA</u> | NA NA | NA NA | NA NA | NA | NA |
| Inchlometheae | 10.11 | 1011 | NA NA | 10 11 | 10 11 | NA | 201 | 100 | 112 | | 10.11 | | NA | NA NA | NA |
| Trichlonfluoromethane | NA | NA NA | NA NA | 10.1 | 10 11 | NA NA | 1011 | 10 11 | NA | NA | NA | NA | NA | NA | NA |
| Vinvl Chloride | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 11 | NA | NA NA | NA NA | NA | NA | NA | NA |
| Semi-Volatile Organic Compounds, | | | | | | | | | | | | | | 1 | |
| micrograms per liter | 1 | | | | | [| | | | | (| { i | | I / | |
| I, I-Biphenyl | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA. | NA | NA | NA | NA |
| 2,2-oxyhis(1-chloropropane) | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| 2.4.5-Trichlorophenol | 26 () | 25 | NA | _25 U | 25 Ū | NA_ | 25 U | R | NA | NA | NA | NA | 24 U | 24 1) | NA |
| 2.4.6-Trichlorophenol | 10 U | 10 | NA NA | 10 U | 1010 . | NA | 10 U | R | NA | NA | NA | NA | 9.4 11 | 94 U | NA |
| 2.4-Dichlorophenol | - 10 U | 10 t) | NA | <u>10 U</u> | 10 1/ | NA | 10 U | R | NA | NA | NA | NA | 9411 | 941) | NA NA |
| 2.4-Lanchylphenol | 10 11 | 10 0 | NA NA | 10 0 | 10 U | NA | 10 0 | R | NA NA | <u>NA</u> | NA | NA | 9.4 U | 940 | - NA |
| 2.4-Sourcoperior | 20 U | 23 U | | 10 11 | 45 0 | NA | | K | NA | <u>NA</u> | NA | NA_ | 24 0 | <u></u> | NA NA |
| 2.6-Dinitmiologne | NA NA | NA NA | | 10 11 | 10 11 | - NA | 10 10 | 10 0 | | <u>NA</u> | | | NA | <u>NA</u> | NA NA |

1 Proprietabilite Part Group from Corporation and Report of Neuroscien and Submitted Alian (SubStates 1:1 QA OF International Parts).

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TABLE 2-3

COMPARISON OF QUALITY CONTROL/QUALITY ASSURANCE SAMPLES - WATER MEDIA

Peter Cooper Site Gowanda, New York

| | T | | | 5 | mole Location, Identific | ation and Dat | e Collected | | | | | | | I | |
|--------------------------------|----------------|--------------------|------------|----------------|--------------------------|---------------|-------------|-------------|-------------|--------------|--------------|------------|-------------|-----------|------------|
| | Creek Water #1 | Creek Water #1 DUP | Relative | Creek Water #3 | Creek Water #3 DUP | Relative | MWEP-15 | MWFP-35 DUP | Relative | MWEP-10 | MWEP, ID DUP | Relative | MW.SD | MW-SD DUP | Relative |
| 1 | 050201134 | 050201133 | Percent | 110700098 | 110700099 | Percent | 110700088 | 110700089 | Percent | 050101126 | 950101127 | Percent | 50101141 | 050301142 | Percent |
| Constituent ² | 5/2/2001 | 5/2/2001 | Difference | 11/7/2000 | 11/7/2000 | Difference | 11/7/2000 | 11/7/2000 | Difference | 5/1/2001 | \$1/2001 | Difference | \$117001 | \$/3/2001 | Difference |
| 2.C blorogaphthalene | NA | NA | NA | 10 11 | 10.1/ | NA | 10.11 | 10.11 | NA | NA | NA | NA | NA | NA | NA |
| 2-Chioropheuol | 10 U | 10 U | NA | 10.17 | 10 U | NA | 10 0 | R | NA | NA | NA | NA | 941 | 9.4 11 | NA |
| 2-Methylnaphthalene | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| 2-Methylphenol | 10 17 | 10 U | NA | 10 U | 10 U | NA | 10 U | R | NA | NA | NA | NA | 9.4 U | 9.4 U | NA |
| 2-Nitmaniline | NA | NA | NA | 25 U | 25 U | NA | 25 U | 25 U | NA | NA | NA | NA | NA | NA | NA |
| 2-Nitrophenol | 10 U | 10 0 | <u> NA</u> | 10 0 | 10 0 | NA | 10 1) | R | NA | NA | NA | NA | 9.4 U | 9.4 U | NA. |
| 1.3-Dichlombenzidine | NA | NA | <u>NA</u> | 10 0 | 10 0 | NA | 10 0 | 10 0 | NA NA | NA | NA | <u>NA</u> | NA | NA | <u>NA</u> |
| 1-Naroanine | 26.11 | | NA NA | 25.0 | 25 U | | 25 0 | <u></u> | - <u>NA</u> | | NA NA | NA | .NA | NA | <u>NA</u> |
| 4-Bromothenyl-Phenylether | NÁ NÁ | NA NA | NA | 10 1 | 10 11 | NA | 1011 | 1011 | NA | NA | NA NA | NA NA | NA | NA NA | NA NA |
| 4. Chloro-3-Methylphenol | 10 U | 10 U | NA | 10 U | 10 U | NA | 10 U | R | NA | NA | NA NA | NA | 940 | 9.4 U | NA |
| 4-Chloroaniline | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| 4-Chlorophenyl-Phenylether | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA _ | NA | NA | NA | NA | NA | NA |
| 4-Methylphenol | 10 (1 | 10 U | NA | 10 U | 10 U | NA | 10 U | R | NA | NA | NA | NA | 9.4 U | 9.4 U | NA |
| 4-Nitmaniline | <u>NA</u> | NA NA | NA | 25 U | 25 U | NA | 25 U | 25 U | NA | NA | NA | NA | NA | NA | NA |
| | 26 0 | 25 U | NA NA | | - 25 U | NA NA | 25 0 | R | | <u>NA</u> | NA NA | NA NA | 24 0 | 24 U | NA |
| Acenaphihene | <u>NA</u> | <u>NA</u> | | 10 0 | 10 0 | | 10 0 | | <u>NA</u> | NA NA | NA NA | NA | <u>NA</u> | NA NA | NA NA |
| Acetonhenone | NA | NA | NA | 10 0 | 10 U | NA | 10 11 | 10 0 | 1 NA | NA - | NA NA | NA | NA | NA NA | NA NA |
| Anthracene | NA | NA | NA | 10 0 | 10 U | NA | 101/ | 100 | NA | NA | NA NA | NA | NA | NA NA | NA |
| Atrazinc | NA | NA | NA | 10 tr | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| Benzaldehyde | NA | NA | NA | 10 1 | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| Benzo(a)anthracene | 10 U | 10 U | NA | 10 U | 10 U | NA | - 10 U | 10 1/ | NĂ | 9.4 U | 9.4 U | NA | NA | NA | NA |
| Benzo(a)pytene | 10 U | 10 U | NA . | 10 U | 10 U | NA | 10 U | 10 U | NA | 9.4 U | <u>9.4 U</u> | NA | NA | NA | NA |
| llenzo(h)fluoranthene | 100 | 10 0 | | 100 | 10 U | NA | 10 () | 10 10 | NA | <u>9.4 U</u> | 9.4 U | NA | NA | NA NA | <u>NA</u> |
| Benzolg, B.I) perylene | <u>NA</u> | NA | <u> </u> | 10 0 | 10 0 | <u>NA</u> | | 10 10 | <u>NA</u> | <u>NA</u> | NA NA | - NA | NA NA | | <u> </u> |
| hirt 2-chlomethoxy)methane | NA NA | NA | NA | 10 10 | 10.0 | NA | 101 | 10 0 | NA NA | NA | NA NA | NA | NA NA | NA NA | NA |
| hist 2-chlomethyliether | NA NA | NA | NA | 10 U | 10 11 | NA | 10 11 | 100 | NA NA | NA | NA | NA | NA | NA | NA |
| his(2-Ethylhexyl)phthalate | NA | NA | NA | 10 U | 10 U | NA | 10 t1 | 10 1 | NA | NA | NA | NA | NA | NA | NA |
| Butyi Benzyi Phthalate | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| Capmlactam | NA | NA | NA | 10 U | 10 U | NA | 10 Ü | 10 U | NA | NA | NA | NA | NA | NA | NA |
| Carbazole | NA | NA | NA | 10 U | 10 11 | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| K.hrysene | NA | NA IO II | NA | 10 0 | 10 0 | NA NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| Dibenzo(a.n.)anthracene | 100 | 10.0 | NA NA | 10 0 | 10.0 | <u>NA</u> | 10 0 | 10 0 | NA NA | 9.4 0 | 9,40 | NA NA | <u>NA</u> | | |
| l hethylohthalate | NA NA | NA | | 10.11 | 10.11 | - MA | | 100 | | NA | NA NA | NA NA | - <u>NA</u> | NA | NA |
| Dimethyl Phthalate | NA | NA | NA | 10 U | 10 U | NA | 10 11 | 10 17 | NA | NA | NA | NA | NA | NA | NA |
| di-N-Boiylphthalate | NA | NA | NA | 10 U | 10 U | NA | | 10 U | NA | NA | NA | NA | NA | NA | NA |
| di-n-Octyl Phthalate | NA | NA | NA | 10 U | 10 U | NA | 10 1 | 10 U | NA | NA | NA | NA | NA | NA | NA |
| Huoranthene | NA | NA | NA | 10 U | 10 11 | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| Fluorene | NA | NA | NA | 10 U | 10 1) | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| Hexachlorobenzene | NA NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | <u>NA</u> | NA | NA |
| Hexachlorobutadiene | NA NA | NA NA | NA NA | 10 0 | 10 (1 | <u>NA</u> | 10 0 | 10 0 | <u>NA</u> | NA | NA | NA NA | NA | <u> </u> | NA NA |
| lievachioroethane | | NA NA | NA | 10 10 | 10.11 | NA NA | 10 0 | 10 0 | | NA | <u>NA</u> | NA | NA . | I NA | |
| Indepo(1.2.3-cd)nyrene | 10 11 | 1017 | NA | 10 01 | 1011 | NA | 1011 | | NA | 9411 | 9411 | NA | NA | NA | NA |
| isophorone | NÁ | NA | NA | 10 U | 10 U | NA | 10 1 | 10 0 | NA | NA | NA | NA | NA | NA | NA |
| Naphthalene | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| Nitrohenzene | NĂ | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| n-Nitroso-di-n-Propyfamine | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| n-Nitrosodiphenylamine | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA NA |
| Pentachlorophenol ⁴ | 26 U | 25 U | NA | 25 U | 25 U | NA | 25 U | R | NA | <u>NA</u> | NA | NA | 24.11 | 24 U | NA |
| Phenanthtene | NA | NA | NA | 10 U | 10 U | NA | 10 U | 10 U | NA | NA | NA | NA | NA | NA | NA |
| i henol | 10 U | 10 U | NA | 10 U | 10 U | NA | <u>io u</u> | R | NA | NA | NA | NA | 9.4 1/ | 9.4.0 | NA NA |
| l'yrene | <u>NA</u> | NA NA | NA | 10 U | 10 U | NA | 10_U | <u>10 U</u> | NA | NA | NA NA | NA | NA | NA | NA |
| Aluminum | | NA | NA | 0111 | <u></u> | NA I | 0.00 | | | | | | NA | NA | NA |
| Antimony | | NA | NA | 0.06 1/ | 0.061 | NA | 0.06.11 | 0.057 | 31.4 NA | NA | NA | NA NA | NA | NA NA | |
| Arsenic | 0.01 U | 0.01 U | NA | 0.01 U | 0010 | NA | 0011 | 0000 | NA | NA | NA | NA NA | 0011 | 0010 | NA |
| Barium | NA | NA | NA | 0618 | 0.0631 | 162.9 | 0.103 | 0.0996 | 3.4 | NA | NA | NA | NA | NA | NA |
| Beryllium | NA | NA | NA | 0.005 U | 0.005 U | NA | 0.005 U | 0.005 U | NA | NA | NA | NA | NA | NA | NA |

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TABLE 2-3

COMPARISON OF QUALITY CONTROL/QUALITY ASSURANCE SAMPLES - WATER MEDIA

Peter Cooper Site Gowanda, New York

| | Sample Location, Identification and Date Collected | | | | | | | | | | | | | | |
|--|--|--------------------|------------|----------------|--------------------|------------|-----------|-------------|------------|-----------|-------------|------------|----------|-----------|------------|
| | Creek Water #1 | Creek Water #1 DUP | Relative | Creek Water #1 | Creek Water #3 DUP | Relative | MWFP-35 | MWFP-35 DUP | Relative | MWFP-1D | MWFP-3D DUP | Relative | MW-SD | MW-SD DUP | Relative |
| | 050201134 | 050201111 | Percent | 110700098 | 110700099 | Percent | 110700088 | 110700089 | Percent | 050101126 | 050101127 | Percent | 50301141 | 050301142 | Petrent |
| Constituent ² | 5/2/2001 | 5/2/2001 | Difference | 11/7/2000 | 11/7/2000 | Difference | 11/7/2000 | 11/7/2000 | Difference | 5/1/2001 | 5/1/2001 | Difference | 5/3/2001 | 5/3/2001 | Difference |
| Cadmisum | NA | NA | NA | 0.005 U | 0.005 U | NA | 0.005 U | 0.005 U | NA | NA | NA | NA | NA | NA | NA |
| Calcium | 51.8 | 54.1 | 4.3 | 58.3 | 57.6 | 1.2 | 360 | 344 | 4.5 | 348 | NA | NA | 586 | NA | NA |
| homium | 0.01 U | 0.01 | NA | 0.01 U | 0.01 U | NA | 0.01 U | 0.01 U | NA | 0.01 U | 0.01 U | ŇĂ | 0.01 U | 0.01 U | NA |
| Cobalt | NA | NA | NA | 0.05 U | 0.05 U | NA | 0.05 U | 0.05 (/ | NA | NA | NA | NA | NA | NA | NA |
| Copper | NA | NA | NA | 0.02 U | 0.02 U | NA | 0.02 U | 0 02 U | NA | NA | NA | NΛ | NA | ŇĂ | NA |
| Hexavalent Chromium | (0.01) R | (0.01) R | NA_ | 0010 | U 10.0 | NA | 001 t | 0010 | NA | (0.01) R | (0.01) R | NA | (0.01) U | (0.01) U | NA |
| Iron | 0.39 | 0 413 | 5.7 | 0.143 | 0.134 | 6.5 | 16 | 14.8 | 7.8 | 17.7 | 17.6 | 0.6 | 71.4 | NA | NA |
| lead | 0.005 U | 0.005 U | NA | 0.005 U | 0.005 U | NA | 0.005 U | 0 005 U | NA | 0.005 () | 0.005 U | NA | NA | NA | NA |
| Magnesium | 9.25 | 9.37 | 1.3 | 9.88 | 10 | 1.2 | 17.5 | 16.7 | 4.7 | 17.9 | NA | NA | 35.4 | NA | NA |
| Manganese | 0.0161 | 0.0165 | 2.5 | 0.0129 | 0.0131 | 1.5 | 2.08 | 2.24 | 7.4 | 1.96 | 1.96 | 0.0 | NA | NA | NA |
| Mercury | NA | NA | NA | 0.0003 U | 0 0003 U | NA | 0.0003 U | 0.0003 U | NA | NA | NA | NA | NA | NA | NA |
| Nicket | NA | NA | NA | 0.04 U | 0.04 U | NA | 0.04 U | 0.04 U | NA | NA | NA | NA | NA | NA | NA |
| Potassium | 2 U | 2 U | NA | 2 U | 2 U | NA | 6.6 | 6.63 | 0.5 | 5.68 | 5.68 | 0.0 | 2076 | NA | NA |
| Selenium | NA | NA | NA | 0.005 U | 0.005 U | NA | 0.0061 | 0.005 U | NA | NA | NA | NA | NA | NA | NA |
| Silver | NA | NA | NA | 0.01 U | 0010 | NA | 0.01 U | 0.01 U | NA | NA | NA | NA | NA | NA | NA |
| Sodium | NA | <u>NA</u> | NA | 13.4 | 13.2 | 1.5 | 122 | 115 | 5.9 | 78.9 | NA | NA | 27 | NA | NA |
| Thallium | NA | NA NA | NA | 0.01 U | 0010 | NA | 0.01 U | 0.01 U | NA | NA NA | NA | NA | NA | NA | NA |
| Vanadium | NA | NA | NA | 0.05 U | 0 05 U | NA | 0.05 U | 0.05 U | NA | NA | NA | <u>NA</u> | NA | NA | NA |
| Zinc | 0.02 U | 0.02 U | NA | 0.02 U | 0.02 U | NA | 0.0551 | 0.0412 | 28.9 | NA | <u>NA</u> | NA NA | 0.02 U | 22.6 | NA |
| Soluble Metals, milligrams per liter | | | | | | | | | | | | | | | I |
| Chromium | NA | NA | NA | NA | NA | NA | NA | <u>NA</u> | NA | 0.01 U | 0.01 U | NA | NA | NA | NA |
| Hexavalent Chromium | NA | NA | NA | NA | NA | NA | NA | NA | NA | (0.01) R | (0.01) R | NA | NA | NA | NA |
| Iron | NA | NA | NA | NA | NA | NA | NA | NA | NA | 16.4 | 16 | 0.6 | NA | NA | NA |
| I zad | NA | NA | NA | NA | NA | NA | NA NA | NA | NA | 0.005 U | 0.005 U | NA | NA | NA | NA NA |
| Manganese | NA | NA NA | NA. | NA | NA | NA | NA | NA | NA | 1.89 | 1.9 | 0.5 | NA | NA | NA |
| Other Geochemical Parameters, milliorante per liter | | | | | | | | | 1 1 | | | | | | 1 |
| Aumonia | 0.05.11 | 0.05 11 | NA | 0.234 | 0.234 | 00 | ŇĂ | NA | NA | NA | NA | NA | NA | NA | NA |
| Bicarbonate Alkalinity | 270 1 | 1121 | 827 | 164 | 164 | 0.0 | \$58 | 550 | 14 | 480 | NA | NA | NA | NA | NA |
| Carbonate Alkalinity | 2.11 | 2 () | NA | 211 | 2 11 | NA | -2 U | -210 | NA | -21/ | NA | NA | NA | NA | NA |
| Chloride | 26.4 | 26.5 | 0.4 | 23.4 | 22.8 | 2.6 | NA | NA | NA | 77.7 | NA | NA | NA | NA | NA |
| Ferrous Iron | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Nitrate Nitrogen | 1.07 | 1.04 | 2.8 | 1.81 | 1.79 | 1.1 | NA | ŇĂ | NA | NA | NA | NA | NA | NA | NA |
| Soluble Organic Carbon | NA | NA | NA | NA | NA | NA | 5.41 U | 5.42 U | NA | 4.92 | NA | NA | NA | NA | NA |
| Sulfate | 24.8 | 24.7 | 0.4 | 27.5 | 27.6 | 0.4 | 651 | 631 | 3.1 | 544 | NA | NA | NA | NA | NA |
| Fotal Alkalinity | 270 J | 112 J | 82.7 | 164 | 164 | 0.0 | 558 | 550 | 1.4 | 480 | NA | NA | NA | NA | NA |
| Total Dissolved Solids | 216 | 215 | 0.5 | 249 | 254 | 2.0 | 1570 | 1590 | 1.3 | 1350 | NA | NA | NA | NA | NA |
| Total Hardness | 166 | 165 | 0.6 | 195 | 191 | 2.1 | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Fotal Kjeldahl Nitrogen | 0.345 | 0.281 | 20.4 | 0.417 | 0.239 | 54.3 | NA | NA . | NA | NA | NA | NA | NA | NA | NA |
| Fotal Organic Carbon | 1.67 | 1.73 | 3.8 | 2.14 | 1.99 | 7.2 | 5.19 U | 5.23 U | NA | 4.36 U | NA | NA | NA | NA | NA |
| Total Suffide | 2 111 | 2 ()) | NA | 10 | iu | NA | -i.i U | -1.1 U | NA | -2 UJ | NA | NA | NA | NA | NA |
| Total Suspended Solids | 66 | 6.4 | 3.1 | 1.30 J | 2.76 J | 71.9 | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| Induity NT11 | NA | NA | NA | NA | NA | NA | NA | NA | NA | 70 7 | 514 | 15.5 | NA | NA | NA |

Notes:

Sample locations provided on Plate 1.
 Zanaple locations provided on Plate 1.
 Data qualifications reflect 100% data validation performed by Data Validation Services.

J = indicates an estimated value.

5

U = indicates compound was not detected at or above the listed detection limit.

R= indicates value was rejected by data validator.

UJ = indicates compound was not detected above the listed detection limit. However, the reported quantitation limit is approximate and may or may

not represent the actual limit of quantilitation necessary to accurately and precisely measure the comprand in the sample. NA = Not applicable; RPD cannot be calculated when analyte is qualified with a U or UJ.

NTU = Nephelometric Turbidity Unit

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VILLAGE OF GOWANDA POPULATION - 1940 TO 2000¹

Peter Cooper Site Gowanda, New York

| Year | Gowanda (Catt. County) | Gowanda (Erie County) | Gowanda (all) |
|------|---------------------------|--------------------------|------------------|
| 1940 | 2,206 | 950 | 3,156 |
| 1950 | 2,221 | 1,068 | 3,289 |
| 1960 | 2,273 | 1,079 | 3,352 |
| 1970 | 2,098 | 1,012 | 3,110 |
| 1980 | 1,864 | 849 | 2.713 |
| 1990 | 2,016 | 885 | 2,901 |
| 2000 | 1,961 | 881 | 2.842 |



Notes:

1. Population data obtained from United States Census Bureau.



VILLAGE OF GOWANDA AND SURROUNDING TOWNSHIPS POPULATION 1990 - 2000

| Towns (includes Villages) | | Population ¹ | |
|--|---------------------------------------|--|--------|
| Census Year | 1990 | 2000 | change |
| Cattaraugus County | | | |
| Ashford | 2162 | 2223 | 61 |
| Dayton | 1931 | 1945 | 14 |
| East Otto | 981 | 1105 | 124 |
| Otto | 799 | 831 | 32 |
| Рептуѕвигд | 1838 | 1771 | (67) |
| Persia | 2514 | 2512 | (2) |
| Erie County | · · · · · · · · · · · · · · · · · · · | | |
| Brant | 2119 | 1906 | (213) |
| Collins ^{2.3} | 6020 | 8307 | 2287 . |
| Collins (less V. of Gowanda & CC Facility) | 4250 | 5112 - • • • • • • • • • • • • • • • • • • | (738) |
| Concord | 8387 | 8526 | 139 |
| North Collins | 3502 | 3376 | (126) |
| Cattaraugus Indian Reservation | 1789 | 2001 | 212 |
| TOTAL | 32,042 | 34,503 | 2,461 |

Peter Cooper Site Gowanda, New York

Notes:

1. Population data obtained from United States Census Bureau.

3. The Town of Collins 2000 population data includes data for population in group quarters (i.e., Collins Correctional Facility, pop. 3,914) and the Village of Gowanda population within Erie County (pop. 881).

^{2.} The Town of Collins 1990 population data includes data for population in group quarters (i.e., Collins Correctional Facility, pop. 885) and the Village of Gowanda population within Erie County (pop. 885).



STRATIFICATION SUMMARY

Peter Cooper Site Gowanda, New York

| | Surface | Cover | | Alluvial | Top of I | Bedrock |
|----------------------------|-----------------|-----------------|-----------------|------------------------|-----------------|-----------------|
| Boring | Elevation | Soil (1) | Fill (2) | Deposit ⁽³⁾ | Elevation | Depth |
| Number | (fmsl) | Thickness (ft) | Thickness (ft) | Thickness (ft) | (fmsl) | (f t) |
| Unit Screened: Alluvial De | posits | | • | | | |
| MW-1S | 778.1 | not encountered | not encountered | 10.0 | 768.1 | 10.0 |
| MW-3(R) | 768.1 | 0.5 | 3.5 | 3.5 | 760.6 | 7.5 |
| MW-7S | 786.1 | 2.5 | 3.5 | 10.0 | not encountered | not encountered |
| MW-8S | 778.1 | not encountered | 8.5 | 10.0 | not encountered | not encountered |
| MWFP-3S | 778.5 | 0.5 | 3.5 | 7.5 | 767.0 | 11.5 |
| Unit Screened: Fill | | | | | | |
| -Cindery Fill | | | | | | |
| MW-5S | 779.1 | 0.5 | 11.5* | not encountered | 764.1 | 12.0 |
| MWFP-2S | 784.3 | 0.2 | 10.8 | not encountered | 773.3 | 11.0 |
| -Sludge Fill | | | | | | |
| MW-2S(R) | 768.2 | 2.0 | 5.5 | not encountered | 760.7 | 7.5 |
| MW-4S(R) | 765.2 | 0.5 | 5.0 | not encountered | 759.7 | 5.5 |
| MW-6 | 781.5 | 3.0 | 12.0 | 3.0 | 763.5 | 18.0 |
| PZ-1 | 770.0 | 0.5 | >13.5 | not encountered | not encountered | not encountered |
| GMW-1 | 787.1 | 1.0 | 23.0 | 1.4 | 761.7 | 25.4 |
| GMW-2 | 787.1 | 1.0 | 18.0 | 5.9 | 763.2 | 23.9 |
| GMW-3 | 788.4 | 1.5 | 16.5 | 3.2 | 767.2 | 21.2 |
| Unit Screened: Shallow Be | drock | | | | | |
| MW-1D | 777.6 | not encountered | not encountered | 10.0 | 767.6 | 10.0 |
| MW-2D | 781.3 | 3.0 | 11.0 | 5.0 | 763.2 | 18.1 |
| MW-4D(R) | 765.0 | 0.5 | 5.0 | not encountered | 759.5 | 5.5 |
| MW-5D | 779.3 | 0.5 | 11.5 | not encountered | 761.3 | 12.0 |
| MW-7D | 785.8 | 2.5 | 3.5 | 10.0 | 769.8 | 16.0 |
| MW-8D | 778.0 | not encountered | 8.5 | 10.0 | 759.5 | 18.5 |
| MWFP-1D | 785.2 | 0.5 | not encountered | 4.0 | 780.7 | 4.5 |
| MWFP-2D | 784.1 | 0.2 | 10.8 | not encountered | 773.1 | 11.0 |
| MWFP-3D | 778.7 | 0.5 | 3.5 | 7.5 | 767.2 | 11.5 |
| Unit Screened: Deep Bedro | ck | | | | | |
| MW-4D2 | 765.1 | 0.5 | 5.0 | not encountered | 759.6 | 5.5 |
| Former Manufacturing Pla | nt Soil Borings | | | | | |
| SB-1 | 789.9 | 1.0 | 9.0 | not encountered | 779.0 | 10.0 |
| SB-2 | 784.0 | not encountered | 8.0 | not encountered | 776.0 | 8.0 |
| SB-3 | 782.0 | concrete pad | 6.0 | >2.0 | not encountered | not encountered |
| SB-4 | 783.6 | 0.5 | 9.0 | >2.5 | not encountered | not encountered |
| SB-5 | 785.4 | 0.5 | 5.0 | >6.5 | not encountered | not encountered |
| SB-6 | 780.3 | concrete pad | 4.5 | 3.5 | not encountered | not encountered |
| SB-7 | 789.9 | 0.5 | >11.5 | not encountered | not encountered | not encountered |
| SB-8 | 787.6 | 0.5 | 7.5 | >4.0 | not encountered | not encountered |
| SB-9 | 778.4 | 0.5 | 4.5 | >7.0 | not encountered | not encountered |
| SB-10 | 779.3 | 0.5 | 5.0 | >6.5 | not encountered | not encountered |

Notes:

1. Cover soil consist of sandy silty topsoil layer.

2. Fill consists of poorly graded granular soil with cinders and other anthropogenic material or sludge-like fill found in the Inactive Landfill Area.

3. Alluvial deposits consist of native clay, silt, sand, and gravel.

ft = feet

fmsl = feet mean sea level

* = indicates estimated from O'Brien and Gere boring logs.

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TABLE 3-4

SUMMARY OF WATER LEVEL DATA AND VERTICAL HYDRAULIC GRADIENTS

Peter Cooper Site Gowanda, New York

| Measuring | Top of Riser | | | | Date | | | | | | | | |
|---------------------|-----------------------|--------------------------|-----------|------------|-----------|------------|----------|-----------|-----------|--------|--------|--|--|
| Point | Elevation (1) (famsl) | 8/14/2000 | 9/29/2000 | 10/30/2000 | 11/6/2000 | 12/21/2000 | 2/2/2001 | 3/20/2001 | 4/30/2001 | Max. | Min. | | |
| MW-ID | 779.49 | 767.18 | 766.77 | 767.28 | 767.23 | 767.79 | 767.65 | 767.73 | 767.72 | 767.79 | 766.77 | | |
| MW-1S | 779.62 | 772.04 | 771.34 | 771.31 | 771.01 | 773.34 | 773.1 | 773.88 | 773.01 | 773.88 | 771.01 | | |
| MW-2S(R) | 770.93 | 763.50 | 763.26 | 762.52 | 762.32 | 762.73 | 762.74 | 762.85 | 762.70 | 763.50 | 762.32 | | |
| MW-2D | 782.82 | 755.57 | 755.09 | 754.90 | 754.87 | 755.66 | 755.62 | 755.63 | 755.34 | 755.66 | 754.87 | | |
| MW-3(R) | 770.70 | 763.47 | 763.18 | 762.90 | 762.65 | 763.12 | 762.41 | 763.36 | 763.31 | 763.47 | 762.41 | | |
| MW-45(R) | 766.97 | 762.23 | 762.25 | 762.57 | 762.49 | 762.09 | 761.67 | 762.08 | 762.53 | 762.57 | 761.67 | | |
| MW-41X(R) | 766.36 | 754.38 | 755.05 | 753.51 | 753.46 | 754.22 | 754.25 | 754.14 | 753.78 | 755.05 | 753.46 | | |
| MW-4D2 | 766.36 | | F 194 | 753.86 | 753.81 | 754.62 | 754.62 | 754.47 | 753.44 | 754.62 | 753.44 | | |
| MW-5 | 781.16 | 771.03 | 770.63 | 770.52 | 770.47 | 770.92 | 770.32 | 771.09 | 771.38 | 771.38 | 770.32 | | |
| MW-5D | 781.04 | 7 | 770.69(2) | 770.54 | 770.22 | 770.96 | 770.35 | 771.16 | 771.43 | 771.43 | 770.22 | | |
| MW-6 ⁽³⁾ | 783.58 | 771.76 | 771.46 | 772.07 | 771.87 | 773.22 | 773.04 | 774.30 | 773.99 | 774.30 | 771.46 | | |
| MW-75 | 787.77 | 19 A. | 773.18(2) | 773.48 | 773.36 | 776.63 | _ 775.22 | 775.13 | 775.35 | 776.63 | 773.36 | | |
| MW-7D | 787.38 | | 769.86(2) | 765.93 | 765.89 | 766.50 | 766.3 | 766.49 | 766.27 | 766.50 | 765.89 | | |
| MW-8S | 777.44 | | | 770.31 | 770.09 | | 771.72 | 772.71 | 771.51 | 772.71 | 770.09 | | |
| MW-8D | 777.64 | A 1 | . 4 | 770.30 | 770.90 | | 765.02 | 774.47* | 770.93 | 770.93 | 765.02 | | |
| DP-1 | 761.38 | -62^{10} $\pm 10^{10}$ | | 757.34 | 757.71 | 758.68 | 758.66 | 758.48 | 753.37 | 758.68 | 753.37 | | |
| PZ-1 | 772.31 | | · • ** | 763.91 | 763.81 | 763.97 | 763.52 | 764.39 | 769.26 | 769.26 | 763.52 | | |
| MWFP-1D | 787.30 | | | 776.12 | 776.05 | 777.38 | 777.14 | 777.49 | 777.45 | 777.49 | 776.05 | | |
| MWFP-2S | 786.00 | - 14 | · | 774.37 | 774.09 | 775.89 | 774.71 | 775.70 | 775.14 | 775.89 | 774.09 | | |
| MWFP-2D | 785.17 | 1; | | 772.82 | 772.82 | 772.46 | 772.38 | 772.36 | 772.22 | 772.82 | 772.22 | | |
| MWFT-3S | 780.69 | | | 772.51 | 772.34 | 773.58 | 772.6 | 774.17 | 773.70 | 774.17 | 772.34 | | |
| MWFP-3D | 780.51 | 7 | | 765.24 | 765.20 | 765.81 | 765.91 | 765.87 | 765.80 | 765.91 | 765.20 | | |
| Creek Upstream | | | | | | 772.40 | 772.39 | 772.19 | 772.08 | 772.40 | 772.08 | | |
| Creek Downstream | | | | | | 754.34 | 754.49 | 754.34 | 754.21 | 754.49 | 754.21 | | |

| Well Pair | Water Elevation ** | Top of Screen | Bottom of Screen | Adjusted Water Elevation for "Delta L" *** | "Detta L" | "Delta II" | Vertical Gradient |
|-----------|--------------------|---------------|------------------|---|-----------|------------|-------------------|
| MW-IS(R) | 773.1 | 772.6 | 767.6 | 770.1 | 26.9 | 5.45 | -0.20 |
| MW-ID | 767.65 | 745.7 | 740.7 | 743.2 | | | |
| MW-2S(R) | 762.74 | 763.7 | 759.7 | 762.74 | 17.24 | 7.12 | -0.41 |
| MW-2D | 755.62 | 748 | 743 | 745.5 | | | |
| MW-45(R) | 761.67 | 760.7 | 756.2 | 758.45 | 13.95 | 7.42 | -0.53 |
| MW-4D(R) | 754.25 | 747 | 742 | 744.5 | | | |
| MW-41X(R) | 754.25 | 747 | 742 | 744.5 | 14.4 | -0.37 | 0.03 |
| MW-4D2 | 754.62 | 735.1 | 725.1 | 730.1 | | | |
| MW-5 | 770.32 | 766.1 | 764.1 | 765.1 | 9.3 | -0.03 | 0.003 |
| MW-5D | 770.35 | 760.8 | 750.8 | 755.8 | | | |
| MW-75 | 775.22 | 782.1 | 769.6 | 775.22 | 19.92 | 8.92 | -0.45 |
| MW-7D | 766.3 | 760.3 | 750.3 | 755.3 | | | |
| MW-8S | 771.72 | 772.1 | 762.1 | 771.72 | 33.72 | 6.70 | -0.20 |
| MW-8D | 765.02 | 743 | 733 | 738 | | | |
| MWFP-2S | 774.71 | 779.3 | 772.3 | 774.71 | 13.61 | 2.33 | -0.17 |
| MWFP-2D | 772.38 | 766.1 | 756.1 | 761.1 | | 1 | |
| MWFP-3S | 772.6 | 773.5 | 767 | 772.6 | 14.9 | 6.69 | -0.45 |
| MWFP-3D | 765.91 | 762.7 | 752.7 | 1 7577 | | 1 | I I |

Notes:

(1) Top of riser or staff gage zero elevation as measured by TVGA Engineering, Surveying P.C., August 28, 2000.
(2) Water levels was measured prior to development.
(3) The riser pipe from MW-6 was cut down (approximately 3-inches) on 10%/00 and the elevation shown is current conditions. fams! = feet above mean sea level

fotor = feet below top of riser

-- = water level not collected as a result of snow cover

* = water was observed in the annular space of the road box.

** = (water levels taken 2/2/01)

*** = (The midpoint of the well screen was used for the elevation if the water level was above the well screen) shaded cells indicate the well was not installed or had been installed less than 24 hours prior to groundwater elevation measurements.



HYDRAULIC CONDUCTIVITY ESTIMATES

Peter Cooper Site Gowanda, New York

| | Material | Screen In | iterval | Estimated Hydraulic Conductivity ² | | | |
|------------------------------|-----------------------|--------------------------------|--------------|--|---|--|--|
| Well I.D. | Screened | Elevation ¹ (famsl) | Depth (fbgs) | (cm/sec) | (ft/day) | | |
| | | | | Range: 2.9x10 ⁻⁶ to | o 4.3x10 ⁻³ cm/sec | | |
| Alluvial Depostits | • | <u> </u> | ····· | : 0.008 to 1 | 2.2 ft/day | | |
| MW-1S | sand, gravel | 772.6-767.6 | 5.5-10.5 | 4.3x10 ⁻³ | 12.2 | | |
| <u>MW-3(R)</u> | silt, sand and gravel | 763.6-759.1 | 4.5-9.0 | 6.5x10 ⁻⁴ | 1.84 | | |
| MW-7S | silt | 782.1-769.6 | 4.0-16.5 | 2.9x10 ⁻⁶ | 0.008 | | |
| MW-8S | silt, sand and gravel | 772.1-762.1 6.0-16.0 | | 1.8x10 ⁻³ | 5.10 | | |
| MWFP-3S | silt, sand and gravel | 773.5-767.0 | 5.0-11.5 | 2.4×10^{-3} | 6.80 | | |
| Fill -Cindery Fill | | | | Range: 1.3x10 ⁻⁴ to : 0.37 to 62 | 2.2x10 ⁻² cm/sec 2.4 ft/day | | |
| MW-5S | sandfill/bedrock | 766.1-764.1 | 13.0-15.0 | 2.2x10 ⁻² | 62.4 | | |
| MWFP-2S | cindery fill | 779.3-772.3 | 5.0-12.0 | 1.3x10 ⁻⁴ | 0.37 | | |
| -Sludge Fill | | | | Range: 3.8x10 ⁻⁴ to : 1.08 to 93 | 3.3x10 ⁻² cm/sec 3.6 ft/day | | |
| MW-2S | sludge fill | 763.7-759.7 | 4.5-8.5 | 1.2x10 ⁻³ | 3 40 | | |
| MW-4S(R) | sludge/bedrock | 760.7-756.2 | 4.5-9.0 | 3.8x10 ⁻⁴ | 1.08 | | |
| MW-6 | sludge/silt and sand | 768.5-763.5 | 13.0-18.0 | 3.3x10 ⁻² | 93.6 | | |
| | | | | Range: 2.2x10 ⁻⁶ to | 3.4x10 ⁻² cm/sec | | |
| Shallow Bedrock | | <u> </u> | | : 0.00615 t | o 96.4 ft/day | | |
| MW-1D | shale | 745.7-740.7 | 31.9-36.9 | 2.5x10 ⁻⁴ | 0.71 | | |
| MW-2D | shale | 748.0-743.0 | 33.3-38.3 | 1.2×10^{-3} | 3.40 | | |
| MW-4D(R) | shale | 747.0-742.0 | 18.0-23.0 | 1.1×10^{-4} | 0.31 | | |
| MW-5D | shale | 760.8-750.8 | 18.5-28.5 | 3.4×10^{-2} | 96.4 | | |
| MW-7D | shale | 760.3-750.3 | 25.5-35.5 | 1.3x10 ⁻³ | 3.69 | | |
| MW-8D | shale | 743.0-733.0 | 35.0-45.0 | 2.2×10^{-6} | 0.006 | | |
| MWFP-1D | shale | 772.7-762.7 | 12.5-22.5 | 4.1×10^{-4} | 1.16 | | |
| MWFP-2D | shale | 766.1-756.1 | 18.0-28.0 | 6.4×10^{-4} | 1.81 | | |
| MWFP-3D | shale | 762.7-752.7 | 16.0-26.0 | 6.7×10^{-3} | 19.0 | | |
| Deep Bedrock | | | | 5.5x10 ⁻⁶ cm/sec; 0. | 016 ft/day | | |
| MW-4D2 | shale | 735.1-725.1 | 30.0-40.0 | 5.5x10 ⁻⁶ | 0.016 | | |

Notes:

1. Survey completed by TVGA Engineering, Surveying P.C., August 28, 2000.

 Hydraulic conductivity estimated by Geomatrix Consultants, Inc. using Bouwer and Rice Methods.

famsl = feet above mean sea level fbgs = feet below ground surface cm/sec = centimeters per second ft/day = feet per day




LANDFILL WASTE LIMITS

Peter Cooper Site Gowanda, New York

| Date | Test Pit Location | Description |
|------------|---|--|
| 10/05/00 | Elevated Fill Area Waste Limit A | 0-2' clay cover soil w/some light brown sandy material. Railroad spur and ties encountered at 2 ft w/ black cindery material, creosote-type odor. Transition from black sludge type material to black cindery material at the elevated fill area limit. |
| 10/12/2000 | Elevated Fill Area Waste Limit B | Black waste sludge with transition to cindery brick slag material |
| 10/12/2000 | Elevated Fill Area Landfill Waste Limit C | Black waste sludge material with transition to ash ballast slag material. Cover soil decrease from approximately 1.5'on the elevated fill area to < 6" off the elevated fill area. |
| 10/12/2000 | Elevated Fill Area Landfill Waste Limit D | Silty/clay cover soils w/black sludge material underlying to a transition of cindery sand and gravel cover soils with railroad tie and ballast. Tar-like odor detected. |
| 10/12/2000 | Elevated Fill Area Landfill Waste Limit E | Transition of black sludge material to cindery ash ballast material. |

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I:\Project\005771 PRP Group Peter Cooper NPL\RI report\FINAL REPORT (November 2003 Submittal)\Tables (Final)\Table 3-6 Peter Cooper Landfill Limits TP Log





SLUDGE FILL MATERIAL TESTING SUMMARY

Peter Cooper Site Gowanda, New York

| Sample Type | Sample Number | Dry D | ensity cf | Water (% | Content 6 | Hydraulic Conductivity |
|-----------------------|------------------|--------|--------------|--------------|--------------|---------------------------|
| | Ivumber | Before | After | Before | After | cm/s |
| Discrete(Shelby Tube) | ST-2 | 54.2 | 54.2 | 40.9 | 41.0 | 1.7 E-5 |

| Sample Type | Sample | Gravel | Sand | Silt | Clay | Liquid Limit | Plasticity Index |
|-------------|--------|--------|------|------|------|--------------|------------------|
| | Number | % | % | % | % | % | % |
| Grab | # 1 | 27.5 | 33.2 | 31.3 | 8 | 62.1 | 17.5 |

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SUMMARY OF TEST HOLES TO EVALUATE EXISTING COVER SOIL THICKNESS

| Test Hole | Cover Soil Thickness | Test Hole Depth | Depth Range | Description of Lithology |
|-----------|-------------------------|--------------------|----------------|--|
| No. | (inches) | (inches) | (inches) | |
| TUI | 13 | > 13 | 0-13" | Grayish Brown silty sand, trace gravel & sand |
| 111-1 | 15 | >13 | 13-?" | Waste cinders |
| THIN | 18 | 30 | 0-18" | Gray sandy silt, trace clay, gravel |
| 111-2 | 10 | 50 | 18-30" | Dark brown sand & gravel fill with little brick, wood |
| | | | 0-18" | Gray silt & fine sand |
| TH-3 | 18 | 44 | 18-32" | Brown sandy waste material |
| | | | 32-44" | Black waste with sand & brick |
| TH-4 | 22 | 32 | 0-22" | Olive gray sandy silt, trace clay & gravel |
| | | | 22-32" | Brown and rust colored fill with wood, glass, gravel |
| TH-5 | 7 | 10 | 0-7" | Gray silt and sand, trace clay and gravel |
| | | | 7-10" | Black sludge, very strong odor |
| TH-6 | 12 | 13 | 0-12" | Gray & dark gray silt and sand, trace clay and grave |
| | | | 12-13 | Black sludge |
| TH-7 | 48 | 53 | U-40 40 52" | Dialy sill and sand, trace clay and graves |
| | | | 48-33 | Dive menter the site trace along the group |
| тн-8 | 38 | 43 | 0-36 | Plack sludge |
| | | | 0.18" | Gray silly sand with trace clay and gravel |
| TH-9 | 18 | 21 | 18-21" | Black sludge |
| | | | 0-14.4" | Olive brown to may silt with trace clay little sand |
| TH-10 | 14.4 | 16.8 | 14.4-16.8" | Gravish black sandy material with odor |
| | | | 0-18" | Gray silt and sand trace gravel |
| TH-11 | 18 | > 18 | 18-?" | Black waste |
| | | | 0-15" | Silt and fine sand, trace gravel |
| TH-12 | 15 | > 15 | 15-?" | Black waste |
| | 10 | 24 | 0-18" | Gray fine sand and silt with trace gravel |
| 111-13 | 18 | 24 | 18-24" | Black sludge |
| TU 14 | 10 | ~ 10 | 0-18" | Sand and gray silt, trace gravel |
| 11-14 | 10 | >10 | 18-?" | Black sludge with odor |
| TU 15 | 23 | 26 | 0-23" | Gray silt and sand, trace clay and gravel |
| 111-15 | 23 | 20 | 23-26" | Black sludge |
| TH 16 | 30 | 34 | 0-32" | Gray silt and sand, trace clay and trace-little gravel |
| 11-10 | 32 | 34 | 32-34" | Black sludge |
| TH 17 | 31.2 | 31.2 | 0-31.2" | Brownish gray sandy silt with little gravel & silty sand |
| 111-17 | 21.2 | | 31.2" | Refusal on metal, likely bottom of 'cover' |
| TH-18 | 17 | 20 | 0-17" | Brownish gray/gray sandy silt w/trace clay & gravel |
| | | 20 | 17-20" | Black waste |
| TH-19 | 12 | >12 | 0-12" | Gray/brown fine sand and silt, trace gravel |
| | · · · · | - 14 | 12-?" | Black waste sludge with odor |
| тн-20 | 24 | 26 | 0-24" | Gray sand and silt |
| | | | 24-26" | Black sludge |
| TH-21 | 18 | > 18 | 0-18" | Silt and fine sand, trace gravel |
| | | | 18-?" | Black sludge waste |
| тн-22 | 22 | 25 | 0-22" | Gray sandy silt with trace clay and gravel |
| | ļ | | 22-25" | Black sludge |
| TH-23 | 41 | 44 | 0-41" | Gray silt and fine sand, little gravel, trace clay |
| | | | 41-44" | Black sludge |
| TH-24 | 20 | 25 | 0-20" | Gray sandy silt with trace clay and gravel |
| | | | 20-25" | Cinders |

Peter Cooper Site Gowanda, New York





EXISTING COVER SOIL TESTING SUMMARY

Peter Cooper Site

| Gowanda, | New | York |
|----------|-----|------|
|----------|-----|------|

| | | ASTM ³ | D421,422 | 2 | ASTM | D4318 | ASTM | D1557 | ASTM D5084 | Compaction | ASTM D2216 | |
|--------------------------------|-------------|-------------------|-----------|-----------|-------------------|-----------------------|------------------------------|------------------------------|------------------------------------|------------|--------------------|-------|
| Sample Number ¹ | Gravel % | Sand % | Silt % | Clay % | Liquid Limit % | Plasticity Index % | Maximum Dry Density (pcf) | Optimum Water Content (%) | Recompacted Permeability (cm/s) | % of MDD | Water Content % | USCS |
| Comp-1, TH-1 through TH-4 | 6.6 | 38.3 | 45.1 | 10 | 27.4 | 7.3 | 125.7 | 10.3 | 1.1 E-6 | 88.5 | 12.5 | CL-ML |
| Comp-2, TH-5 through TH-8 | 4.8 | 26.2 | 52.6 | 16.4 | 25.4 | 7.9 | 125.9 | 10.4 | 3.9 E-7 | 89.4 | 11.3 | CL-ML |
| Comp-3, TH-9 through TH-12 | 4.1 | 28.1 | 45.7 | 22.1 | 26.5 | 9.1 | 128.0 | 10.4 | 9.0 E-7 | 88.3 | 12.1 | CL |
| Comp-4, TH-13 through TH-16 | 2.7 | 26.8 | 50.5 | 20.0 | 23.0 | 6.0 | 130.1 | 9.0 | 1.8 E-6 | 86.5 | 12.4 | CL-ML |
| Comp-5, TH-17 through TH-20 | 5.3 | 36.7 | 40.8 | 17.2 | 23.9 | 4.7 | 124.7 | 11.4 | 3.2 E-6 | 85.9 | 11.6 | CL-ML |
| Comp-6, TH-21 through TH-24 | 5.4 | 27.5 | 45.7 | 21.4 | 22.9 | 6.9 | 130.3 | 9.2 | 3.8 E-7 | 87.3 | 10.9 | CL-ML |

| | ASTM | D3080 | ASTA | 1 D2216 | ASTM D5084 | |
|----------------------------|--------|--------------|--------|--------------|---------------------------|--|
| Sample Number ² | Dry D | ensity :f | Water | Content % | Hydraulic Conductivity | |
| | Before | After | Before | After | cm/s | |
| ST-1 | 117.2 | 115.5 | 15.9 | 16.1 | 1.1 E-5 | |
| ST-2 | 105.6 | 102.1 | 20.4 | 24.5 | 7.5 E-7 | |
| ST-3 | 112.5 | 119.2 | 19.1 | 15.7 | 9.0 E-8 | |
| ST-4 | 107.6 | 114.3 | 20.5 | 15.9 | 1.0 E-6 | |
| ST-5 | 126.2 | 123.9 | 11.0 | 13.1 | 2.4 E-7 | |
| ST-6 | 103.0 | 105.1 | 19.5 | 20.4 | 3.6 E-5 | |

Notes:

1. Samples identified as "Comp-#" represent composites of the identified test hole (TH) locations

2. Samples identified as "ST-#" represent shelby tube samples

3. ASTM followed by the letter D#### is the Method used for testing

pcf = pounds per cubic foot USCS = Unified Soil Classification System MDD = Maximun Dry Density

cm/s = centimeters per second % = percentage

301112

FORMER MANUFACTURING PLANT AREA GEOTECHNICAL TESTING SUMMARY

Peter Cooper Site Gowanda, New York

| Samala | | | Cravel | Sand | Fines | | |
|------------------|---------------------------------------|----------|--------|------|----------|--------|--|
| Number | Depth | Strata | Gravei | Suna | Silt | Clay | |
| Inumber | | | % | % | % | % | |
| Surface | · · · · · · · · · · · · · · · · · · · | | | | | | |
| SB-1 | 0-2' | Fill | 23.9 | 57.4 | 14.8 | 3.9 | |
| SB-2 | 0-2' | Fill | 49.3 | 43.1 | 7. | 6* | |
| SB-4 | 0-2' | Fill | 26.3 | 35.6 | 24.6 | 13.5 | |
| SB-5 | 0-2' | Fill | 18.0 | 30.4 | 36.6 | 15.0 | |
| SB-6 | 0-2' | Fill | 10.6 | 24.2 | 38.4 | 26.8 | |
| SB-7 | 0-2' | Fill | 21.4 | 52.7 | 15.4 | 10.5 | |
| SB-8 | 0-2' | Fill | 26.8 | 30.8 | 28.7 | 13.7 | |
| SB-10 | 0-2' | Fill | 29.6 | 37.7 | 22.4 | 10.3 | |
| Subsurface | | | | | | | |
| SB-1 | 5-7' | Fill | 27.0 | 61.1 | 11 | .9* | |
| SB-4 | 4-6' | Fill | 32.5 | 57.2 | 10 | .3* | |
| SB-2 | 6-8' | Fill | 42.4 | 43.5 | 14 | .1* | |
| SB-5 | 6-8' | Alluvial | 17.5 | 62.1 | 16.4** | 4.0** | |
| SB-3 | 3-5' | Alluvial | 0.0 | 42.0 | 39.3 | 18.7 | |
| MW-FP-2 | 0.5-2.5' | Fill | 56.3 | 28.9 | 14 | .8* | |
| MW-FP-2 | 5-7' | Fill | 17.9 | 69.9 | 12 | .2* | |
| SB-7 | 7-9' | Fill | 15.0 | 58.8 | 16.2** | 10.0** | |
| SB-8 | 10-12' | Alluvial | 11.8 | 35.2 | 39.0 | 14.0 | |
| SB-9 | 0.5-2.5' | Fill | 35.2 | 29.5 | 25.0 | 10.3 | |
| SB-9 | 7-9' | Alluvial | 1.1 | 58.2 | 32.7** | 8** | |
| SB-10 | 7-9' | Alluvial | 0.0 | 56.6 | 35.2 | 8.2 | |
| MW-FP-3 | 0.5-2.5' | Fill | 39.5 | 40.5 | 16.0 | 4.0 | |
| MW-FP-3 | 5-7' | Alluvial | 17.1 | 54.0 | 23.9 | 5.0 | |
| SB-6 | 4-6' | Fill | 15.9 | 52.4 | 22.7 | 9.0 | |
| Wetland Sediment | | | | | <u> </u> | | |
| #1 | | | 0.2 | 45.8 | 47.5 | 6.5 | |
| #2 | | | 0.7 | 35.8 | 55.9 | 7.6 | |
| # 3 | | | 0.3 | 42.1 | 44.2 | 13.4 | |
| Creek Sediment | | J | | | L | | |
| #1 | | | 16 | 82.3 | 1 | .7 | |
| # 2 | | | 20.1 | 78.9 | | 1 | |
| # 3 | | | 3.9 | 94.5 | 1 | .6 | |
| # 4 | | | 0.3 | 95.4 | 4 | .3 | |

Notes:

% = Percent

* Predominantly Silt

** Approximate percentages based on shape of curve.



WETLAND AND CREEK SEDIMENT GEOTECHNICAL TESTING SUMMARY

Peter Cooper Site Gowanda, New York

| Wetland Sediment | | | | | | | |
|---------------------|-------------|-----------|-----------|-----------|--|--|--|
| Sample Number | Gravel % | Sand % | Silt % | Clay % | | | |
| Wetland Sediment #1 | 0.2 | 45.8 | 47.5 | 6.5 | | | |
| Wetland Sediment #2 | 0.7 | 35.8 | 55.9 | 7.6 | | | |
| Wetland Sediment #3 | 0.3 | 42.1 | 44.2 | 13.4 | | | |

| Creek Sealment | | | | | | | |
|-------------------|--------|------|-------------|--|--|--|--|
| Sample | Gravel | Sand | Silt & Clay | | | | |
| Number | % | % | % | | | | |
| Creek Sediment #1 | 16.0 | 82.3 | 1.7 | | | | |
| Creek Sediment #2 | 20.1 | 78.9 | 1.0 | | | | |
| Creek Sediment #3 | 3.9 | 94.5 | 1.6 | | | | |
| Creek Sediment #4 | 0.3 | 95.4 | 4.3 | | | | |

Creek Sediment

GEOM

Page 1 of 2

TABLE 4-1

ANALYTICAL RESULTS FOR INACTIVE LANDFILL SLUDGE FILL SAMPLES VOLATILE ORGANIC COMPOUNDS

| | Sample Location, Identification, Depth, and Date Collected | | | | | |
|--|--|---------------|--------------|--|--|--|
| | GMW-3 | GMW-2 | GMW-1 | | | |
| | 100900042 | 100900043 | 100900045 | | | |
| | 16-20 (ft-bgs) | 8-12 (ft-bgs) | 4-8 (ft-bgs) | | | |
| Constituent | 10/9/2000 | 10/9/2000 | 10/9/2000 | | | |
| Volatile Organic Compounds, milligrams | | | | | | |
| per kilogram | | | | | | |
| Acetone | 15 J | 15 J | 2.5 DJ | | | |
| Benzene | 0.0067 J | 0.013 J | 0.034 J | | | |
| Bromodichloromethane | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| Bromoform | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| Bromomethane | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| 2-Butanone (MEK) | 2.4 J | 3.2 J | 1 DJ | | | |
| Methyl tert-Butyl Ether | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| Carbon Disulfide | 0.079 J | 0.22 J | 0.24 J | | | |
| Carbon Tetrachloride | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| Chlorobenzene | 0.026 UJ | 0.024 UJ | 0.059 J | | | |
| Chloroethane | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| Chloroform | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| Chloromethane | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| 1.2-Dibromo-3-Chloropropane | 0.026 UJ | 0.024 UJ | 2.8 UJ | | | |
| Cyclohexane | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| Dibromochloromethane | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| 1,2-Dibromoethane | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| 1,2-Dichlorobenzene | 0.026 UJ | 0.024 UJ | 0.54 J | | | |
| 1,4-Dichlorobenzene | 0.026 UJ | 0.024 UJ | 2.2 DJ | | | |
| 1.3-Dichlorobenzene | 0.026 UJ | 0.024 UJ | 2.8 UJ | | | |
| Dichlorodifluoromethane | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| 1,1-Dichloroethane | 0.026 UJ | 0.024 UJ | 0.01 J | | | |
| 1,2-Dichloroethane | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| 1,1-Dichloroethene | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| trans-1,2-Dichloroethene | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| cis-1,2-Dichloroethene | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| 1,2-Dichloropropane | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| trans-1,3-Dichloropropene | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| cis-1,3-Dichloropropene | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| Ethylbenzene | 0.019 J | 0.054 J | 0.12 J | | | |
| 2-Hexanone | 0.076 J | 0.078 J | 0.1 J | | | |
| Isopropylbenzene | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| Methyl Acetate | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| Methylcyclohexane | 0.0054 J | 0.0096 J | 0.03 UJ | | | |
| Methylene Chloride | 0.026 UJ | 0.024 UJ | 0.03 UJ | | | |
| 4-Methyl-2-Pentanone | 0.086 J | 0.016 J | 0.24 J | | | |
| Styrene | 0.0034 J | 0.024 UJ | 0.03 UJ | | | |
| 1,1,2,2-Tetrachloroethane | 0.026 UJ | 0.024 UJ | 2.8 UJ | | | |
| Tetrachloroethene | 0.026 UJ | 0.024 UJ | 0.054 J | | | |
| Toluene | 1.7 DJ | 0.12 J | 0.37 J | | | |

Peter Cooper Site Gowanda, New York

I: Project 005771 PRP Group Peter Cooper NPL/RI report/FINAL REPORT (November 2003 Submittal) (Tables (Final) \Table 4-1, 2, & 4-4 landfill subsurface soils FINAL



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TABLE 4-1

ANALYTICAL RESULTS FOR INACTIVE LANDFILL SLUDGE FILL SAMPLES VOLATILE ORGANIC COMPOUNDS

| | Sample Location, | Identification, Depth, ar | nd Date Collected ¹ |
|---------------------------------------|---|--|---|
| Constituent | GMW-3 100900042 16-20 (fi-bgs) 10/9/2000 | GMW-2 100900043 8-12 (ft-bgs) 10/9/2000 | GMW-1 100900045 4-8 (ft-bgs) 10/9/2000 |
| 1,2,4-Trichlorobenzene | 0.026 UJ | 0.024 UJ | 2.8 UJ |
| 1,1,1-Trichloroethane | 0.026 UJ | 0.024 UJ | 0.03 UJ |
| 1,1,2-Trichloroethane | 0.026 UJ | 0.024 UJ | 0.03 UJ |
| Trichloroethene | 0.026 UJ | 0.024 UJ | 0.03 UJ |
| Trichlorofluoromethane | 0.026 UJ | 0.024 UJ | 0.03 UJ |
| 1,1,2-Trichloro-1,2,2-Trifluoroethane | 0.026 UJ | 0.024 UJ | 0.03 UJ |
| Vinyl Chloride | 0.026 UJ | 0.024 UJ | 0.03 UJ |
| m-/p-Xylene | 0.014 J | 0.046 J | 0.16 J |
| o-Xylene | 0.0067 J | 0.027 J | 0.047 J |

Peter Cooper Site Gowanda, New York

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

J = indicates an estimated value.

U = indicates compound was not at or above the listed detection limit.

D = indicates spike diluted out.

UJ = indicates compound was not detected above the listed detection limit. However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitation necessary to accurately and precisely measure the compound in the sample.

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TABLE 4-2

ANALYTICAL RESULTS FOR INACTIVE LANDFILL AREA SLUDGE FILL SAMPLES SEMI-VOLATILE ORGANIC COMPOUNDS AND METALS

| | Samula Logation |
|----------------------------------|------------------------|
| | Jampie Location, |
| | Dete Collected |
| | |
| | <i>COMP GMW-1,-2-3</i> |
| | 100900046 |
| - · | Composite 4-20 fbgs |
| Constituent | 10/9/2000 |
| Semi-Volatile Organic Compounds, | |
| milligrams per kilogram | |
| Acenaphthene | 5.6 U |
| Acenaphthylene | 5.6 U |
| Acetophenone | 5.6 U |
| Anthracene | 5.6 U |
| Atrazine | 5.6 U |
| Benzaldehyde | 5.6 U |
| Benzo(a)anthracene | 5.6 UJ |
| Benzo(a)pyrene | 5.6 U |
| Benzo(b)fluoranthene | 5.6 UJ |
| Benzo(g.h.i)perylene | 5.6 UJ |
| Benzo(k)fluoranthene | 5.6 UJ |
| 1,1-Biphenyl | 5.6 U |
| Butyl Benzyl Phthalate | 5.6 U |
| di-N-Butylphthalate | 5.6 U |
| Caprolactam | 5.6 U |
| Carbazole | 5.6 U |
| Indeno(1,2,3-cd)pyrene | 5.6 UJ |
| 4-Chloroaniline | 5.6 U |
| bis(2-chloroethoxy)methane | 5.6 U |
| bis(2-chloroethyl)ether | 5.6 U |
| 2-Chloronaphthalene | 5.6 U |
| 2-Chlorophenol | 5.6 U |
| 2,2-oxybis(1-chloropropane) | 5.6 U |
| Chrysene | 5.6 U |
| Dibenzo(a,h)anthracene | 5.6 UJ |
| Dibenzofuran | 5.6 U |
| 3,3-Dichlorobenzidine | (5600 U) R |
| 2,4-Dichlorophenol | 5.6 U |
| Diethylphthalate | 5.6 U |
| Dimethyl Phthalate | 5.6 U |
| 2,4-Dimethylphenol | 5.6 U |
| 2,4-Dinitrophenol | 14 UJ |
| 2,4-Dinitrotoluene | 5.6 U |
| 2,6-Dinitrotoluene | 5.6 U |
| bis(2-Ethylhexyl)phthalate | 5.6 U |
| Fluoranthene | 5.6 U |
| Fluorene | 5.6 U |
| Hexachlorobenzene | 5.6 U |

Peter Cooper Site Gowanda, New York

I:VProjectV005771 PRP Group Peter Cooper NPL/RI report/FINAL REPORT (November 2003 Submittal)/Tables (Final)/Table 4-1, 2, & 4-4 landfill subsurface soils FINAL

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ANALYTICAL RESULTS FOR INACTIVE LANDFILL AREA SLUDGE FILL SAMPLES SEMI-VOLATILE ORGANIC COMPOUNDS AND METALS

Sample Location, Identification, Depth, and Date Collected¹ COMP GMW-1,-2-3 100900046 Composite 4-20 fbgs 10/9/2000 Constituent Hexachlorobutadiene 5.6 U Hexachlorocyclopentadiene 28 U Hexachloroethane 5.6 Ū Isophorone 5.6 U 2-Methylnaphthalene 5.6 U 4,6-Dinitro-2-Methylphenol 14 U 4-Chloro-3-Methylphenol 5.6 U 2-Methylphenol 5.6 U 4-Methylphenol 150 D Naphthalene 22 14 U 2-Nitroaniline (14000 U) R 3-Nitroaniline 4-Nitroaniline 14 U 5.6 U Nitrobenzene 2-Nitrophenol 5.6 U 4-Nitrophenol 14 UJ n-Nitrosodiphenylamine 5.6 U di-n-Octyl Phthalate 5.6 UJ Pentachlorophenol 6.8 J Phenanthrene 1 J Phenol 15 4-Bromophenyl-Phenylether 5.6 U 4-Chlorophenyl-Phenylether 5.6 U n-Nitroso-di-n-Propylamine 5.6 U Pyrene 5.6 U 2,4,6-Trichlorophenol 5.6 U 2,4,5-Trichlorophenol 14 U Metals, mg/kg Aluminum 3780 Antimony 57.6 J Arsenic 34.8 175 Barium 0.83 Ū Beryllium Cadmium 1.5 122,000 Calcium 9280 Chromium Cobalt 8.3 U 156 Copper Hexavalent Chromium 6.75 Ū

Peter Cooper Site Gowanda, New York

E: Project/005771 PRP Group Peter Cooper NPL/RI report/FINAL REPORT (November 2003 Submittal)/Tables (Final)/Table 4-1, 2, & 4-4 landfill subsurface soils FINAL

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TABLE 4-2

ANALYTICAL RESULTS FOR INACTIVE LANDFILL AREA SLUDGE FILL SAMPLES SEMI-VOLATILE ORGANIC COMPOUNDS AND METALS

Peter Cooper Site

Gowanda, New York Sample Location, Identification, Depth, and Date Collected¹ COMP GMW-1,-2-3 100900046 Composite 4-20 fbgs 10/9/2000 14800 J 97.4

| Constituent | 10/9/2000 |
|-------------------------|-----------|
| lron | 14800 J |
| Lead | 97.4 |
| Magnesium | 9740 |
| Manganese | 250 |
| Mercury | 6.2 |
| Nickel | 10.6 |
| Potassium | 334 U |
| Selenium | 1.8 J |
| Silver | 1.7 U |
| Sodium | 1020 |
| Thallium | 1.7 U |
| Vanadium | 8.3 U |
| Zinc | 6060 |
| Other | |
| Percent Solids, % | 59.3 |
| рН | 7.86 |
| Total Organic Carbon, % | 10.0 |

Notes:

- 1. Sample locations provided on Plate 1.
- 2 Data qualifications reflect 100% data validation performed by Data Validation Services.

J = indicates an estimated value.

U = indicates compound was not detected.

D = indicates spike diluted out.

UJ = indicates compound was not detected above the listed detection limit. However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitation necessary to accurately and precisely measure the compound in the sample.

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ANALYTICAL RESULTS FOR SURFACE SOIL SAMPLES FROM THE INACTIVE LANDFILL AREA

Peter Cooper Site Gowanda, New York

| | | | | | | 5 | iample Locati | on, Identificatio | n, Depth, and | Date Collected | | | |
|--|---------------------------|---------------------------|-----------------------------|-------------|-------------|-------------|---------------|-------------------|---------------|----------------|-------------|-------------|-------------|
| | (| | , j | LFSS-I | LFSS-2 | LFSS-3 | LFSS-4 | LFSS-5 | LFSS-6 | LFSS-7 | LFSS-8 | LFSS-9 | LFSS-10 |
| · · · · · · | Se Se | oil Criteria ⁾ | ! | 101100058 | 101100059 | 101100060 | 101100061 | 101100062 | 101100069 | 101100064 | 101100065 | 101100066 | 101100067 |
| Constituent ² | Eastern USA Background | Region 9 PRGs | Soil Screening Levels | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs |
| | | ليستوجدهم | | 10/11/2000 | 10,11,2000 | 10,112000 | 10/11/2000 | | 10/11/2000 | 10/11/2000 | 10/11/2000 | 10/11/2000 | 10/11/2000 |
| Volatile Organic Compounds, milligrams per kilogram | | | | | <u> </u> ' | | | l! | | | | | |
| 1,2-dichlorobenzene | | 370 | 17 | 0.013 U J | 0.014 UJ | 0.014 U J | 0.012 U | 0.01 U | 0.015 U J | (0.015 U) R | 0.021 U J | 0.015 U J | 0.014 U J |
| 1,4-dichlorobenzene | | 7.9 | 2 | 0.013 U J | 0.014 U J | 0.014 UJ | 0.012 U | 0.01 U | 0.015 UJ | (0.015 U) R | 0.021 U J | 0.015 U J | 0.014 U J |
| Benzene | | 1.3 | 0.03 | 0.013 U J | 0.0016 J | 0.0042 J | 0.00511 | 0.0032 J | 0.0029 J | 0.015 U J | 0.0022 J | 0.015 U | 0.014 U J |
| Chlorobenzene | | 530 | | 0.013 U J | 0.014 U J | 0.014 U J | 0.012 U | 0.01 U | 0.015 Ū J | 0.015 U J | 0.021 U | 0.015 U | 0.014 U J |
| Ethylbenzene | | 20 | 13 | 0.013 UJ | 0.014 U J | 0.0018 J | 0.012 U | 0.01 U | 0.015 U J | 0.015 U J | 0.021 U | 0.015 U | 0.014 U J |
| m/p-Xylene | | 420 | 210 | 0.0016 J | 0.0023 J | 0.006 J | 0.0047 1 | 0.004 J | 0.015 U J | 0.015 U J | 0.0028 J | 0.015 U | 0.014 U J |
| o-Xylene | | 420 | 210 | 0.013 U J | 0.014 U J | 0.002 J | 0.0015 J | 0.01 U | 0.015 U J | 0.015 U J | 0.021 U | 0.015 U | 0.014 U J |
| Toluene | | 520 | 12 | 0.002 J | 0.0037 J | 0.0082 J | 0.0086 J | 0.0052 J | 0.015 U J | 0.015 U J | 0.0058 J | 0.0016 J | 0.014 U J |
| Metals, milligrams per kilogram | | | | | | | | | | | | | |
| Arsenic | 3-12** | 1.6 | 29.0 | 9.3 | 8.7 | 10.2 | 6.6 | 10.6 | 919 | · 2[1] | 7.2 | 11 | 8.7 |
| Chromium | 1.5-40** | 210 | 38 | 18.4 | 15.4 | 267 | 13 | 32.8 | 341 | 208 | 550 | 33.8 | 36.4 |
| Hexavalent Chromium | <u> </u> | 64 | 38 | 5.03 U | 5.35 U | 5.03 U | 5.28 U | <u>5.1 U</u> | 5.17 U | 5.12 U | 5.62 U | 5.42 U | 5.44 U |
| Zinc | 9-50 | 100,000 | 12,000 | 81.8 | 79.3 | 163 | 55 | 91.4 | 165 | 77.5 | 137 | 96.6 | 89.2 |
| Other | | | | | [| | | | | | | | |
| Percent Solids, % | | <u> </u> | [/ | 79.5 | 74.8 | 79.5 | 75.8 | 78.4 | 77.3 | 78.1 | 71.2 | 73.8 | 73.5 |
| Total Organic Carbon, % | | l' | <u> </u> | 2.30 | 2.40 | 2.80 | 2.10 | 2.70 | 5 | 5.60 | 3.40 | 2.40 | 3.70 |
| pH | | (, -) | (| 7.42 | 7.36 | 7.78 | 8.07 | 7.97 | 6.61 | 7.53 | 7.35 | 7.18 | 5.76 |

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

3. Soil criteria from U.S.EPA, Region 9 Preliminary Remediation Goals (PRGs) for Industrial Soil (October 2002) and from range of background metals concentrations measured in soil found in the

eastern United States from NYSDEC Division of Technical and Administrative Guidance Memorandum (TAGM) #4046.

** A New York State Background value

in. hgs =inches below ground surface.

-- = indicates value does not exist.

SB = Site Background

 $\label{eq:U} \begin{gathered} \texttt{U} = \texttt{indicates compound was not detected above the listed detection limit. However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitation necessary to accurately and precisely measure the compound in the sample. \end{gathered}$

 $\label{eq:J} J = indicates an estimated value. \\ U = indicates compound was not detected. \\ R = indicates data rejected by data validator. \\ (values) = indicates value reported before rejected. \\$

indicates concentration above soil criteria.

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TABLE 4-3

ANALYTICAL RESULTS FOR SURFACE SOIL SAMPLES FROM THE INACTIVE LANDFILL AREA

Peter Cooper Site Gowanda, New York

| | | | | | | | Samela Lagati | un Identificati | an Danth and D | ate Cellested! | | | | | |
|---------------------------------|-------------|------------|-----------|-------------|-------------|-------------|---------------|-----------------|----------------|----------------|-------------|--------------|-------------|----------|----------|
| | | | | 1 2 2 2 11 | I ECC 12 | 1866 13 | | I ESS 10 | I ESS 14 | EEC 12 | LECC 18 | 1 5 6 10 | 1.566.30 | | |
| | | 104.43 | | LF 33-11 | LF 33-12 | LF 53-13 | LF 55-14 | LF 33-13 | LF 35-10 | LF 55-17 | LF 33-18 | LF 35-19 | LF 55-20 | | |
| | | ou Crueria | | 101100068 | 101100070 | 1011000/1 | 1011000/2 | 101100073 | 101100074 | 101100075 | 101100076 | 101100077 | 101200078 | | |
| | Eastern USA | Region 9 | Soil | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | 0-6 in. bgs | Maximum | Minimum |
| | | | Screening | | | | |]] | | | | | | | |
| Constituent ² | Background | PRGs | Levels | 10/11/2000 | 10/11/2000 | 10/11/2000 | 10/11/2000 | 10/11/2000 | 10/11/2000 | 10/11/2000 | 10/11/2000 | 10/11/2000 | 10/12/2000 | Conc. | Conc. |
| Volatile Organic Compounds, | | | | | | | | | | | | | | | |
| milligrams per kilogram | | | | | | | | | | | | | | | |
| 1,2-dichlorobenzene | | 370 | 17 | 0.016 U J | 0.015 U J | 0.01 U J | 0.013 U J | 0.012 U J | 0.013 U J | 0.015 U J | 0.012 U J | 0.015 U J | 0.012 U J | 0.021 UJ | 0.01 UJ |
| 1,4-dichlorobenzene | | 7.9 | 2 | 0.016 U J | 0.015 U J | U 10.0 | 0.013 U J | 0.012 U J | 0.013 U J | 0.015 U J | 0.012 U J | 0.015 U J | 0.012 U J | 0.021 UJ | 0.01 UJ |
| Benzene | | 1.3 | 0.03 | 0.016 U J | 0.015 U J | 0.0021 J | 0.0025 J | 0.0026 J | 0.0019 J | 0.015 U J | 0.0036 J | 0.005 J | 0.004 J | 0.016 UJ | 0.0016 J |
| Chlorobenzene | | 530 | 1 | 0.016 U J | 0.015 U Ĵ | 0.01 U J | 0.013 U J | 0.012 U J | 0.013 U J | 0.015 U J | 0.012 U J | 0.015 U J | 0.012 U J | 0.021 U | 0.01 UJ |
| Ethylbenzene | | 20 | 13 | 0.016 U J | 0.015 U J | 0.01 U J | 0.0014 J | 0.012 U J | 0.013 U J | 0.015 U J | 0.0012 J | 0.015 U J | 0.001 J | 0.021 U | 0.0012 J |
| m/p-Xylene | | 420 | 210 | 0.016 U J | 0.015 U J | 0.0023 J | 0.0041 J | 0.0037 J | 0.0031 J | 0.015 U J | 0.0049 J | 0.015 U J | 0.006 J | 0.016 J | 0.0016 J |
| o-Xylene | | 420 | 210 | 0.003 J | 0.015 U J | 0.01 U J | 0.0018 J | 0.012 U J | 0.013 U J | 0.015 U J | 0.0016 J | 0.015 U J | 0.002 J | 0.021 U | 0.0015 J |
| Toluene | | 520 | 12 | 0.005 J | 0.015 U J | 0.0033 J | 0.0051 J | 0.0051 J | 0.0041 J | 0.015 U J | 0.007 J | 0.015 U J | 0.01 J | 0.015 UJ | 0.0016 J |
| | | | | | | | | | | | | | | | |
| Metals, milligrams per kilogram | | | | | | | | | | | | | | | |
| Arsenic | 3-12** | 1.6 | 29.0 | 9.1 J | 7.5 J | 7.2 J | 21.5 J | <u>6.5 J</u> | 9.4 J | 38.8 J | 6.9 J | 128' 7 | 4 J | 919 | 43 |
| Chromium | 1.5-40** | 210 | 38 | 40.1 J | 92 J | 15.5 J | 134 J | 11 J | 17.2 J | S 117 J 🗟 | 17.I J | <u>169 J</u> | 10.6 J | 550 | 10.6 J |
| Hexavalent Chromium | | 64 | 38 | 5.53 U J | 5.49 U J | 4.67 U J | 5.1 U J | 4.88 U J | 4.94 U J | 5.78 U J | 4.83 U J | 5.31 U J | 4.73 U J | 5.78 UJ | 4.67 UJ |
| Zinc | 9-50 | 100,000 | 12,000 | 75.1 J | 96.9 J | 54 J | 67.1 J | 46.9 J | ': 61.3 J " | : 85.9 J | 53.8 J | 103 J | 52.2 J | 165 | 46.9 J |
| | | | • | | | | | | | | | | | | |
| Other | | | | | | | | | | | | | | | |
| Percent Solids, % | | | | 72.3 | 72.8 | 85.6 | 78.5 | 82.0 | 81.0 | 69.2 | 82.9 | 75.3 | 84.6 | 85.6 | 69.2 |
| Total Organic Carbon, % | | | | 1.80 | 2.60 | 0.680 | 1.90 | 1.70 | 1.90 | 3.40 | 1.10 | 4.10 | 0.990 | 5.6 | 0.68 |
| pH | | | · | 6.92 | 7.02 | 8.23 | 8.42 | 8.50 | 7.92 | 7.62 | 8.16 | 7.31 | 7.64 | 8.5 | 5.76 |

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

3. Soil criteria from U.S.EPA, Region 9 Preliminary Remediation Goals (PRGs) for Industrial Soil (October 2002) and from range of background metals concentrations measured in soil found in the eastern United States from NYSDEC Division of Technical and Administrative Guidance Memorandum (TAGM) #4046.

** A New York State background value.

in. bgs =inches below ground surface.

-- = indicates value does not exist.

SB = Site Background

UJ = indicates compound was not detected above the listed detection limit. However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitation necessary to accurately and precisely measure the compound in the sample. J = indicates an estimated value.

U = indicates compound was not detected.

R = indicates data rejected by data validator.

(values) = indicates value reported before rejected.

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TABLE 4-4

ANALYTICAL RESULTS FOR SUBSURFACE SOIL FROM THE INACTIVE LANDFILL AREA

Peter Cooper Site Gowanda, New York

| | 1 | | | | | | Samp | le Location, Ide | ntification, Dep | th and Date Col | lected | | | ···· | 1 | |
|---|-------------|---------------|-------------------|------------------|--------------------------|------------------|------------------|------------------|------------------|------------------|------------------|------------------|-------------------|-----------------------------|----------|-------------|
| | | Soil Criteria | | TP-1 10090025 | TP-2 1009002 4 | TP-3 10090023 | TP-4 10090026 | TP-5 10100028 | TP-6 10100030 | TP-7 10060022 | TP-8 10060021 | TP-9 10060020 | TP-10 10120031 | settling basin 101060029 | | |
| | Eastern USA | Region 9 | Soil Screening | 6.5-7 fbgs | 12.5 fbgs | 8.5-9 fbgs | 7 fbgs | 9.5 fbgs | 5 fbgs | 3-4 fbgs | 4-5 fbgs | 6.5 fbgs | l fbgs | 7 fbgs | Maximum | Minimum |
| Constituent | Background | PRGs | Levels | 10/6/2000 | 10/6/2000 | 10/6/2000 | 10/6/2000 | 10/6/2000 | 10/6/2000 | 10/6/2000 | 10/6/2000 | 10/6/2000 | 10/6/2000 | 10/6/2000 | Conc. | Conc. |
| Volatile Organic Compounds, milligrams per kilogram | | | | | | | | | | | | | | | | |
| Benzene | | 1.3 | 0.03 | 0.021 UJ | 0.014 UJ | 0.017 UJ | 0.003 J | 0.0023 J | 0.025 U | 0.021 UJ | 0.013 UJ | 0.03 UJ | 0.0032 J | 0.0046 J | 0.03 UJ | 0.0023 J |
| Chlorobenzene | | 530 | 1 | 0.021 UJ | 0.014 U | 0.017 UJ | 0.011 UJ | 0.01 UJ | 0.025 U | 0.021 UJ | 0.013 UJ | 0.03 UJ | 0.017 UJ | (0.012) R | 0.03 UJ | (0.012 U) R |
| 1,2-Dichlorohenzene | | 370 | 17 | 0.021 UJ | 0.014 U | 0.017 UJ | 0.011 UJ | 0.01 UJ | 0.025 U | 0.021 UJ | UU E10.0 | 0.03 UJ | 0.017 UJ | (0.012) R | 0.03 UJ | (0.012 U) R |
| 1.4-Dichlorobenzene | | 7.9 | 2 | 0.021 UJ | 0.014 U | 0.017 UJ | U 110.0 | 0.01 UJ | 0.025 U | 0.021 UJ | 0.013 UJ | 0.03 UJ | 0.017 UJ | (0.012) R | 0.03 UJ | (0.012 U) R |
| Ethylbenzene | | 20 | 13 | 0.021 UJ | 0.014 U | 0.017 UJ | 0.011 UJ | 0.01 UJ | 0.025 U | 0.021 UJ | 0.013 UJ | 0.03 UJ | 0.017 UJ | 0.0013 J | 0.03 UJ | 0.0013 J |
| Toluene | | 520 | 12 | 0.031 J | 0.0015 J | 0.0069 J | 0.005 J | 0.0039 J | 0.029 | 0.021 UJ | 0.013 UJ | 0.03 UJ | 0.0074 J | 0.0082 J | 0.031 J | 0.0015 J |
| m-/p-Xylene | | 420 | 210 | 0.021 UJ | 0.014 U | 0.017 UJ | 0.004 J | 0.0023 J | 0.025 U | 0.021 UJ | 0.013 UJ | 0.03 UI | 0.005 J | 0.0063 J | 0.03 UJ | 0.0023 J |
| o-Xylene | | 420 | 210 | 0.021 UJ | 0.014 U | 0.017 UJ | 0.011 UJ | 0.01 UJ | 0.025 U | 0.021 UJ | 0.013 UJ | 0.03 UJ | 0.017 UJ | 0.0019 J | 0.03 UI | 0.0019 J |
| Metals, milligrams per kilogram | | | | | | | | | | | | | | | İ | |
| Arsenic | 3-12** | 1.6 | 29.0 | 13.5 | 9.1 | 60.5 | 4.3 | 6.5 | 29.8 | 58.4 | 294 | 22.1 | 67.1 U | 9.8 | 67.IU | 4.3 |
| Chromium | 1.5-40** | 210 | 38 | 270 | 9.1 | - 137 | 10.3 J | 15.3 | 149 | .623 | 55 | 7.9 | 8610 U | 12.5 | 8619 U | 7.9 |
| Hexavalent Chromium | | 64 | 38 | 5.41 U | 4.76 U | 5.87 U | 4.78 U | 4.55 U | 5.49 U | 4.83 U | 4.98 U | 4.6 U | 5.58 UJ | 4.85 U | 5.87 U | 4.55 U |
| Zinc | 9-50 | 100,000 | 12,000 | 277 | 58.6 | 214 | 57.3 | 70.2 | · 1390 | <u>77.9</u> J | 58:6 J | 99 J | 445 | 68.6 | 1390 | 57.3 |
| Others | | | | | | | | | | | | | | | | |
| Percent Solids | | | | 74.0 | 84.1 | 68.2 | 83.7 | 88.0 | 72.8 | 82.8 | 80.4 | 86.9 | 71.7 | 82.5 | 88 | 68.2 |
| pН | | | | 8.05 | 7.88 | 6.61 | 8.35 | 8.10 | 9.24 | 7.31 | 8.43 | 7.47 | 7.80 | 8.05 | 9.24 | 6.61 |
| Total Organic Carbon | í | [| 1 | 0.900 | 0.570 | 2.00 | 0.1 UI | 0.330 | 2.80 | 0.620 | 1.00 | 0.220 | 1.50 | 0.390 | 2.8 | 0.1 UJ |

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

3. Soil criteria from U.S.EPA, Region 9 Preliminary Remediation Goals (PRGs) for Industrial Soil (October 2002) and from range of background metals concentrations measured in soil found in the

eastern United States from NYSDEC Division of Technical and Administrative Guidance Memorandum (TAGM) #4046.

** A New York State Background value

fbgs = feet below ground surface

SB = Site Background

-- = indicates value does not exist.

J = indicates an estimated value,

U = indicates compound was not detected above the listed detection limit.

R= indicates data rejected by data validator.

UJ = indicates compound was not detected above the listed detection limit.

However, the reported quantitation limit is approximate and may or may

not represent the actual limit of quantitiation necessary to accurately

and precisely measure the compound in the sample.

(value) = indicates value reported before rejected.

ISDivisionDS771 PRP Group Deer Cooper NPI SRI equalsENAL REPERT (November 2003 SideminiATables (Eind/STable 4.1, 2, & d.4 Iund/El subsorface soils EDAL

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ANALYTICAL RESULTS FOR SURFACE SOIL SAMPLES FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Site Gowanda, New York

| | | | | Γ | | | Sample Loca | tion, Identificatio | n, Depth, and D | ate Collected | | | | r | |
|-----------------------------|-------------|---------------------------|-----------|------------|-----------|---------------|-------------|---------------------|-----------------|---------------|-----------|-----------|--------------|---------|-----------|
| 1 | | | | SB-1 | SB-2 | SB-4 | SB-5 | SR-7 | SB-8 | SB-9 | SB-10 | MWFP-2 | MWFP-3 | 1 | |
| 1 | l sa | oil Criteria ^J | | 1005000006 | 100500008 | 100500010 | 100600012 | 100600018 | 100600032 | 100600034 | 100600036 | 100600015 | 100900018 | | |
| | | Re | eion 9 | 0-2' | 0-2' | 0-2' | 0-2' | 0-2' | 0-2' | 0.5-2.5' | 0-2' | 0.5-2.5' | 0.5-2.5 | Maximum | Minimum |
| | Eastern USA | | Sail | 1 | | | | | | 1 | | | | | |
| | Rackground | | Screening | | | | | | 1 | | | | | | |
| Constituent ² | | PRGs | Levels | 10/05/00 | 10/05/00 | 10/05/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/09/00 | Conc. | Conc. |
| Volatile Organic Compounds, | Î. | 1 | | [| | | | 1 | | 1 | | | 1 | 1 | |
| milligrams per kilogram | 1 | | l | | | | | | l | | | | | | |
| Acetone | | 6000 | 16 | 0.045 J | 0.053 J | 0.12 | 0.056 | 0.053 J | 0.058 J | 1.4 J | 0.21 J | 0.056 J | 1.4 U | 1.4 J | 0.045 J |
| Benzene | | 1.3 | 0.03 | 0.0027 J | 0.0021 J | 0.01 UJ | 0.0036 J | 0.0025 J | 0.0016 J | 0.0023 J | 0.0082 J | 0.0076 J | 1.4 U | 1.4 U | 0.0016 J |
| Bromodichloromethane | | 1.8 | 0.6 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | U 10.0 | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| Bromoform | | 220 | 0.8 | 0.0094 UJ | 0.011 UJ | LU 10.0 | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| Bromomethane | | 13 | 0.2 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 UJ | 1.4 UJ | 0.0094 UJ |
| 2-Butanone (MEK) | | 27000 | | 0.0088 J | 0.0094 J | 0.018 J | 0.011 | 0.0071 J | 0.0057 J | 0.28 J | 0.017 J | 0.022 UJ | 1.4 U | 1.4 U | 0.0057 J |
| Methyl tert-Butyl Ether | | 160 | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| Carbon Disulfide | | 720 | 32 | 0.013 J | 0.0023 J | 0.0072 J | 0.01 | 0.072 J | 0.015 J | 0.0031 J | 0.016 UJ | 0.01 J | 1.4 U | 1.4 U | 0.0023 J |
| Carbon Tetrachloride | | 0.55 | 0.07 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 10× | 10 | 0.0094 UJ |
| Chlorobenzene | Ι | 530 | 1 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | I.4 U | 1.4 U | 0.0094 UJ |
| Chloroethane | | 6.5 | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| Chloroform | 1 | 12 | 0.6 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01.03 | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 5.7 | 5.7 | 0.0094 UJ |
| Chloromethane | | 2.6 | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| 1.2-Dibromo-3-Chloropropane | | 2 | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| Cyclohexane | | 140 | | 0.013 J | 0.011 J | 0.01 UJ | 0.0058 J | 0.0065 J | 0.003 J | 0.0036 J | 0.016 UJ | 0.0095 J | 0.47 J | 0.47 J | 0.003 J |
| Dibromochloromethane | | 2.6 | 0.4 | 0.0094 UI | 0.011 UJ | 0.01 UJ | 10.01 | 0.0099 UI | 0.0094 UJ | 0.014 UJ | 0.016 UI | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| 1.2-Dibrornocthane | | 0.028 | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| 1,2-Dichlorohenzene | | 370 | 17 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UI | 0.022 UJ | <u>1.4 U</u> | 1.4 U | 0.0094 UJ |
| 1.4-Dichlorobenzene | | 7.9 | 2 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 111 | 0.0099 UJ | 0.0094 UJ | 0.0017 J | 0.016 UJ | 0.022 UJ | <u>1.4 U</u> | 1.4 U | 0.0017 J |
| 1,3-Dichlorobenzene | | 63 | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.00941)J |
| Dichlorodifluoromethane | | 310 | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| 1.1-Dichloroethane | | 1700 | 23 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | U 10.0 | UJ 9900.0 | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 0.16 J | 0.16 J | 0.0094 UJ |
| 1.2-Dichloroethane | | 0.6 | 0.02 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 0.24 J | 0.24 J | 0.0094 UJ |
| 1,1-Dichloroethene | | 410 | 0.06 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| trans-1,2-Dichloroethene | L | 230 | 0.7 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| cis-1,2-Dichloroethene | | 150 | 0.4 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 0.26 J | 0.26 J | 0.0094 U1 |
| 1.2-Dichloropropane | | 0.71 | 0.03 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 1/J | 1.4 U | 1.4 U | 0.0094 UJ |
| trans-1.3-Dichloropropene | I | 1.8 | 0.004 | 0.0094 UJ | 0.011 UJ | 0.01 UI | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | <u>1.4 U</u> | 1.4 U | 0.0094 UJ |
| cis-1,3-Dichloropropene | | 1.8 | 0.004 | 0.0094 UJ | tU 110.0 | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UI |
| Ethylbenzene | | 20 | 13 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UI | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| 2-Hexanone | | - | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.064 J | 0.025 J | 0.016 1)J | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| Isopropylbenzene | <u> </u> | 2000 | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| Methyl Acetate | | 92000 | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.0048 J | 0.21 J | 0.022 UJ | 1.4 U | 1.4 U | 0.0048 J |
| Methylcyclohexane | | 8700 | | 0.022 J | 0.014 J | U 10.0 | 0.0096 J | 0.01 J | 0.0042 J | 0.0023 J | 0.0021 J | 0.015 J | 1.6 | 1.6 | 0.0021 J |
| Methylene Chloride | l | 21 | 0.02 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 U1 | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| 4-Methyl-2-Pentanone | | | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.005 J | 0.039 J | 0.022 UJ | 1.4 U | 1.4 U | 0.005 J |
| Styrene | | 1700 | 4 | 0.0094 UJ | U 110.0 | 0.01 UJ | U 10.0 | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0 0094 UJ |
| 1.1.2,2-Tetrachloroethane | | 0.93 | 0.003 | 0.0094 UJ | 0.011 UJ | <u>U 10.0</u> | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| Tetrachloroethene | 1 | 3.4 | 0.06 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0031 J | 0.0094 UJ | 0.0097 J | 0.18 J | 0.022 UJ | 54 ··· | 54 | 0.00311 |
| Toluene | I | 520 | 12 | 0.005 J | 0.0032 J | · 0.01 UJ | 0.0061 J | 0.0046 J | 0.0032 J | 0.0023 J | 0.019.1 | 0.015.1 | 0.38 J | 0.38 J | 0.0023 J |
| 1,2,4-Trichlorobenzene | I | 3000 | 5 | 0.0094 UJ | 0.011 UJ | 0.01 UI | 0.01 UI | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 0.36 1 | 0.36 J | 0.0094 UJ |
| 1.1.1-Trichloroethane | I | 1200 | 2 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.0064 J | 0.022 ()] | \$.5 | 5.5 | 0.0064 J |
| 1.1.2-Trichloroethane | • | 1.6 | 0.02 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | LU 10.0 | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |



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TABLE 4-5

ANALYTICAL RESULTS FOR SURFACE SOIL SAMPLES FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Site Gowanda, New York

| | | | | | | | Sample Locat | ion, Identificatio | n, Depth, and De | ate Collected | | | | | |
|---------------------------------------|-------------|--------------------------|-------------|------------|---------------------|-----------|--------------|--------------------|------------------|---------------|-----------|-----------|-----------|---------|-----------|
| | 1 | | | SB-1 | SB-2 | SB-4 | SB-5 | SB-7 | SB-8 | SB-9 | SB-10 | MWFP-2 | MWFP-3 | | |
| | So So | il Criteria ⁾ | | 1005000006 | 100500008 | 100500010 | 100600012 | 100600018 | 100600032 | 100600034 | 100600036 | 100600015 | 100900038 | | |
| | | Reg | tion 9 | 0-2' | 0-2' | 0-2' | 0-2' | 0-2' | 0-2' | 0.5-2.5' | 0-2' | 0.5-2.5' | 0.5-2.5 | Maximum | Minimum |
| | Eastern USA | | Soil | |] | | | | | | | í i | | | |
| 1 | Background | | Screening | | | | | Ì | l | | | | | | |
| Constituent ² | | PRGs | Levels | 10/05/00 | 10/05/00 | 10/05/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/09/00 | Conc. | Солс. |
| Trichloroethene | | 0.1 | 0.06 | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UI | 0.022 UJ | 0.51 1 | 0.51 J | 0.0094 UJ |
| Trichforofluoromethane | | 2000 | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| 1,1,2-Trichloro-1,2,2-Trifluoroethane | | 5600 | | 0.0094 UJ | 0.011 UJ | 0.01 UJ | 0.01 UJ | 0.0099 UI | 0.0094 UJ | 0.014 UJ | 0.016 111 | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| Vinyl Chloride | | 0.75 | | 0.0094 UJ | 10.011 03 | 0.01 UJ | 0.01 UJ | 0.0099 UJ | 0.0094 UJ | 0.014 UJ | 0.016 UJ | 0.022 UJ | 1.4 U | 1.4 U | 0.0094 UJ |
| m-/p-Xylene | | 420 | 210 | 0.0045 (JJ | 0.0031 UJ | 0.01 UJ | 0.0036 J | 0.0048 J | 0.0033 1 | 0.003 J | 0.0044 J | 0.0071 J | 0.52 1 | 0.52 J | 0.003 J |
| o-Xylene | | 420 | 210 | 0.0014 UI | 0.0011 UJ | 0.01 UJ | 0.0011 J | 0.00151 | 0.00098 J | 0.014 UJ | 0.004 J | 0.0039 J | 0.42 J | 0.42 J | 0.00098 J |
| | | | | | | | | | | | | ļ | | | |
| Semi-Volatile Organic Compounds, | [| | | | | | | · · · | | | | | | | |
| milligrams per kilogram | | | | | | | | | · | | | | | | |
| Acenaphthene | | 29000 | 570 | 0.37 U | 1.8 J | 0.16 J | 0.36 U | 2.6 J | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 2.63 | 0.161 |
| Acenaphthylene | | | | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 0.4] | 0.38 U | 0.41 U | 0.39 U | 0.29 J | 0.38 U | 0.41 U | 0.29 J |
| Acetophenone | L | | | 0.37 0 | 3.9 U | 0.4 U | 0.36 0 | 3.90 | 0.38 0 | 0.41 U | 0.39 0 | 0.42 U | 0.38 0 | 3.90 | 0.36 U |
| Anthracene | | 100000 | 12000 | 0.04 J | 5.9 | 0.47 J | 0.36 U | 14 | 0.38 0 | 0.044] | 0.055 J | 0.24] | 0.049 J | 14 | 0.04 J |
| Atrazine | Į | 1.8 | | 0.37 0 | 3.90 | 0.4 0 | 0.36 U | 3.90 | 0.38 0 | 0.41 0 | 0.39 0 | 0.42 0 | 0.38 0 | 3.90 | 0.36 0 |
| Benzaldehyde | | 62000 | <u> </u> | 0.370 | 3.9 U | 0.4 0 | 0.36 0 | 3.90 | 0.38 0 | 0.41 0 | 0.39 0 | 0.42 0 | 0.38 0 | 3.90 | 0.36 0 |
| Benzo(a)anthracene | | 2.1 | 2 | 0.16 1 | - 10 | 1.6 J | 0.36 () | 20 | 0.38 0 | 0.111 | 0.237 | 0.4/ | 0.20 1 | 24 | 0.011 |
| Benzo(a)pyrene | | - 0.21 | <u> </u> | 0.10 1 | | | 0.30 0 | | 0.38 0 | 0.0871 | 0.23 3 | 0.11 | 0.27.1 | - 20 | 0.087 J |
| Benzo(b)Huoraninche | | 2.1 | | 0.14 J | 0.4 | | 0.36 U | | 0.38 0 | 0.079 1 | 0.24 J | 0.3 1 | 0.23 1 | | 0.0793 |
| Denzo(k)(fugraethage | { | | 40 | 0.11 J | 7.2 | | 0.36 U | 10 | 0.38 0 | 0.41 0 | 0.24 J | 0.351 | 0.21 1 | 14 | 0.001 |
| L 1-Biobenyl | | 150 | - 47 | 0.14 3 | 301 | 0411 | 0.36 U | 10 | 0.38 U | 0.07 J | 0.23 1 | 0.381 | 0.25 5 | 10 | 0.36.11 |
| Butyl Bennyl Dathalate | | 100000 | 010 | 0.37 111 | 1011 | 0.4 0 | 0.36.0 | 3.90 | 0.38 0 | 0.41 0 | 0.37 0 | 0.42.0 | 0.38 U | 3911 | 0.16 U |
| di-N-Butylobthalate | | 100000 | | 0.37 11 | 3.90 | 0.40 | 0.36 U | 3.70 | 0.38 0 | 04111 | 0.39 0 | 0.42 0 | 0.38 11 | 3911 | 0.36 U |
| Caprolactam | | 100000 | | 0.37 11 | 391 | 0.4 U | 0.36 [] | 3911 | 0.38 0 | 0411 | 0.39.11 | 0.42 U | 0 38 U | 1.911 | 0.36 11 |
| Carbazole | | 86 | 0.6 | 0.37 U | 231 | 0.391 | 0.36.11 | 351 | 0.38 U | 0.41 0 | 0.044 1 | 0.42.0 | 0.18 1/ | 3.51 | 0.044 1 |
| Indeno(1.2.3-cd)nytene | | - 21 | 14 | 0.084 1 | 4.0 | 0.921 | 0.36 (/ | 13 | 0 18 11 | 0.043.1 | 0.19.1 | 0.271 | 0.19.1 | 13 | 0.043 J |
| 4-Chloroaniline | | 2500 | 0.7 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| bis(2-chloroethoxy)methane | | | | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9U | 0.36 U |
| bis(2-chloroethyl)ether | | 0.55 | 0.0004 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 2-Chloronaphthalene | | | | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 2-Chlorophenol | | 240 | 4 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 2,2-oxybis(1-chloropropane) | | | | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| Сhrysene | | 210 | 160 | 0.17 J | 9.3 | 1.7 5 | 0.36 U | 22 | 0.38 U | 0.14 J | 0.36 J | 0.6 | 0.29 J | 22 | 0.14 J |
| Dibenzo(a,h)anthracene | | 0.21 | 2 | 0.37 U | a ²¹ 191 | 0.35 J | 0.36 U | 5.2 | 0.38 Ū | 0.41 U | 0.076 J | 0.13 J | 0.078 1 | 5.2 | 0.076 J |
| Dibenzofuran | | 3100 | | 0.37 U | 1.1.7 | 0.12 J | 0.36 U | 2.2 J | 0.38 U | 0.41 U | 0.055 J | 0.42 U | 0.38 U | 2.2 J | 0.055 J |
| 3,3-Dichlorobenzidine | | 3.8 | 0.007 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | R | 3.9 U | 0.36 U |
| 2,4-Dichlorophenol | | 1800 | i | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| Diethylphthalate | | 100000 | | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| Dimethyl Phthalate | | 100000 | | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 2.4-Dimethylphenol | | 12000 | 9 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.910 | 0.36 U |
| 2.4-Dinitrophenol | | 1200 | 0.3 | 0.93 U | 9.8 U | 1.0 | 0.92 U | 9.8 U | 0.95 U | I U | 0.98 U | 1.1 U | 0.97 U | 3.9 () | 0.36 U |
| 2,4-Dinitrotolucne | | 1200 | 0.0008 | 0.37 U | <u>3.9 U</u> | 0.4 U | 0.36 Ú | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 2,6-Dinitrotoluene | | 620 | 0.0007 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| his(2-Ethylhexyl)phthalate | | 120 | | 0.068 1 | 3.9 U | 0.069 J | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.91/ | 0.068 J |
| Fluoranihene | | 22,000 | 4300 | 0.31 J | 23 | 3.8 J | 0.36 U | 60 | 0.38 U | 0.26 J | 0.51 | 0.62 | 0.41 | 60 | 0.26 J |

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TABLE 4-5

Page 3 of 4

ANALYTICAL RESULTS FOR SURFACE SOIL SAMPLES FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Site Gowanda, New York

| | | | | | | | Sample Locat | ion, Identificatio | n, Depth, and Do | ate Collected | | | | | |
|---------------------------------|---------------|---------------|-----------|------------|-----------|-----------|--------------|--------------------|------------------|---------------|-----------|-----------|-----------|---------|---------|
| | | | | SB-1 | SB-2 | SB-4 | SB-5 | SB-7 | SB-8 | SB-9 | SB-10 | MWFP-2 | MWFP-3 | | |
| | Sa | il Criteria) | | 1005000006 | 100500008 | 100500010 | 100600012 | 100600018 | 100600032 | 100600034 | 100600036 | 100600015 | 100900038 | | |
| | | Rej | rion 9 | 0-2' | 0-2' | 0-2' | 0-2' | 0-2' | 0-2' | 0.5-2.5' | 0-2' | 0.5-2.5' | 0.5-2.5 | Maximum | Minimum |
| | Eastern USA | | Soil | | | | | | | | | | | | |
| | Background | | Screening | | | | | | | | | | | | |
| Constituent ² | | PRGa | Levels | 10/05/00 | 10/05/00 | 10/05/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/09/00 | Conc. | Conc. |
| Fluorene | 1 | 26,000 | 560 | 0.37 U | 2.3 J | 0.17 J | 0.36 U | 4.2 | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 4.2 | 0.17 J |
| Hexachiorobenzene | | 1.1 | 2 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| Hexachlorobutadiene | | 22 | 2 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| Hexachlorocyclopentadiene | | 3700 | 400 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 1.9 U | 3.9 U | 0.36 U |
| Hexachloroethane | | 120 | 0.5 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| Isophorone | | 1800 | 0.5 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 2-Methylnaphthalene | | | | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.1 J | 0.18 J | 0.083 J | 0.079 J | 3.9 U | 0.079 J |
| 4.6-Dinitro-2-Methylphenol | | | | 0.93 U | 9.8 U | 10 | 0.92 U | 9.8 U | 0.95 U | 1 U | 0.98 U | 1.I U | 0.97 U | 9.8 U | 0.92 U |
| 4-Chloro-3-Methylphenol | | | | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 2-Methylphenol | | | | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 4-Methylphenol | 1 | 3100 | | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.059 J | 0.42 U | 0.38 U | 3.9 U | 0.059 J |
| Naphthalene | | 190 | 84 | 0.37 U | 3.9 U | 0.051 J | 0.36 U | 3.9 U | 0.38 U | 0.069 J | U 11.0 | 0.044 J | 0.047 J | 3.9 U | 0.044 J |
| 2-Nitroaniline | | 18 | | 0.93 U | 9.8 U | 10 | 0.92 U | 9.8 U | 0.95 U | 10 | 0.98 U | 1.1 U | 0.97 U | 9.8 U | 0.92 U |
| 4-Nitroaniline | | | | 0.93 U | 9.8 U | 10 | 0.92 U | 9.8 U | 0.95 U | 10 | 0.98 U | 1.1 U | 0.97 U | 9.8 U | 0.92 U |
| Nitrobenzene | | 100 | 0.1 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 2-Nitrophenol | | •• | •• | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 4-Nitrophenol | | | | 0.93 U | 9.8 U | 10 | 0.92 U | 9.8 U | 0.95 U | 10 | 0.98 U | L.I U | 0.97 U | 9.8 U | 0.92 U |
| n-Nitrosodiphenylamine | | 350 | 1 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| di-n-Octyl Phthalate | | 25000 | 10000 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| Pentachlorophenol | | 9 | 0.03 | 0.93 U | 9.8 U | 10 | 0.92 U | 9.8 U | 0.95 U | 10 | 0.98 U | 1.I U | 0.97 U | 9.8 U | 0.92 U |
| Phenanthrene | | | | 0.18 J | 21 | 2.4 J | 0.36 U | 45 | 0.38 U | 0.23 J | 0.34 J | 0.3 J | 0.24 J | 45 | 0.18 J |
| Phenol | | 100,000 | 100 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 4-Bromophenyl-Phenylether | | | | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 4-Chlorophenyl-Phenylether | | | | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| n-Nitroso-di-n-Propylamine | | 0.25 | 0.00005 | 0.37 U | 3.9 U | 0.4 U | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| Рутспе | | 29,000 | 4,200 | 0.32 J | 20 | 3.4 J | 0.36 U | 44 | 0.38 U | 0.22 J | 0.49 | 0.86 | 0.43 | 44 | 0.22 J |
| 2,4,6-Trichlorophenol | | 62 | 0.2 | 0.37 U | 3.9 U | 0.4 Ú | 0.36 U | 3.9 U | 0.38 U | 0.41 U | 0.39 U | 0.42 U | 0.38 U | 3.9 U | 0.36 U |
| 2,4,5-Trichlorophenol | | 62000 | 270 | 0.93 U | 9.8 U | 10 | 0.92 U | 9.8 U | 0.95 U | 10 | 0.98 U | 1.1 U | 0.97 U | 9.8 U | 0.92 U |
| Metals, milligrams per kllogram | | | | | | | | | | | | | | | |
| Aluminum | 33,000 | 100,000 | [| 6210 | 5440 | 7570 | 8000 | 8280 | 6810 | 2010 | 4220 | 6490 | 5190 | 8280 | 2010 |
| Antimony | ~~ | 410 | 5 | 6.5 UJ | 6.9 UJ | 7.2 UJ | 6.5 UJ | 6.9 UJ | 6.8 UJ | 7.2 UJ | 7 UJ | 7.6 UJ | 7 UJ | 7.6 UJ | 6.5 UJ |
| Arsenic | 3-12** | 1.6 | 29.0 | 8.5 | 168 3 | 10.7 | 8 | 9,5 | 6.6 | 16.2 | 16.2 | 729.9 | 22.7 | 168 | 6.6 |
| Barium, | 15-600 | 67,000 | 1.600 | 72.8 | 65.1 | 80.2 | 58.4 | 92.8 | 63.9 | 68.7 | 59.2 | 64.2 | 117 | 117 | 58.4 |
| Beryllium | 0-1.75 | 1.900 | 63 | 0.54 U | 0.58 U | 0.6 U | 0.54 U | 0.57 U | 0.57 U | 0.6 U | 0.58 U | 0.87 | 0.64 | 0.87 | 0.54 U |
| Cadmium | 0.1-1 | 450 | 8 | 0.54 U | 0.58 U | 0.6 U | 0.54 U | 0.57.11 | 0.57.11 | 0.6 U | 0.58 U | 0.64 U | 1.6 | 1.6 | 0.54 U |
| Calcium | 130-35,000** | | | 29000 | 6880 | 23800 | 44200 | 30200 | 33600 | 1050 | 1870 | 2490 J | 8210 | 44200 | 1050 |
| Chromium | 1.5-40** | 210 | 38 | 34.2 | 59.7 | 18.2 | 10.8 | 33.3 | | 59.3 | 54.5 | 198.1 | 52.5 | 198 | 91 |
| Cobalt | 2.5-60** | 1,900 | | 6.4 | 6.9 | 8.2 | 7.7 | 7.2 | 6.6 | 6 U | 6.6 | 7,1 | 7.6 | 8.2 | 6 U |
| Copper | 1.50 | 41,000 | | 26.6 | 37 | 43.6 | 21 | 73.3 | 20.9 | 56.7 | 30.7 | 29.3 | 171 | 171 | 20.9 |
| Hexavalent Chromium | | 64 | 38 | 4.46 UJ | 4.74 UJ | 4.9 U1 | 4.42 UJ | 4.71 UI | 4.57 UI | 4,92 UJ | 4.74 UJ | 5.08 UI | 4.66 U | 5.08 UJ | 4.42 UJ |
| tron | 2,000-550,000 | 100,000 | | 18200 | 18900 | 23000 | 16900 | 12600 | 15300 | 31300 | 18500 | 18900 | 30100 J | 31,300 | 12600 |
| Lead | 4-61*** | 750 | | 50.11 | 79.4 3 | 169.1 | 8.2 1 | 74.2 1 | 8.2.1 | 193.1 | 269 1 | 41 5 | 202 | 2691 | 8.23 |
| Magnesium | 100-5,000 | | | 4470 | 3130 | 6260 | 12600 | 5740 | 9300 | 225 | 1520 | 1730 | 2270 | 12600 | 225 |
| Manganese | 50-5,000 | 19,000 | | 332 | 251 | 449 | 489 | 451 | 469 | 64.7 | 132 | 160 | 314 | 489 | 64.7 |



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ANALYTICAL RESULTS FOR SURFACE SOIL SAMPLES FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Site Gowanda, New York

| | T | | | | | | Sample Locat | ion, Identificatio | n. Depth, and D | ate Collected | | | | | |
|--------------------------|----------------|--------------------------|-----------|------------|-----------|-----------|--------------|--------------------|-----------------|---------------|-----------|-----------|-----------|---------|---------|
| | ļ | | | SB-1 | SB-2 | SB-4 | SB-5 | SB-7 | SR-8 | SR-9 | SB-10 | MWFP-2 | MWFP-3 | | |
| | Sa | il Criteria ³ | | 1005000006 | 100500008 | 100500010 | 100600012 | 100600018 | 100600032 | 100600034 | 100600036 | 100600015 | 100900038 | | |
| | | Ree | tion 9 | 0-2' | 0-2' | 0-2' | 0-2' | 0-2' | 0-2' | 0.5-2.5' | 0-2' | 0.5-2.5' | 0.5-2.5 | Maximum | Minimum |
| | Eastern USA | ° | Soil | | | 1 | | | | | | | | | |
| | Background | | Screening | | | | | | | | | | | | |
| Constituent ² | | PRGs | Levels | 10/05/00 | 10/05/00 | 10/05/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/09/00 | Conc. | Conc. |
| Mercury | 0.001-0.2 | 310 | | 0.08 | 0.13 | 0.13 | 0.05 U | 0.17 | 0.06 U | 3.1 | 0.47 | 0.16 J | 0.4 113 | 3.1 | ND |
| Nickel | 0.5-25 | 20,000 | 130 | 17.8 | 17.9 | 19.4 | 18 | 19.1 | 15.9 | 13.1 | 14.7 | 179 | 27.2 | 27.2 | 13.1 |
| Potassium | 8,500-43,900** | | | 951 | 527 | 805 | 1060 | 912 | 755 | 239 U | 399 | 542 | 622 | 1060 | 239 U |
| Sclenium | 0.1-3.9 | 5,100 | 5 | 1.6 | 1.8 | 1.3 | 1.4 | 0.95 | 1.6 | 1.7 | 2 | 2.7 | 2.1 J | 2.7 | 0.95 |
| Silver | | 5100 | 34 | 1.I U | 1.2 U | 1.2 U | 1.1 U | 1.1 U | 1.I U | 1.2 U | 1.2 U | 1.3 U | 1.2 U | 1.3 U | 1.1 U |
| Sodium | 6,000-8,000 | | | 377 | 372 | 439 | 425 | 479 | 389 | 398 | 458 | 411 | 514 | 514 | 372 |
| Thallium | | 67 | | 1.1 0 | I.I U | 1.2 U | 1.1 U | 1.2 U | 1.I U | 1.2 U | 1.2 U | 1.3 U | 1.I Ŭ | 1.3 U | 1.1 U |
| Vanadium | 1-300 | 7,200 | 6,000 | 14.8 | 12.8 | 16.7 | 18 | 17.7 | 14.6 | 17.5 | 17.8 | 15.3 | 20.2 | 20.2 | 12.8 |
| Zinc | 9-50 | 100,000 | 12,000 | 152 J | 109 J | 132 J | 45.6 J | 124 J | 51.67 | 161 | 728 3 | 84.61 | 246 | 728 | 45.6 J |
| | | • | | | | | | | | | | | | | |
| Others | | | | | | | | | | | | | | | |
| Percent Solids, % | | | | 89.7 | 84.3 | 81.7 | 90.6 | 84.9 | 87.6 | 81.3 | 84.4 | 78.7 | 85.8 | 90.6 | 78.7 |
| pH | | | | 8.01 | 8.2 | 8.34 | 7.85 | 8.24 | 7.81 | 7.34 | 7.61 | 7.7 | 7.46 | 8.34 | 7.34 |
| TOC, % | | | | 0.47 | 0.94 | 1.8 | 0.25 | 1.3 | 0.35 | 1.1 | 1.7 | 1.7 | 1.5 | 1.8 | 0.25 |

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services. The analytical results for the SVOC, 3-Nitroaniline, was rejected during data validation for each sample.

3. Soil criteria from U.S.EPA, 2000 Region 9 Preliminary Remediation Goals (PRGs) for Industrial Soil (October 2002) and from range of background metals concentrations measured in soil found in the

eastern United States from NYSDEC Division of Technical and Administrative Guidance Memorandum (TAGM) #4046.

** A New York State Background value

*** Background levels for lead vary widely, average levels in undeveloped, rural areas range from 4-61 ppm while metropolitan/suburban areas range from 200-500 ppm.

J = indicates a laboratory estimated value or estimated as a result of data validation.

U = indicates compound was not detected at or above the listed detection limit.

R= indicates data rejected by data validator.

UJ = indicates compound was not detected above the listed detection limit. However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitiation necessary to accurately

-- = indicates value does not exist. fbgs = feet below ground surface indicates concentration above soil criteria.

SB = Site Background

and precisely measure the compound in the sample.

201771 PRP Group Pres Conper NPL/R1 report/FINAL REPORT (November 2003 Solonistal #Tobles (Famil/KToble 4-5 FINAL

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TABLE 4-6

ANALYTICAL RESULTS FOR SUBSURFACE SOILS FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Site Gowanda, New York

| l | r | | | T | | | | Semale Locat | on Identificatio | on Death and D | ate Collected | | | | | r | |
|--|-------------|--------------------------|-----------|-----------|------------|------------|-----------|--------------|------------------|----------------|---------------|-----------|-----------|-----------|-----------|-----------|-----------|
| 1 | 1 | | | 58-1 | 58-2 | 58.1 | 58.4 | Sample Local | SR.6 | SR.7 | CR.R | 58.0 | 58.10 | MWEP.1 | MWEP. | 1 | |
| | So | il Criteria ³ | | 100500007 | 100500009 | 100600014 | 100500011 | 100600013 | 100900040 | 100600019 | 100600011 | 100600035 | 100600037 | 100600017 | 100900039 | | |
| | Eastern USA | Region 9 | Soil | 5.7 | 6.8 | 2.5' | 44 | | 4.6 | 7.0' | 10.12 | 7.0' | 7 9' | 5.7' | 6.7' | Marinna | Minimum |
| Construction 2 | | | Screening | | | | 4-0 | 0-4 | •-0 | / | 10-12 | 7.9 | / | 3-7 | 3-7 | MALIMAM | MINIMAM. |
| Constantena | Background | PRGA | Levels | 10/05/00 | 10/05/00 | 10/06/00 | 10/05/00 | 10/06/00 | 10/9/2000 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/09/00 | Conc. | Conc. |
| Volatile Organic Compounds, milligrams | | | 1 | | | | | | | | | 1 | | | | | |
| Acatuma | | 6000 | 16 | 0.07.1 | 0.011 | 0.080.1 | | | | 0.01 | | | 00/51 | - 0.12 | 0.0000 1 | | 0.010.1 |
| Henzene | f | 1 13 | 0.03 | 0.07 | 0.0088.11 | 0.00971/1 | 0.029 1 | 0.491 | 0.147 | 0.067 | 0.0423 | 0.004 | 0.0031 | 0.12 | 0.098 1 | 0.031 111 | 0.0241 |
| Bromodichloromethane | <u> </u> | 1.5 | 0.6 | 0.01510 | 0.0088 U | 0.0092 1/1 | 001910 | 0.031 11 | 0.013 (0 | 001710 | 0.0006 11 | 0.01331 | 0.011111 | 0.014 11 | 0.011.00 | 0.031 11 | 0.008817 |
| Bromoform | t | 220 | 0.8 | 0.015 UJ | 0.0088 U | 0 0092 UJ | 0.01910 | 0031111 | 001301 | 0.017 10 | 0.0096 UI | 0.014 0 | 001110 | 0.014 U | 001111 | 0.01111 | 0.008811 |
| Bromomethane | | 13 | 0.2 | 0.015 UJ | 0.0088 U | 0.0092 UJ | 0.019 111 | 0.031 10 | 0.013 00 | 0.017 10 | 0.0096 UI | 0.014 1/ | 0 011 (1) | 0.014 U | 0.011 1// | 0 031 10 | 0.0088.0 |
| 2-Butanone (MEK) | | 27000 | - | 0.0096 J | 0.0088 U | 0.011 J | 0.00691 | 0.12.1 | 0.026 J | 0.01 J | 0 0076 1 | 0.013.1 | 00141 | 0.031 | 0.016 J | 0.12 J | 0.00691 |
| Methyl tert-Butyl Ether | | 160 | | 0.015 UJ | 0.0088 U | 0.0092 ()] | 0.019 UJ | 0.031 UJ | 0.013 (U | 0.017 UJ | 0.0096 UJ | 0.014 U | 0.011 UJ | 0.014 U | 0.011 ()) | 0.031 UJ | 0.0088 U |
| Carbon Disulfide | | 720 | 32 | 0.0021 J | 0.0088 U | 0.0092 UJ | 0.00551 | 0.031 1// | 0.024 J | 0.017 UJ | 0.0044 J | 0.032 | 0.014 J | 0.014 U | 0.011 UJ | 0.032 | 0.0021 J |
| Carbon Tetrachluride | | 0.55 | 0.07 | 0.008 J | 0 0088 U | 0.0092 UJ | 0.019 (1) | 0.031 UJ | 0.013 UJ | 0.012 J | 0.0096 UJ | 0.014 U | 001110 | 0.014 U | 0 025 J | 0.031 UJ | 0.008 J |
| Chlorobenzene | | 530 | 1 | 0.015 UJ | 0.0088-11 | 0.0092 1/J | 0.019 (1) | 0 031 UJ | 0.013 ()) | 0.017 (1) | 0.0096 111 | 0.014 U | 0.01110 | 0014 U | 0.011 UJ | 0.031 1/J | 0.008810 |
| Chloroethane | | 6.5 | | 0.015 UJ | 0.0088 U | 0.0092 UJ | 0.019 UI | 0.031 UJ | 0.013 UJ | 0.017 ()] | 0.0096 UJ | 0.014 17 | 0.011 []] | 0.014 U | 0.011 1/ | 0.031.13 | 0.0088 U |
| Chloroform | | 12 | 0.6 | 0.015 UJ | 0.0088 U | 0.0092 (1) | 0.019 ()) | 0.031 UJ | 0.013 UJ | 0.0067 J | 0.0096 UJ | 0.014 1/ | 0.011 UJ | 0.014 U | 0.0081 J | 0.031.10 | 0 0067 J |
| Chloromethane | | 2.6 | | 0.015 10 | 0.0088 U | 0.0092 (J) | 0.019 [1] | 0.031 UJ | 0.013 UJ | 0.017 U | 0.0096 UJ | 0.014 1/ | 0.011 UI | 0014 U | 0.011 (1) | 0.031 UJ | 0.008811 |
| 1.2-Dibromo-3-Chloropropane | | 2 | | 0015 0 | 0.0088 U | 0.0092 UJ | 0.019 (1) | 003110 | 0.013 UJ | 0.017 UJ | 0 0096 (1) | 0.014 13 | 0.011 UJ | 0.014 UJ | 0.011 ()) | 0.031 ()) | 0.0088 U |
| Cyclohexane | | 140 | | 0.015 () | 0.0088 U | 0.0092.03 | 00191/ | 0.031 U | 0.013 UJ | 0.0047 J | 0.0096 (1) | 0.0095 J | 0.0035 J | 0014 U | 0.011 () | 0.031 ()) | 0.0035 J |
| 1.2 15k-semicitare | | 2.0 | U.4 | 0.015 07 | 0.0088 U | 0 0092 03 | 0.01910 | 0031 () | 0.013 ()) | 001710 | 0.0096 UI | 0.014 () | 0.011 UJ | 00140 | 0.011 00 | 0.031 (J) | 0.0088.0 |
| 1,2-Dichlorobenzene | | 170 | | 001510 | 0.0088 U | 0.0092 UJ | 0.019 (0 | 0.0.31 10 | 001300 | 001700 | 0.0096 00 | 0.014 U | 0.011 (0 | 00140 | | 0.031 (3 | 0.0088.0 |
| 1,2-Dichlorobenzene | | 3/0 | <u> </u> | 0.015 UI | 0.0088 11 | 0.0002.00 | 0.019 00 | 003110 | 001310 | 0.017 00 | 0.00% U | 0.014 10 | | 0.014 () | 0.011 1/1 | 0.031 00 | 0.008911 |
| 1.3-Dichlorobenzene | | 63 | | 0.015 111 | 0.0088.0 | 0.009210 | 0.019 11 | 003110 | 0.013 03 | 0.017 () | 0.000610 | 0.014 ()) | 0.01110 | 0.014 01 | 0.01110 | 003110 | 0.008811 |
| Dichlorodifluoromethane | | 310 | | 0.015 111 | 0.0088.11 | 0.009211 | 0.019 11 | 003110 | 0013 0 | 0.017 () | 0.0006.111 | 0.014 1/ | 0.011(1) | 0.014 11 | 0.011 (1) | 0.031.111 | 0.008811 |
| 1.1-Dichloroethane | | 1700 | 23 | 0.01510 | 0.0088.11 | 0.0092.111 | 0.019 11 | 00010 | 0.013 (1) | 0.017 (0) | 0.00% 11 | 0.01401 | 0.01110 | 001411 | 0.011 18 | 0.031.111 | 0.0141/ |
| 1,2-Dichlorocthane | | 0.6 | 0.02 | 0.015 UJ | 0.0088 11 | 0.0092 111 | 001910 | 001110 | 0.013 1/1 | 0.017.00 | 0.0096 111 | 0.014.1/ | 0.011 (1) | 0.014 U | 0.011 (1) | 0.031 UJ | 0 0088 1/ |
| 1,1-Dichloroethene | | 410 | 0.06 | 0.015 UJ | 0.0088 U | 0.0092 UJ | 0.019 10 | 0.031 (1) | 0.013 (1) | 001710 | 0.0096 1// | 0.014 U | 0.011 (1) | 0.014 U | 0.011 UJ | 0.031111 | 0 0088 1/ |
| trans-1,2-Dichloroethene | | 230 | 0.7 | 0.015 (J) | 0.0088 U | 0.0092 (1) | 001910 | 0.031 (1) | 0.013 U | 0.017 10 | 0.0096 UJ | 0.014 1/ | 0.011 ()) | 0.014 U | 0.011 UJ | 0.031.111 | 0.0088 U |
| cis-1,2-Dichloroethene | | 1.50 | 0.4 | 0.015 UJ | 0 0088 U | 0.0092 UJ | 001910 | 0.031 10 | 0.013 (0) | 0.017 UJ | 0.0096 1/J | 0.014 U | 0.0041.J | 0014 U | 0.011 (1) | 0.031 111 | 0.0041 J |
| 1,2-Dichloropropane | | 0.71 | 0.03 | 0 015 UJ | 0.0088 () | 0.0092 UJ | 0.019 10 | 0.031 UJ | 0.013 UJ | 0.017 (1) | 0.00% 1.0 | 0.014 U | 0.011 UJ | 0.014 U | 0.011 (1) | 0.031 11 | 0 0088 11 |
| trans-1,3-Dichloropropene | | 1.8 | 0.004 | 0.015 UJ | 0.0088 U | 0 0092 UJ | 0.019 (J) | 0.031 UJ | 0.013 UJ | 0.017 UJ | 0.00% UI | 0.014 U | 0.011.03 | 00141/ | 0.011 UJ | 0 031 111 | 0.0088 U |
| cis-1,3-Dichloropropene | | 1.8 | 0.004 | 0.015 UI | 0 0088 U | 0.0092 UJ | 0.019 UJ | 0.031 UJ | 0.013 ()) | 0 017 (1) | 0.0096 13 | 0.014 U | 0.011111 | 0.014 U | 0.011 1/1 | 0 031 10 | 0.008811 |
| Ethylbenzene | | 20 | 13 | 0.015 UJ | 0.0088 U | 0.0092 UJ | 0.019 1/J | 0.031 (1) | 0.013 ()) | 0.017 (1) | 0.0096 UI | 0.014 U | 0.011.00 | 0.014 U | 0.011 UI | 0.031 UJ | 0.0088 U |
| 2-Hexanone | | | | 0.0094 J | 0.0088 U | 0.0092 (J) | 0.019 UJ | 0.031 UJ | 0.013 UI | 0.017 UJ | 0.0096 UJ | 0.014 U | 001111 | 0014 U | 0.011 (1) | 0 021 03 | 0.0088 U |
| Isopropylbenzene | | 2000 | | 0.015 UJ | 0.0088 U | 0.0092.111 | 0019 UJ | 0.031 UJ | 0.013 UJ | 0.017 UJ | 0.00%6 1// | 0.014 U | 001103 | 0.014.11 | 0.011 [J] | 0 031 111 | 0.008811 |
| Methyl Acetale | | 92000 | | 0.015 UI | 0.0088_U | 0.0092 ()J | 0.019 ()) | 0.031 UJ | 0.013 UI | 0.017 (1) | 0.0096.UJ | 0.014 U | 0.011 UI | 0.014 U | 0.011 (0) | 0 031 01 | 0.0088.0 |
| Methylcyclohexane | | 8700 | | 0.0029 1 | 0.0088 U | 0.0092.10 | 0.0047 J | 0.031 10 | 0.0082 J | 0.0076 J | 0.0096 UI | 0.015 | 0.0049 J | 0014 0 | 0.011 01 | 0.031 () | 0.00291 |
| Methylene Chionde | | <u></u> ZI | 0.02 | 0.015 () | 0.0088 U | 0.0092 (1) | 0.019 (1) | 003110 | 0.013 U | 0017 0 | 0.0096 UJ | 0.014 () | 0.011 01 | 0.014 U | 0.011 00 | 0.031 (0) | 0.00881 |
| 4-Meinyl-2-Pentanone | | 1700 | | 0.015 10 | 0.0088.0 | 0 0092 10 | 0.019 () | 0.031 (J) | 001319 | 001710 | 0.0096 UJ | 0014 0 | 0.011 (1) | 0.014 1 | 00110 | 0.031 (1) | 0.008811 |
| 111.7.2.Tetrachloroethane | | - 1/00 | 0.003 | 0.015 UI | 0.0088.0 | 0.009210 | 0.019 01 | 0.03110 | 0.013 00 | 0.01710 | 0.0096 10 | 0.014 0 | | 0014 0 | 0.011 11 | 003101 | 0.0088.0 |
| Tetrachloroethene | | 1.4 | 0.00.1 | 0.015 () | 0.0088 U | 0.009210 | 0.019 0 | 003110 | | 001710 | 0.0004 1/1 | 0.014 () | | 0.014 (1) | 0.0110 | 11 | 0.00381 |
| Tolucne | | 520 | 17 | 0.015 00 | 0.0088.11 | 0.0092 (1) | | 0.031 0 | 0.013 00 | 0.017.01 | 0.0038 1 | 0.0063.1 | 0.00361 | 00141 | 0.011.111 | 1 0 0 0 | 0.0032J |
| 1.2.4-Trichlorobenzene | | 1000 | - 12 | 0015 00 | 0.0088.11 | 0.0092 (1) | 0019 01 | 0.03107 | 0.077 | 0.017 11 | 0.0006.113 | 0.014 111 | 0.01110 | 0.014 111 | 001110 | 0.031 10 | 0.008811 |
| 1.1.1-Trichtoroethane | | 1200 | 3 | 0.015 11 | 0.0088 U | 0.0092 (1) | 0.019 (1) | 0.031 10 | 0.013 (0 | 0.017 0 | 0.0006.10 | 0.014 1/ | 0.01110 | 0.01411 | 0.0551 | 0.0551 | 0.0088 U |
| 1.1.2-Trichloroethane | | 16 | 0.02 | 0.015 11 | 0.0088.11 | 0.0097 [1] | 11 810.0 | 0.031 1/1 | 0.013 11 | 0.017.00 | 0.0096 111 | 00141 | 0.01110 | 0.014 11 | 0.011 UF | 0.031 UJ | 0 0088 U |
| Trichloroethene | | 0.1 | 0.06 | 0.015 10 | 0.0088 [] | 0.0092.111 | 001910 | 001110 | 001310 | 0.017.11 | 0.009610 | 0.014 1/ | 0.011.111 | 0.014 1/ | 0.011 00 | 0.031 UJ | 0.00881/ |
| Trichlorofluoromethane | | 2000 | | 0.015 UJ | 0.0088 U | 0.0092 10 | 0.019 UJ | 0.031 111 | 001310 | 0.017.11 | 0.0096.117 | 0.014 U | 0011 UI | 0.014 U | 0.011 (1) | 0.031 UI | 0.0088 U |
| 1,1,2-Trichloro-1,2,2-Trifluoroethane | | 5600 | | 0.015 (1) | 0.0088 U | 0 0092 (1) | 0019 UJ | 0 031 10 | 0013 (1) | 0.017 11 | 0.0096 11/ | 0.014 1/ | 001111 | 0.014 U | 0.011 00 | 0.031.03 | 0.0088 U |
| Vinyl Chloride | | 0.75 | - | 0 015 UI | 0.0088.13 | 0.0092 1/J | 001910 | 0.031 ()) | 0.013 UJ | 0.017 (1) | 0.00% 1// | 0.014 U | 001111 | 0.014.11 | 0.011 UJ | 0.031.03 | 0.0088.11 |
| m-/p-Xylene | | 420 | 210 | 0.015 UJ | 0.0088 U | 0.0092 UJ | 0.019 UJ | 0.031 UI | 0.013 1// | 001711 | 0.00% UI | 0 0043 J | 0.0029 J | 0.014 U | 0.011.01 | 0.011 (1) | 0.0029 J |
| o-Xylene | | 420 | 210 | 0.015 UJ | _ 0.0088 U | 0.0092 UJ | 0.019 UJ | 0.031 1// | 0.013 U | 0.017 ()) | 0.0096.11 | 0.0014 J | 0.011.03 | 0.014 U | 0.011.10 | 0.031.00 | 0.0014.J |
| Semi-Volatile Organic Compounds, | | | | | | | | | | | | | | | | | |
| milligrams per kilogram | | | | | | | | | | | | | | | _ | | , |
| Acenaphihene | | 29000 | 570 | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 0.48 U | 0.4 1 | 042 U | 0.38 U | 0.42 U | 0.39 U | 046 () | 0410 | 0.48 1/ | 0.36 11 |
| Acenaphthylene | | | | 0.38 U | 0.36 U | 0.4 U | 0.45 13 | 0.48 U | 041 | 0 061 J | 0.38 1/ | 042 () | 0.39 U | 0.46 U | 0411/ | 0.4811 | 0.0011 |
| Acetophenone | | | | 0.38 1/ | 0.36 U | 0.4 U | 0.45 U | 0.48 U | 0.4 U | 0 42 1 | 0.38 U | 042 U | 0.3911 | 0.46 [] | 0410 | 0 48 11 | ND |
| Anihracene | | 100000 | 12000 | 0.75 | 0.36 U | 0.4 U | 0.45 U | 048 U | 040 | 0 32 1 | 0.041 J | 0 42 () | 0.39 U | 0.46.11 | 0411 | 0.75 | 0.0411 |
| Atrazine | | 7.8 | | 0.38 U | 0.36 U | 0.4 U | 0.45 U | I 0.48 U | 040 | 0.42 () | 0.38 U | 1 0.42 () | 0.39 U | 046 0 | 1 04111 | E 0.48 U | 0,36.0 |

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Total Manufacture Control

PANA A SUBACT TO PAN - LABOR CONTRACTOR

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TABLE 4-6

ANALYTICAL RESULTS FOR SUBSURFACE SOILS FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Sile Gowanda, New York

| | | | | | | | | Samula Locati | on Identificatio | a Death and D | te Collected | | | | | | |
|------------------------------|-------------|--------------------------|-----------|-----------|-----------|-----------|-----------|---------------|------------------|---------------|--------------|-----------|-----------|-----------|-----------|---------|----------|
| | | | | SB-1 | SB-2 | SR-3 | 58.4 | Sample Facal | SB-6 | SR.7 | SR.A | 58.9 | SB-10 | MWFP-2 | MWFP-3 | | |
| | Sø | ll Criteria ^J | | 100500007 | 100500009 | 100600014 | 100500011 | 100400013 | 100900040 | 100600019 | 100600033 | 100600035 | 100600037 | 100600017 | 100900039 | | |
| | Eastern USA | Region 9 | Soil | 5-7' | 6-8' | 3.5' | 4.6' | 6-8' | 4.6' | 7.9' | 10-12 | 7.9' | 7-9' | 5-7' | 5-7' | Maximum | Minimum |
| Constituent ² | Reckeround | PRGa | Screening | 10/05/00 | 10/05/00 | innenn | 1000000 | 100000 | 1400000 | 1000000 | 1000000 | 1000000 | 10/06/00 | 10/06/00 | 100000 | Come | Cont |
| Benzaldehyde | | 62000 | Lanu | 0 38 11 | 0.36.12 | 0416 | 10/05/00 | 10/06/00 | 10/9/2000 | 100000 | 1000/00 | 10/06/00 | 0.10.11 | 0.46.11 | 0.41.11 | 0.48.11 | 0.36.11 |
| Benzo(a)Anthracene | 1 | 2.1 | 2 | 1.3.1 | 0.36 U | 0411 | 0.45 0 | 0.062.1 | 041 | 0.02.0 | 0.58 0 | 0.42.0 | 0.39 11 | 0.291 | 04111 | 1.11 | 0.062.1 |
| Benzo(a)pyrene | 1 | 0.21 | 8 | 2.3 | 0.36 U | 0.4.11 | 04511 | 0.0671 | 041 | 12 | 0.058 / | 0421 | 0 19 11 | 0.13 / | 0.41 U | 2.3 | 0.058 J |
| Benzoch)fluoranthene | | 2.1 | 5 | 2.5 J | 0.36 U | 0.4 U | 04517 | 0.48 11 | 0411 | 11 | 0.049.1 | 04211 | 0.39 U | 0.26 1 | 0.41 U | 2.51 | 0.049 J |
| Benzo(g,h,i)perylene | | | | 1.7.1 | 0.36 U | 0.4 U | 0.45 11 | 048.0 | 041 | 11 | 0.38 U | 0.42 U | 0.39 U | 0.31.1 | 041 U | 1.71 | 0.31 J |
| Benzo(k)fluoranthene | | 21 | 49 | 2.3 J | 0.36 U | 04 U | 045 U | 0 48 13 | 0.4 U | 1.2 | 0.057 1 | 0.42 U | 0.19 U | 0.26 1 | 041 U | 2.3.1 | 0.057 3 |
| 1,1-Biphenyl | | 3.50 | | 0.38 U | 0.36 U | 0.4 U | 045 U | 048 U | 0.4 U | 042 U | 0.38 U | 0.42 U | 0.39 U | 0.46 U | 0.41 U | 0.48 U | 0.36 U |
| Butyl Benzyl Pinhalate | | 100000 | 930 | 0.38 U | 0.36 U | 0.4 U | 0 45 11 | 0.48 U | 0.4 11 | 0.42 U | 0.38 17 | 042 U | 0.39 U | 0.46 U | 0.41 U | 0 48 U | 0.3611 |
| di-N-Butylphthalate | | | | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 0 48 U | 041 | 0.42 U | 0.38 U | 0.42 U | 0.39 U | 046 U | 0.41 U | 0.48 U | 0.36 U |
| Caprolaciam | | 100000 | | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 048 U | 04 U | 0.42 U | 0.38 U | 0 42 U | 0.39 U | 0.46 U | 0.41 U | 048 U | 0.36 U |
| Carbazole | | 86 | 0.6 | 0.38 () | 0.36 U | 0.4 U | 0.45 U | 0121 | 0.4 U | 0.11.1 | 0.38 U | 0.42 U | 0.39 U | 0.46 U | 0.41 U | 0.46 [] | 0.11 J |
| Indeno(1,2,3-cd)pyrene | | 2.1 | 14 | 1.4.1 | 0.36 U | 0.4 U | 0.45 U | 0.48 U | 0.4 U | 0.91 | 0.38 U | 0.42 U | 0.39 U | 0.24 1 | 04I U | 1.4.1 | 0.24 J |
| 4-Chloroaniline | | 2500 | 0.7 | 0.38 U | 0.36 U | 040 | 0.45 U | 0.48.1/ | 0.4 U | 0.42 () | 0.38 U | 0.42 U | 0.39 U | 0.46 U | 0.41 U | 0.48 () | 0.36 U |
| his(2-chloroethoxy)methane | | ** | | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 048 U | 0.4 U | 0.42 U | 0.38 U | 042 U | 0.39 U | 0.46 U | 0.41 U | 0.48 U | 0.36 U |
| his(2-chloroethyl)ether | | 0.55 | 0.0004 | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 0.48 U | 0.4 U | 0.42 U | 0.38 U | 0.42 U | 0.39 U | 0.46 U | 0.41 U | 0.48 U | 0.36 U |
| 2-Chloronaphthalene | | | | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 0 48 LI | 0.4 1 | 0.42 1 | 0.38 U | 0.42 U | 0.39 U | 0.46 U | 0.41 U | 0.48 (/ | 0.36 () |
| 2-4. hlorophenol | | 240 | | 0.38 U | 0.16 U | 0.4 () | 0.45 U | 0.48 U | 0.4 U | 0.42 U | 0.38 U | 0.42 U | 0.39 U | 0.46 U | 0.41 1/ | 0.48 () | 0.36 0 |
| 2.2-oxynist i-cnioropropane) | | | | 0.18 0 | 0.36 U | 0.4 () | 0.45 U | 0.48 U | 0.4 () | 0.42 U | 0.38 U | 0.42 U | 0.39 0 | 0.46 0 | 0.41 U | 0.48.0 | 0.36 0 |
| Cillysenc | | 210 | 100 | - 1.3 | 0.36 U | 0.4 U | 0.088) | 0.0617 | 0.058 J | 1.4 | 0.073 J | 0.42 0 | 0.39 0 | 0.291 | 0410 | 0641 | 0.0.16 1 |
| Dibensofuran | | | <u> </u> | 0.04 1 | 0.36 U | 0.4 U | 0.45 0 | 0.48 0 | 040 | 0.371 | 0.38 U | 0.42 0 | 0.39 0 | 0.11 1 | 0.41 U | 0.041 | 0.0511 |
| 3.1 Dichlorobenzidine | | | 0.007 | 0.257 | 0.36 1 | 0.4 0 | 0.0511 | 0.0721 | 0.4 0 | 0.111 | 0.38 () | 0.42 U | 0.39 () | 0.40 0 | | 0.400 | 0 38 11 |
| 2 4.1 Schlorophenol | | 1800 | | 0.38 U | 0.36 U | 041 | 0.45 U | 0.48.0 | | 0.42 0 | 0.38.0 | 0.42.0 | 0.39 () | 0461 | 04111 | 0.48 | 0.3612 |
| Diethvinhthalate | | 100000 | | 0 38 11 | 0.361 | 041 | 0.45 0 | 0.48 U | 0.4 0 | 0.42 0 | 0.38 0 | 0.42 1 | 0.39 () | 0.46 11 | 0410 | 0.4811 | 0.36 U |
| Dimethyl Phthalate | | 100000 | | 0.38 1/ | 0.36 U | 0411 | 0.45 () | 0481 | 0411 | 0.42 U | 0.38 11 | 0.42 U | 0.39 11 | 0.46 1 | 0.41 U | 048 U | 0.36 U |
| 2,4-Dimethylphenol | | 12000 | 9 | 0.38 U | 0.36 U | 0.4 U | 04511 | 048 U | 0.4 U | 04211 | 0.36 U | 0.42 11 | 0 39 11 | 046 U | 0.41 U | 048 U | 0.36 U |
| 2.4-Dinitrophenol | | 1200 | 0.3 | 0.96 U | 0.91 U | 10 | 111 | 1211 | 10 | | 0.94 1/ | | 0 97 11 | 1.2.U | 10 | 0481/ | 0.36 U |
| 2,4-Dinitrotolucne | | 1200 | 0.0008 | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 0481/ | 04 U | 0.42 U | 0.38 U | 0.42 U | 0.39 U | 046 U | 0410 | 0.48 U | 0.36 U |
| 2.6-Dinitrotoluene | | 620 | 0.0007 | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 0.48 11 | 0.4 U | 0.42 U | 0.38 U | 0.42 U | 0.39 U | 046 U | 0.41 U | 0481 | 0.36 U |
| his(2-Ethylhexyl)phthalate | | 120 | | 0.076 J | 0.36 U | 0.4 U | 0.45 () | 0.48 (/ | 0.4 U | 0.049 / | 0.38 U | 0.42 U | 0.39 U | 0.46 U | 0.41 U | 048 U | 0.0491 |
| Fluoranthene | | 22,000 | 4300 | 3.1 | 0.36 U | 0.4 U | 0.12 J | 0.075 1 | 0.073 J | 2.5 | 0.15 J | 0.42 U | 0.39 U | 0.38 1 | 0.41 U | 3.1 | 0.073 J |
| Huorene | | 26,000 | 560 | 0.45 | 0.36 U | 0.4 U | 0.45 U | 0.48 U | 0.4 U | 0.12 J | 0.38 () | 0.42 U | 0.39 U | 0.46 U | 0.41 U | 048 U | 0.12 J |
| Hexachlorobenzene | | <u> </u> | 2 | 0.38 U | 0.36 U | 0.4 1 | 0.45 U | 0.48 U | 0.4 U | 0.42 U | 0.38 Ú | 0.42 11 | 0.39 U | 046 U | 0.41 U | 048 U | 0.3617 |
| lexachlorobutadiene | | 22 | 2 | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 0.48 11 | 040 | 0.42 U | 0.38 U | 0.42 U | 0.39 U | 0 46 U | 0.4I U | 048 U | 0.36 U |
| Hexachlorocyclopentadiene | | 3700 | 400 | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 0.48 U | 2 U | 0.42 U | 0.38 U | 0.42 U | 0.39 U | 0.46 1/ | 20 | 2.0 | 0.36 U |
| Hexachloroethane | | 120 | 0.5 | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 0.48 U | 0.4 U | 0.42 1/ | 0.38 U | 0.42 U | 0.39 U | 046 U | 0.41 () | 0.48 (J | 0.36 U |
| Isophorone | | 1800 | 0.5 | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 048 U | 0.4 U | 0.42 U | 0.38 U | 0 42 11 | 0 39 11 | 0.46.11 | 0.41 U | 0.48 U | 0.36 () |
| 2-Methylnaphthalene | | | | 0.16.1 | 0.36 (1 | 0.4 U | 0.15 J | 0.2 J | 0.04.1 J | 0.151 | 0.38 U | 0.42 U | 0.39 U | 0.46 U | 0.41 U | 046.0 | 0.04.11 |
| 4.6-1 Anitro-2-Methylphenol | | | | 0.96 () | 0.91 () | 10 | | 1.2 U | 10 | 1.1.0 | 0.94 U | 1.1.0 | 0.97 U | 1.2 0 | | 1.20 | 0.91 () |
| 4-1 moro-3-Meinyiphenol | | | | 0.38 () | 0.36 U | 0.4 U | 0.45 U | 0.48 1/ | 0.4 U | 0.42 U | 0.3811 | 0.42 U | 0.39 U | 0 46 U | 0.4110 | 0.48 () | 0.36 0 |
| | | | | 0.38 U | 0.36 U | 0.4 U | 0.45 U | 0.48 U | 0.4 U | 0.42 () | 0.38 U | 0.42 U | 0.19 U | 0.46 11 | 0410 | 0.48 U | 0.36 () |
| Nashthalana | | | | 0.38 U | 0.36 U | 0.4 0 | 0.45 U | 0.48 U | 0.47 | 0.42 U | 0.38 () | 0.42 17 | 0.39 U | 0.46 U | 0.41 0 | 0.48 U | 0.0861 |
| 7 Netroapiline | | - 190 | - 84 | 0.111 | 0.36 U | 0.4 0 | 0.1 J | 0.11 J | 0371 | 0.086 | 0.38 () | 0.42 0 | 0.39 U | 0.46 U | - 0.410 | 1211 | 0.0807 |
| d-Nitmaniline | | 18 | | 0.96 0 | 0.91 U | <u> </u> | 1.1.0 | 1.2.0 | | 1.10 | 0.94 () | | 0.97 0 | 1.2 0 | <u> </u> | 1.2.17 | 0.9111 |
| Nitroben tene | | | | 0.90 0 | 0.910 | | 0.1.0 | 1.2.0 | | 1.10 | 0.94 17 | 0.1.1.1 | 0.97 () | 0.46 11 | 0411 | 0.481 | 0.3611 |
| 2-Nitronhenol | | 100 | 0.1 | 0.38 U | 0.36 U | 0.4 0 | 0.45 U | 048.0 | 0.4 0 | 0.42 0 | 0.38 0 | 0.42 0 | 0.39 0 | 0.46 0 | 0410 | 0.4811 | 0.361 |
| 4-Nitroniscool | | | | - 0.16 U | 0.50 0 | 0.4 0 | 0.45 0 | 0.48 () | | 0420 | 0.38 0 | 0.42.0 | 0.39 () | 1211 | | 1211 | 0.911 |
| u.Nitrosodinbenylamine | | - 150 | | 0.70 0 | 0.16 11 | | 1.1 0 | 0.42.11 | <u> </u> | 1.1 0 | 0.94 0 | | 0970 | 0.46.11 | 0411 | 0.48.11 | 0.3612 |
| lin-Octy/Phihalate | | 25000 | 10000 | 0.38 UI | 0.36 U | 0.4 0 | 0.45 U | 0.48 () | 0.40 | 0.42 0 | 0.38 1 | 0.42.0 | 0.19 0 | 0.46 1 | 0411 | 04811 | 0.36 U |
| Pentachlorophenol | | | 0.01 | 0.96 11 | 0.50 0 | 111 | | 1211 | 0.40 | 0.42 0 | 0.38 0 | 1111 | 0.390 | 1,7,11 | 1 11 | 1211 | 0.91 U |
| Phenanthrene | | | | 16 | 0.36.11 | | 0121 | 0141 | 0.088 1 | | 0.71 | 0421 | 0.97 0 | 1 0151 | 1 04111 | 3.6 | 0 088 J |
| Phenol | | 100.000 | 100 | 0.38 U | 0.36 U | 0411 | 0.12.5 | 0.48.11 | 0.000 1 | 04211 | 0.171 | 0.42 1 | 0.3911 | 0.46 11 | 0410 | 0481 | 0.361 |
| 4-Bromophenyl-Phenylether | | | | 0.38 1/ | 0.36 11 | 0411 | 0.45 11 | 04811 | 0411 | 04211 | 0.38.11 | 047 11 | 0 10 11 | 04611 | 04111 | 048.0 | 0.36 U |
| 4-Chlorophenyl-Phenylether | | · | | 0.38 U | 0.36 U | 0.4 11 | 0451 | 0.48.17 | 041/ | 04211 | 0 38 11 | 0421 | 0.3511 | 0.46 1/ | 0411 | 048 U | 0.36 (/ |
| n-Nitroso-di-n-Propylamine | | 0.25 | 0.00005 | 0.38 U | 0.36 U | 0.4 U | 045 12 | 0.48 1 | 0.4 1 | 0.42 1/ | 0.18 U | 0.42 11 | 0.19.11 | 0.46 U | 041 U | 0.48 () | 0.36 U |
| Рутепе | | 29,000 | 4,200 | 2.5 | 0.36 U | 041 | 011 | 0.991 | 0 072 1 | 22 | 0151 | 042 () | 0 19 11 | 0 38 / | 0411 | 2.5 | 0.072 J |
| 246 Trichlomohenol | | - 43 | 0.1 | 0.28.11 | 0.76 11 | | | | | 1 1 1 | | t | 0.00 | + | 0 41 11 | 0.48.11 | 0.36.01 |

Page 2 of 3

Page 3 of 3

ANALYTICAL RESULTS FOR SUBSURFACE SOILS FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Site Gowanda, New York

| | | | | | | | | Samala Locati | | a Death and De | to Collected | | | | | | |
|---------------------------------|----------------|--|--------|------------|-----------|-------------|--------------|---------------|-----------|----------------|--------------|--------------|--------------|--------------|------------|---------|----------|
| | | | • | 501 | 68.2 | CP 1 | F | Sample 12can | CP 4 | C. 7 | CD . | CR A | 68.10 | WWED 2 | AUTOR 1 | | |
| | | ana da | | 10050007 | 30-4 | 38-3 | 100000011 | 38-3 | 100000040 | 100400010 | 100400011 | 30-9 | 100400017 | 100400017 | 100000010 | | |
| | | | | 100/30000/ | 100500009 | 100000014 | 100300011 | 10000013 | 10040000 | 100000000 | 100000033 | 100000033 | 1000000337 | 1004444/17 | 1007900039 | | |
| | Eastern USA | Region 9 | | \$-7° | 6-8' | 3-5' | 4-6' | 6-8' | 4.6' | 7.9' | 10-12' | 7.9' | 7.9' | 5.7" | 5-7' | Maximum | Minimum |
| Constituent ² | Background | PRGs | Levels | 10/05/00 | 10/05/00 | 10/06/00 | 10/05/00 | 10/06/00 | 10/9/2000 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/06/00 | 10/09/00 | Conc. | Conc. |
| 2,4,5-Trichlorophenol | | 62000 | 270 | 0.96 U | 0.91 U | 1 U | LEU | 1.2 U | 1.0 | 1.1 U | 0.94 U | 1.1 0 | 0.97 U | 1.2 U | <u> </u> | 1.2 U | 0.91 U |
| Metals, milligrams per kilogram | | | | | | _ | | | | | | | | | | | |
| Aluminum | 33,000 | 100,000 | | 7210 | 6570 | 6700 | 3210 | 3940 | 6310 | 3280 | 6670 | 8050 | 5990 | 3900 | 5890 | 8050 | 3210 |
| Antimony | | 410 | 5 | 6.6 UJ | 6.5 UJ | 7.I UJ | 8.3 U) | 8.5 UJ | 7.1 UJ | 7.4 (J) | 6.6 UJ | 7.4 UJ | 6.7 UJ | 9.71 | 7.1 10 | 9.7 | 6.5 |
| Arsenic | 3.12** | 1.6 | 29.0 | 12.5 | 8.9 | 8.8 | 12.8 | 3.7 | 6.1 | 6.6 | 5.8 | 14.6 | 6.9 | 23.6 | 10 | 23.6 | 3.7 |
| Barium | 15-600 | 67,000 | 1,600 | 76.4 | 69 | 56.7 | 51.4 | 71.8 | 54.8 | 550 | 53.2 | 47.2 | 41.2 | 145 | 46 | 550 | 41.Z |
| Beryllium | 0-1.75 | 1,900 | 63 | 0.55 U | 0.54 U | 0.59 U | 0.69 U | 0.71 U | 0.59 UJ | 0.62 U | 0.55 U | 0.61 U | 0.56 U | 069 U | 0.59 U | 0.71 | 0.54 |
| Cadmium | 0.1-1 | 450 | 8 | 0.55 U | 0.54 U | 0.59 U | 0.69 U | 0.71 U | 0.59 UI | 1.3 | 0.55 U | 0.61 U | 0.56 U | 0.96 | 0.59 U | 1.3 | 0.54 |
| Calcium | 130-35,000** | " | | 4800 | 2020 | 1270 | 6600 | 4600 | 14200 | 10100 | 1930 | 7110 | 5640 | \$7000 R | 1550 | 67000 | 1270 |
| Chromium | 1.5-40** | 210 | 38 | 11.2 | 9.5 | 8.9 | 25.5 | 6.2 | 9 | 48.3 | 8.2 | 13.2 | 8.5 | : 155 | 10.7 | 155 | 6.2 |
| Hexavalent Chromium | | 1,900 | | 4.63 UJ | 4.4 J | 4.9 UJ | 5.5 UI | 5.79 UJ | 4.81 U | 5.07 UJ | 4.55 UJ | 5.12 UI | 4.67 UJ | 5.56 UJ | 4.91 U | 5.79 | 4.4 |
| Cobalt | 2.5-60** | 41,000 | - | 7.5 | 7.6 | 7.6 | 6.9 U | 7.1 U | 7.5_ | 6.2 U | 6.7 | 7.5 | 6.6 | 6.9 11 | 7,4 | 7.6 | 6.2 |
| Coppet | 1-50 | 64 | 38 | 17.8 | 19.9 | 11.5 | 1 0 - | 11.3 | 19 | F. 187 F | 13.5 | 25.8 | 15.2 | 94.6 | 22.3 | 187 | 11.3 |
| Iron | 2,000-550,000 | 100,000 | - | 18400 | 19400 | 18200 | 12600 | 6650 | 17600 J | 18100 | 15800 | 16800 | 15700 | 24900 | 17800 1 | 24900 | 6650 |
| l cad | 4-61*** | 750 | - | 37 J | 8.8 J | 8.4 J | 37.1 J | 7.2 J | 8.8 | 457 J | 101 | 12.9 1 | 11.9 J | 1950 J | 9.1 | 1950 | 7.2 |
| Magnesium | 100-5,000 | | | 2370 | 2760 | 2340 | 851 | 1250 | 3070 | 1790 | 1750 | 5620 | 3800 | 4710 | 2,340 | 5620 | 851 |
| Manganese | 50-5,000 | 19,000 | - | 328 | 453 | 366 | 63.4 | 59.6 | 351 | 173 | 290_ | 126 | 278 | 373 | 243 | 453 | 59.6 |
| Mercury | 0.001-0.2 | 310 | | 0.06 U | 0.05 U | 0.06 U | 0.17 | 0 07 U | 0.17 | 0.18 | 0.06 U | 0.06 U | 0.06 U | 3.1 | 0.06 U | 3.1 | 0.05 |
| Nickel | 0.5-25 | 20,000 | (30 | 16.5 | 17.3 | 16.8 | 10.5 | 13 | 16 | 13.3 | 13.7 | 21.1 | 15.5 | 13.1 | 17.2 | 21.1 | 10.5 |
| Potassium | 8,500-43,900** | - | | 764 | 767 | 675 | 337 | 354 | 516 | 818 | 452 | 767 | 762 | 411 | 534 | 818 | 337 |
| Setenium | 0.1-3.9 | 5,100 | 5 | 2 | 1.7 | 2.2 | <u></u> | 1.2 | 2 1 | 1.9 | 1.4 | 1.6 | 2 | 1.5 | 1.7.1 | 2.2 | <u> </u> |
| Silver | | 5100 | 34 | 110 | <u> </u> | 1.2 0 | 1.4 U | 1.4 U | 1.2 0 | 1.2 U | 1.1.0 | <u>1.2 U</u> | 1.1 U | <u>1.4 U</u> | 1.2 0 | 1.4 | 1.1 |
| Sodium | 6,000-8,000 | | | 302 | 345 | 393 | 608 | 538 | 757 | 476 | 409 | .563 | 422 | 762 | 460 | 762 | 302 |
| Thallium | | 67 | | 1.2 U | 1.I U | 1.2 U | 1.4 U | 1.4 U | 1.2.0 | · 1.2 U | 1.1 U | 1.2 U | L <u>L U</u> | 1.3 U | 1.2 U | 1.4 | 1.1 |
| Vanadium | 1-300 | 7,200 | 6,000 | 15.8 | 12.3 | 13.8 | 13.4 | 9.3 | 12.4 | 10.1 | 15.5 | 16.4 | 13.7 | 17 | 12.5 | 17 | 9,3 |
| Zinc | 9-50 | 100,000 | 12,000 | 81.3] | 405 J | 48.71 | 294 3 | 37.8 J | 69.6 | 154 J ~ | 47.7 1 | * 222 1 | 84.2 J | 605 1 | 64.2 | 605 | 37.8 |
| Others | | | | | | | | | | | | | | | | | |
| Percent Solids, % | | | | 86.3 | 90.9 | 81.6 | 72.7 | 69.1 | 83.1 | 78.9 | 87.9 | 78.2 | 85.6 | 72 | 81.4 | 90.9 | 69.1 |
| pH | | | | 8.32 | 8.46 | 7.14 | 8.49 | 7.05 | 10.1 | 7.74 | 7.69 | 7.91 | 7.26 | 7.95 | 7.5 | 10.1 | 7.05 |
| Total Organic Cathon, 96 | | | | 0.56 | 0.18 | 0.5 | 1.8 | 1.6 | 1.3 | 0.63 | 0.35 | 0.49 | 0.48 | 1.7 | 0.29 | 1.8 | 0.18 |

Note

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 1007 data validation performed by Data Validation Services. The analytical results for the SVOC, 3. Nitroaniline, was rejected during data validation for each sample.

3 Soil criteria from U.S. EPA. Region 9 Preliminary Remediation Goals (PROs) for Industrial Soil (October 2003) and from range of background metals concentrations measured in soil found in the

eastern United States from NYSDEC Division of Technical and Administrative Ouidance Memorandum (TAGM) \$4046.

** A New York State Background value

*** Background levels for lead vary widely, average levels in undeveloped, rural areas range from 4-61 ppm while metropolitar/suburban areas range from 200-500 ppm.

J = indicates a taboratory estimated value or estimated as a result of data validation.

U = indicates compound was not detected at or above the listed detection limit.

UI = indicates compound was not detected above the listed detection limit.

However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitiation necessary to accurately

and precisely measure the compound in the sample.

R= indicates data rejected by data validator. Mgs = feet below ground surface SB = Site Background -- = indicates value does not exist. indicates concentration above soil criteria.

MATRIX GEO

Page 1 of 2

TABLE 4-7

SUMMARY OF LANDFILL GAS ANALYTICAL RESULTS

Peter Cooper Site Gowanda, New York

| | | Sample ID and | Landfill Gas Monitorin | g Well Location |
|---------------------------|-------------------|---------------|------------------------|-----------------|
| | | 101200081 | 101200080 | 101200082 |
| | | GMW-1 | GMW-2 | GMW-3 |
| Constituent | Unit | 10/12/00 | 10/12/00 | 10/12/00 |
| Field Measured Parameters | | | | |
| Lower Explosive Limit | % | >100 | 45 | >100 |
| Hydrogen Sulfide Gas | ppm (7 | >1000 | >1000 | /10 |
| Oxygen Cathon Monoxide | %, V/V | 0.5 | 21.3 | 6 |
| PID Measurements | ppm | 2.5 | 325 | 13 |
| Laboratory Parameters | | | | |
| Oxygen + Argon | %, v/v | 6.58 | 22.1 | 12.9 |
| Nitrogen | %, v/v | 52.5 | 77.6 | 57.2 |
| Methane | %, v/v | 31.1 | 0.145 | 18.7 |
| Carbon Dioxide | %, v/v | 9.8 | 0.136 | 11.2 |
| Chloromethane | ug/m' | <25 | <2 | <25 |
| Vinyl Chloride | ug/m ³ | <25 | <2 | <25 |
| Bromomethane | ug/m ³ | <25 | <2 | <25 |
| Chioroethane | ug/m ³ | <25 | <2 | <25 |
| Acetone | ug/m ³ | 1200 | 150 | 2900 |
| Trichlorofluoromethane | ug/m ³ | <25 | 1.7 TR | <25 |
| 1,1-Dichloroethene | ug/m ³ | <25 | <2 | <25 |
| Methylene chloride | ug/m ³ | <25 | <2 | <25 |
| Trichlorotrifluoroethane | ug/m ³ | <25 | <2 | <25 |
| Carbon Disulfide | ug/m ³ | 250 | 93 | 3200 |
| trans-1,2-Dichloroethene | ug/m ³ | <25 | <2 | <25 |
| 1.1-Dichloroethane | ug/m ³ | <25 | <2 | <25 |
| Methyl tert-Butyl Ether | ug/m ³ | <25 | <2 | <25 |
| Vinyl Acetate | ug/m ³ | <25 | 14 | <25 |
| 2-Butanone | ug/m ³ | 290 | 43 | 1100 |
| cis-1,2-Dichloroethene | ug/m ³ | <25 | <2 | <25 |
| Chloroform | ug/m ³ | <25 | <2 | <25 |
| 1,2-Dichloroethane | ug/m ³ | <25 | <2 | <25 |
| 1,1,1-Trichloroethane | ug/m ³ | <25 | <2 | <25 |
| Benzene | ug/m ³ | 180 | <2 | 74 |
| Carbon Tetrachloride | ug/m ³ | <25 | <2 | <25 |
| 1,2-Dichloropropane | ug/m ³ | <25 | <2 | <25 |
| Bromodichloromethane | ug/m ³ | <25 | <2 | <25 |
| Trichloroethene | ug/m ³ | <25 | <2 | <25 |
| cis-1,3-Dichloropropene | ug/m ³ | <25 | <2 | <25 |
| 4-Methyl-2-pentanone | ug/m ³ | 370 | 3.4 | 140 |
| trans-1.3-Dichloropropene | ug/m ³ | <25 | <2 | <25 |
| 1,1,2-Trichloroethane | ug/m ³ | <25 | <2 | <25 |
| Toluene | ug/m ³ | 2600 | 41 | 270 |
| 2-Hexanone | ug/m ³ | <25 | 7 | <25 |
| Dibromochloromethane | ug/m ³ | <25 | <2 | <25 |

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SUMMARY OF LANDFILL GAS ANALYTICAL RESULTS

Peter Cooper Site Gowanda, New York

| | | Sample ID and | Landfill Gas Monitori | ng Well Location |
|---------------------------|-------------------|--------------------|-----------------------|--------------------|
| Constituent | | 101200081 GMW-1 | 101200080 GMW-2 | 101200082 GMW-3 |
| Constituent | Unu | 10/12/00 | 10/12/00 | 10/12/00 |
| 1,2-Dibromoethane | ug/m ³ | <25 | <2 | <25 |
| Tetrachloroethene | ug/m ³ | <25 | <2 | <25 |
| Chlorobenzene | ug/m ³ | <25 | <2 | <25 |
| Ethylbenzene | ug/m ³ | 66 | 3.5 | 84 |
| m- & p-Xylenes | ug/m ³ | 9 9 | 3.3 | 130 |
| Bromoform | ug/m ³ | <25 | <2 | <25 |
| Styrene | ug/m ³ | <25 | <2 | 20 TR |
| o-Xylene | ug/m ³ | 51 | 1.4 TR | 60 |
| 1,1.2.2-Tetrachloroethane | ug/m ³ | <25 | <2 | <25 |
| 1,3-Dichlorobenzene | ug/m ³ | <25 | <2 | <25 |
| 1,4-Dichlorobenzene | ug/m ³ | 48 | <2 | <25 |
| 1.2-Dichlorobenzene | ug/m ³ | <25 | <2 | <25 |

Notes:

1. Qualifications reflect the 100% data validation performed by Data Validation Services.

2. Sample locations shown on Plate 1.

< = none detected

TR = trace value

| 4. Provide the first state of a second state bit was a terred when the state only environment of the second state. | A CONTRACTOR OF A DATA STREET, AND A |
|--|--|
|--|--|

20

2*

Compound²

Potassium

Sodium

Zinc

Volatile Organic Compounds,

| 286 | 213 | 160 | 209 | 127 | 164 | 116 | 209 | 323 | 473 | 203 | 213 | 106 | 235 | 179 | 167 | 473 |
|--------|------------|--------|------------|--------|------------|--------|------------|--------|------------|--------|------------|---------|----------|--------|------------|----------|
| 0.01 U | 0.01 U | 0.143 | 0.251 | 0.436 | 0.366 | 0.209 | 0.371 | 0.01 U | 0.01 U | 0.0293 | 0.0228 | 0.0137 | 0.01 UJ | 0.01 U | 0.01 U | 0.436 |
| 0.01 U | (0.01 U) R | 0.01 U | (0.02 U) R | 0.04 U | (0.01 U) R | 0.0215 | (0.04 U) R | 0.01 U | (0.01 U) R | 0.01 U | (0.01 U) R | 0.01 UJ | 0.0172 J | 0.01 U | (0.01 U) R | 0.04 U |
| 0.I U | 0.1 U | 0.107 | 0.1 U | 0.1 U | 0.13 | 0.I U | 0.14 | 23 | 41 5 5 | 13.4 | 16.6 | 9.04 | 2.29 | 10.5 | 11.78.52 | 41 |
| 25 | 16.8 | 90.2 | 154 | 167 | 136 | 83.6 | 150 | 41.6 | 37 | 73.9 | 61.8 | 22.9 | 34 | 25.7 | 20.7 | 167 |
| 6.4 | 4.28 | 4.07 | 5.74 | 5.83 | 5.93 | 8.88 | 9.49 | 9.86 | 7.87 | 5.85 | 4.67 | 37.6 | 22.2 | 5.1 | 4.28 | 37.6 |
| 11.6 | 9.08 | 17.6 | 22.1 | 20.9 | 18.5 | 22.1 | 26.1 | 25.8 | 12.4 | 8.31 | 5 U | 1670 | 229 | 28.2 | 28.6 | 1670 |
| 0223 | 0.0297 | 0.0208 | 0.03 U | 0.02 U | 0.0234 J | 0.02 U | 0.02 U | 0.178 | 0.02 U | 0.02 U | 0.03 U | 0.151 | 0.03 U | 0.0656 | 0.204 | 0.204 |
| | | | | | | | | | | | | | | | | |
| NA | NA | NA | NA | NA | 0.0538 J | NA | NA | NA | NA | NA | NA | 0.0145 | NA | NA | NA | 0.0538 J |
| NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | 114 | NA | NA | NA | 114 |
| NA | NA | NA | NA | NA | 0.354 | NA | NA | NA | NA | NA | NA | U 10.0 | NA | NA | NA | 0.354 |
| NA | NA | NA | NA | NA | (0.01 U) R | NA | NA | NA | NA | NA | NA | 0.013 J | NA | NA | NA | 0.013 J |
| NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | ŇA | NA | 4.61 | NA | NA | NA | 4.61 |
| NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | 23.7 | NA | NA | NA | 23.7 |
| NA | NA | NA | NA | NA | 'NA | NA | NA | NA | NA | NA | NA | 38.5 | NA | NA | NA | .38.5 |
| NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | 1630 | NA | NA | NA | 1630 |
| NA | NA | NA | NA | NA | 0.105 J | NA | NA | NA | NA | NA | NA | 0.079 | NA | NA | NA | 0.105 J |
| | | | | | | | | | | | | | | | | |

TABLE 4-8 ANALYTICAL RESULTS FOR OVERBURDEN GROUNDWATER SAMPLES FROM THE INACTIVE LANDFILL AREA

Peter Cooper Site Gowanda, New York

MW-45

111000117 050301144

5/3/2001

11/10/2000

Sample Location, Identification and Date Collected

MW-5S

110900112 050301143

5/3/2001

11/9/2000

计推动分词补偿

| micrograms per liter | 1 | 1 | | 1 | 1 | 1 | 1 | 1 | 1 | | 1. | 1 | | | | | | | 1 |
|--|----------|--------|------------|--------|------------|--------|------------|--------|------------|--------|------------|--------|------------|---------|----------|--------|------------|----------|---------|
| Benzene | 1 | 10 U | 10 U | 100 U | 10 U | 100 U | 10 UJ | 100 U | 1.3 J | 10 U | 10 U | 1161 | (1.5.1) | 10 U | 10 U | 10 U | 10 U | 100 U | 1.3.1 |
| Chlorobenzene | 5 | 10 U | 10 U | 100 U | 10 U | 100 U | 10 UJ | 100 U | 47 | 10 U | 10 U | 160 | 190 | 10 U | 10 U | 10 U | 10 U | 190 | 10 UJ |
| 1.2-Dichlorobenzene | 3 | 10 U | 10 U | 100 U | 10 U | 100 U | 10 UJ | U 001 | 51 | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 100 U | 51 |
| 1,4-Dichlorobenzene | 3 | 10 U | 10 U | 100 U | 10 U | 100 U | 10 11 | 100 U | 2.4 J | 10 U | 10 U | 100 | 10 U | 10 U | 10 U | 10 U | 10 U | 100 U | 2.4 J |
| Ethylbenzene | 5 | 10 U | 10 U | 100 U | 10 U | 100 U | 1.6.1 | 100 U | 10 0 | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 100 U | 1.6 |
| m/p-Xylene | 5 | 10 U | 10 U | 100 U | 10 U | 100 U | 10 UI | 100 U | 1.1 | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 100 U | 11 |
| o-Xylene | 5 | 10 U | 10 U | 100 U | 10 U | 100 U | 10 UJ | 100 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 100 U | 10 U |
| Toluene | 5 | 10 U | 10 U | 100 U | · 10 | 2 171 | 10 1 | 100 U | 3.2 J | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 100 U | 3.2 J |
| Semi-Volatile Organic Compounds, | 1 | 1 | 1 | 1 | | | 1 | | 1 | | 1 | | | | | | | | 1 |
| micrograms per liter | | | · · | | | • | | 1 | 1 | 1 | 1 | · | | | | | | | |
| 2-Chlorophenol | | R | 9.4 U | 10 U | 9.4 U | 20 U | 9.7 U | 100 | 9.4 U | 10 U | 9.4 U | 1.4 J | 1.8.1 | 10 U | 9.4 U | 10 U | 9.4 U | 20 U | 1.4 J |
| 2,4-Dichlorophenol | 5* | R | 9.4 U | 10 U | 69.4 U | 20 U | 9.7 U | 10 U | 9.4 U | 10 U | 9.4 Ŭ | 100 | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 69.4 U | 9.4 U |
| 2,4-Dimethylphenol | 50* | R | 9.4 U | 10 U | 9.4 U | 2.6 J | 31 | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 100 | 9.4 U | 10 U | 2.6 J |
| 2,4-Dinitrophenol | 10* | R | 24 U | 50 U | 24 U | 100 UJ | 24 U | 50 UJ | 24 U | 50 UJ | 24 U | 50 UJ | 24 U | 50 UJ | 24 U | 50 U | 24 U | 100 UI | 24 U |
| 4,6-Dintiro-2-methylphenol | | R | 24 U | 50 U | 24 U | 100 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 100 U | 24 U |
| 4-Chloro-3-Methylphenol | | R | 9.4 U | 10 U | 9.4 U | 20 U | 9.7 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 20 U | 9.4 U |
| 2-Methylphenol | | R | 9.4 U | 1.3 J | 8.2 J | 18 J | 8.1 J | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 18 J | 1.3.1 |
| 4-Methylphenol | | R | 9.4 U | 96 | 1400 D | 210 | 2400 D | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 Ú | 9.4 U | 10 U | 9.4 U | 2400 D | 9.4 U |
| 2-Nitrophenol | | R | 9.4 U | 10 U | 9.4 U | 20 U | 9.7 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 20 U | 9.4 U |
| 4-Nitrophenol | | R | 24 U | 50 U | 24 U | 100 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 100 U | 24 U |
| Pentachiorophenol | | R | 24 U | 50 U | 24 U | 100 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 100 U | 24 U |
| Phenol | <u> </u> | R R | 9.4 U | 15 | 220 DJ | 38 | 480 DJ | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 480 DJ | 9.4 U |
| 2,4,6-Trichlorophenol | | R | 9.4 U | 10 U | 9.4 U | 20 U | 9.7 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 20 U | 9.4 U |
| 2,4,5-Trichlorophenol | | R | 24 U | 10 U | 24 U | 20 U | 24 U | 10 U | 24 U | 10 U | 24 U | 24 U | 10 U |
| Total Metals, milligrams per liter | | ļ | j | 1 | 1 |] | | | | | | | | | | | | | |
| Arsenic | 0.025 | 0.01 U | 0.01 U | 0.151 | 0.196 | 0.0621 | 0.0479 J | 0.0714 | 0.0582 | 0.01 U | 0.01 U | 0.0338 | 0.025 U | 0.0172 | 0.025 U | 0.01 U | 0.01 U | 0.196 | 0.01 U |
| Calcium | | 286 | 213 | 160 | 209 | 127 | 164 | 116 | 209 | 323 | 473 | 203 | 213 | 106 | 235 | 179 | 167 | 473 | 106 |
| Chromium | 0.05 | 0.01 U | 0.01 U | 0.143 | 0.251 | 0.436 | 0.366 | 0.209 | 0.371 | 0.01 U | 0.01 U | 0.0293 | 0.0228 | 0.0137 | 0.01 UJ | 0.01 U | 0.01 U | 0.4.36 | 0.01 U |
| Hexavalent Chromium | 0.05 | 0.01 U | (0.01 U) R | 0.01 U | (0.02 U) R | 0.04 U | (0.01 U) R | 0.0215 | (0.04 U) R | 0.01 U | (0.01 U) R | 0.01 U | (0.01 U) R | 0.01 UJ | 0.0172 J | 0.01 U | (0.01 U) R | 0.04 U | 0.01 U |
| Iron | 0.3 | 0.I U | 0.1 U | 0.107 | 0.1 U | 0.1 U | 0.13 | 0.I U | 0.14 | 23 | 41 5 5 | 13.4 | 16.6 | 9.04 | 2.29 | 110.5 | 11.78.52 | 41 | 0.10 |
| Magnesium | 35* | 25 | 16.8 | 90.2 | 154 | 167 | 136 | 83.6 | 150 | 41.6 | 37 | 73.9 | 61.8 | 22.9 | 34 | 25.7 | 20.7 | 167 | 16.8 |
| Potassium | | 6.4 | 4.28 | 4.07 | 5.74 | 5.83 | 5.93 | 8.88 | 9.49 | 9.86 | 7.87 | 5.85 | 4.67 | 37.6 | 22.2 | 5.1 | 4.28 | 37.6 | 4.07 |
| Sodium | 20 | 11.6 | 9.08 | 17.6 | 22.1 | 20.9 | 18.5 | 22.1 | 26.1 | 25.8 | 12.4 | 8.31 | 5 U | 1670 | 229 | 28.2 | 28.8 | 1670 | 50 |
| Zinc | 2• | 0.0223 | 0.0297 | 0.0208 | 0.03 U | 0.02 U | 0.0234 J | 0.02 U | 0.02 U | 0.178 | 0.02 U | 0.02 U | 0.03 U | 0.151 | 0.03 U | 0.0656 | 0.204 | 0.204 | 0.02 U |
| Soluble Metals ⁴ , milligrams per liter | | | | | | | 1 | | | | | | | | | [| | | |
| Arsenic | 0.025 | NA | NA | NA | NA | NA | 0.0538 J | NA | NA | NA | NA | NA | NA | 0.0145 | NA | NA | NA | 0.0538 J | 0.0145 |
| Calcium | · · · | NA | NA NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | 114 | NA | NA | NA | 114 | 114 |
| Chromium | 0.05 | NA | NA | NA | NA | NA | 0.354 | NA | NA | NA | NA | NA | NA | 0.01 UJ | NA | NA | NA | 0.354 | 0.01 UJ |
| Hexavalent Chromium | 0.05 | NA | NA | NA | NA | NA | (0.01 U) R | NA | NA | NA | 1 NA | NA | NA | 0.013 J | NA | NA | NA | 0.013 J | 0.013 J |
| tron | 0.3 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA NA | NA | NA | 4.61 | NA | NA | NA | 4.61 | 4.61 |
| Magnesium | 35* | NA | NA | 23.7 | NA | NA | NA | 23.7 | 23.7 |
| Potassium | | NA NA | NA | NA | NA NA | NA NA | NA NA | NA | NA | NA | N.A | NA | NA | 385 | NA | NA | NA | 38.5 | 38.5 |

Page 1 of 2

Minimum

Conc.

1630

0.079

301132

Maximum

Conc.

MW-75

050401151

5/4/2001

111000116

11/10/2000

MW-6S

110700110 050401152

5/4/2001

11/7/2000

MW-8S

110800091 043001121

4/30/2001

11/8/2000

A MARY CONTRACT OF

MW-2SR

050401147

5/4/2001

110700108

11/7/2000

MW-3SR

110700109 050201136

5/2/2001

11/7/2000

MW-ISR

111000120 050101123

5/1/2001

11/10/2000

Groundwater

Criteria '





TABLE 4-8

Page 2 of 2

ANALYTICAL RESULTS FOR OVERBURDEN GROUNDWATER SAMPLES FROM THE INACTIVE LANDFILL AREA

Peter Cooper Site Gowanda, New York

| | 1 | Sample Location, Identification and Date Collected | | | | | | | | | | | | | | | | | |
|--|-------------|--|-----------|-----------|-----------|-----------|-----------|------------|-----------|---------------|-----------|-----------|-----------|------------|-----------|-----------|------------|---------------|----------|
| | | MW. | ISR | MW | -2SR | MW | -35R | MW | -45 | MW- | 55 | МИ | -65 | МИ | V-75 | MW | -85 | | |
| f · | Groundwater | 111000120 | 050101123 | 110700108 | 050401147 | 110700109 | 050201136 | 111000117 | 050301144 | 110900112 | 050301143 | 110700110 | 050401152 | 111000116 | 050401151 | 110800091 | 043001121 | Maximum | Minimum |
| Compound ² | Criteria ' | 11/10/2000 | 5/1/2001 | 11/7/2000 | 5/4/2001 | 11/7/2000 | 5/2/2001 | 11/10/2000 | 5/3/2001 | 11/9/2000 | 5/3/2001 | 11/7/2000 | 5/4/2001 | 11/10/2000 | 5/4/2001 | 11/8/2000 | 4/30/2001 | Conc. | Conc. |
| Other Geochemical Data, milligrams | | | | | | | | | | | | | | | | | | | |
| per liter | | | | | | | | | | | | | | - | | | | | L |
| Ammonia | 2 | 3.26 | 1.05 | 523 | 633 | 837 | 693 | NA | 810 | 23.9 | 6.32 | 219 | 153 | 181 | 593.7 | 2.49 | 42.19 | 837 | 1.05 |
| Bicarbonate Alkalinity | | 469 | 433 | 2250 | 3200 | 3720 | 3350 | 2570 | 3850 | 622 | 410 | 1610 | 1280 | 1480 | 1000 | 371 | 321 | 3850 | 321 |
| Carbonate Alkalinity | | 2 U | 2 U | 2 U | 2 Մ | 2 U | 2 U | 20 | 2 U | 20 | 20 | 20 | 2 U | 20 | 20 | 20 | 2 U | 20 | 20 |
| Chloride | 250 | 8.13 | 9.74 | 21.7 | 17.2 | 32.8 | 22.7 | NA | 26.4 | 6.82 | 6.9 | 10.6 | 3.82 | 2310 | 587 | 22.3 | 61.5 | 2310 | 3.82 |
| Ferrous Iron | | NA | 0.1 U | 0.128 J | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | 0.128 J | 0.1 U |
| Nitrate Nitrogen | 10 | 1.16 | 1.72 | 0.5 U | 0.04 U | 0.5 U | 0.05 U | NA | 0.05 U | 0.5 UI | 0.05 U | 0.5 UJ | 0.0502 | 22.7 | 0.05 U | 0.5 U | 0.5 U | 22.7 | 0.04 U |
| Soluble Organic Carbons | | 4.2225 U | NA | 56.0 | NA | 113 | 102.5 | 57.4 | NA | 5.98 | NA | 16.5 | NA | 78.3 | NA | 5.72 U | NA | 112.75 | 4.2225 U |
| Sulfate | 250 | 416 | 168 | 463 | 48.2 | 24.0 | 99.3 | NA | 209 | 575 | 966 | 2.64 | 4.22 | 127 | 236 | 260 | 181 | 966 | 2.64 |
| Total Alkalinity | | 469 | 433 | 2250 | 3200 | 3720 | 3350 | 2570 | 3850 | 622 | 410 | 1610 | 1280 | 1480 | 1000 | 371 | 321 | 3850 | 321 |
| Total Dissolved Solids | | 1070 | · NA | 729 | NA | 995 | NA | NA | NA | 1290 | NA | 839 | NA | 4900 | NA | 770 | NA | 4900 | 729 |
| Total Hardness | | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | 570 | NA | 570 | 570 |
| Total Kjeldahl Nitrogen | | 3.2 | 1.51 | 494 | 625 | 762 | 698 | NA | 839 | 22.8 | 6.41 | 227 | 148 | 165 | 90.3 | 2.37 | 2.65 | 839 | 1.51 |
| Total Organic Carbon | | 2.8775 U | 2.62 | 56.45 | 187.50 | 112.75 | 105.00 | 55.35 | 90.03 | 5.85 U | 7.59 | 15.45 | 12.08 | 70.13 | 21.25 | 3.36 U | 2.43 | 187.5 | 2.43 |
| Total Sulfide | 0.05* | 10 | 2 U | 38.0 | : 55 | 52.0 | 37.0 J | 34.0 | 19 J | 10 | 2 UJ | 10 | 1 U1 | 10 | 1.03 | 1.1.0 | <u>2 U</u> | 55 | 101 |
| Total Suspended Solids | | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA NA | NA | NA | 13.1 | NA | 13.1 | 13.1 |
| Turbidity, NTU | | NA | 0.37 | NA | NA | NA | 150 | NA | NA | NA | NA | NA | NA | NA | NA | NA | 43.8 | 150 | 0.37 |
| Field Measured Parameters ⁵ | | | | | | | | | | | | | | | | | | | |
| Conductivity (uS/cm) | | 998 | 1063 | 3684 | 5780 | 5119 | 5704 | 3993 | 7117 | 1376 | 2185 | 2244 | 2230 | 6358 | 2986 | 1102 | 1085 | 7117 | 998 |
| Dissolved Oxygen (ppm) | | 4.34 | 1.04 | 1.43 | 3.33 | 1.63 | 0.28 | 3.79 | 5.37 | 2.28 | 5.26 | 9.34 | 0.2 | | 1.22 | 0.63 | 0.5 | 9.34 | 0.2 |
| Ferrous Iron (mg/l) | | 0 | 0 | NA | 0 | 0 | 0 | 0 | 0 | 17.5 (5x dil) | 6 | 12 | 5 | | 2 | | 6 | 17.5 (5X dil) | 0 |
| Oxidation Reduction Potential (mV) | | 11.9 | 291.1 | -371.6 | -291.4 | -369.1 | -327.8 | -365.5 | -356.3 | -95.2 | -27.9 | -196.4 | -60.3 | -96.9 | 53.5 | -74.7 | -37.7 | 291.1 | -371.6 |
| pH (pH units) | | 6.41 | 6.63 | 7.03 | 7.9 | 7.08 | 5.21 | 7.16 | 6.01 | 6.68 | 4.39 | 6.69 | 5.55 | 7.18 | 5.79 | 6.71 | 6.08 | 7.9 | 4.39 |
| Temperature (°C) | | 12.91 | 11.6 | 14.19 | 11.32 | 15.46 | 15.96 | £1.77 | 13.7 | 15.53 | 11.35 | 14.93 | 10.31 | 12.81 | 12.26 | 16.7 | 14.93 | 16.7 | 10.31 |
| Turbidity, NTU | | 3 | 127.9 | 4.72 | 46.1 | 3.24 | 25.5 | 26 | 1.5 | 14 | 3.4 | 1.07 | | >1000 | | 0.84 | 70.6 | 127.9 | 0.84 |

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

3. Groundwater critiera for Class GA groundwater as provided in Division of Water Technical and Operational Series (1.1.1), Ambient Water Quality Standards and Guidance Values and Groundwater Effluent Limitations, October 22, 1993, reissued June 1998.

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values are guidance values.

4. Samples collected for soluble metals analysis were field filtered.

5. The YSI 600XL was used in the November and May sampling events for temperature, pH, specific electrical conductance, dissolved oxygen.and redox potential measurements. Ferrous iron was field measured with the HACHHR.R field kit (for QC, 10% were sent to analytical laboratory). The turbidity treasurements on the YSI 600 XL were not accurate during the May sampling event and as such, the LaMotte turbidity meter was used to measure turbidity. Turbidity measurements were collected with the TURB2020 meter during the November sampling event. November sampling events.

mg/l = milligrams per liter

NA = not analyzed (*alocs) = laboratory reported value prior to data validation. NTI = Nephelometric Turbidity Unit uSVcm = microxicmens per centimeter at 25°C. mV = millivotts ppm = parts per million J = indicates an estimated value. U = indicates compound was not detected. D = indicates spike dibuted out. R= indicates spike dibuted out. indicates succedance of groundwater criteria. UJ = indicates compound was not detected above the listed detection limit. However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitation necessary to accurately and precisely measure the compound in the sample.

2.1

TABLE 4-9

ANALYTICAL RESULTS FOR BEDROCK GROUNDWATER SAMPLES FROM THE INACTIVE LANDFILL AREA

Peter Cooper Sile Gowanda, New York

| | | Sounds Lecition and Date Collected | | | | | | | | | | | | | | | |
|--|-------------|------------------------------------|------------|---------------|-----------|-----------|------------|-----------------|------------------|-----------|------------|-----------|-----------|-----------|------------|----------|---------|
| | | | | r | | | Sa | mple Location a | d Date Collected | | | | | | | - 1 | |
| | | MV | V-1D | MV | V-2D | MW- | 4D2 | MW. | 41)(R) | M | N-5D | MW- | 7D | М₩ | -8D | | |
| _ , | Groundwater | 111000119 | 050101124 | 110800107 | 050401148 | 110900115 | 050301146 | 111000118 | 050301145 | 110900111 | 050301141 | 110700105 | 050401149 | 110900114 | 040301122 | Maximum | Minimum |
| Constituent | Criteria | 11/10/2000 | 5/1/2001 | 11/8/2000 | 5/4/2001 | 11/9/2000 | 5/3/2001 | 11/10/2000 | 5/3/2001 | 11/9/2000 | 5/3/2001 | 11/7/2000 | 5/4/2001 | 11/9/2000 | 4/30/2001 | Conc. | Conc. |
| Volatile Organic Compounds, | | | | | | | | | | | | | | | | | • |
| Pensana | | 10 11 | 10.11 | 10.12 | | | 10.11 | | 10.11 | 1011 | | 10.11 | | | | 10.11 | - 10.11 |
| Belizene | | 10 0 | 100 | 100 | 10 0 | 10 0 | 100 | 100 | 100 | 100 | 100 | 100 | 10 0 | 100 | | 001 | (0) |
| | 3 | 10 0 | 100 | | 100 | 10 0 | 100 | 10 | 0.8 J | 100 | 10 0 | 100 | 100 | 10 0 | 100 | 10 | 10.0 |
| 1,2-Dichlorobenzene | | 10 0 | 100 | 100 | 10 0 | 100 | 100 | 10 0 | 10.0 | 100 | 100 | 100 | 10 0 | 100 | 100 | 100 | 100 |
| 1.4-Dichlorobenzene | | 10 0 | 100 | 100 | - 10 0 | 100 | 100 | 100 | 100 | 100 | 100 | 10.0 | 10 0 | 10 0 | 10 0 | 100 | 100 |
| Einyinenzene | | 10 0 | 10 0 | 100 | 100 | 100 | 100 | 10 0 | 10 0 | 10.0 | 10 0 | 100 | 10 0 | 10 0 | | 100 | 1011 |
| nvp-Aylene | 5 | | 100 | | 100 | 10 0 | 100 | 10 0 | 10 0 | 10 0 | 100 | 10 0 | 100 | 100 | | | 100 |
| Tuluana | | 10 0 | 101 | | 100 | 100 | 100 | 100 | 100 | 100 | | 10 0 | 1011 | 100 | 100 | 100 | 100 |
| Toluene | | 100 | 100 | 100 | 100 | 1.3.7 | 100. | 10.0 | 10 0 | 10.0 | 100 | 100 | 100 | | 10.0 | 100 | 1.3.1 |
| Semi-Volatile Organic Compounds. | | | | | | | [| | | 1 | | | | | | | |
| micrograms per liter | | | | | | | | | | | | | | | | | |
| 2-Chorophenol | | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 10 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U |
| 2,4,5-Trichlorophenol | | 10 U | 24 U | 10 U | 24 U | 10 U | 26 U | 10 U | 24 U | 10 U | 24 U | 10 U | 24 U | 10 U | 24 U | 26 U | 10 U |
| 2,4,6-Trichlorophenol | | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 10 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U |
| 2,4-Dichlorophenol | 5* | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 10 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U |
| 2,4-Dimethylphenol | 50* | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 10 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U |
| 2.4-Dinitrophenol | 10* | 50 UJ | 24 U | 50 U | 24 U | 50 UJ | 26 U | 50 UJ | 24 U | 50 UJ | 24 U | 50 U | 24 U | 50 UJ | 24 U | 50 UJ | 24 U |
| 2-Methylphenol | | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 10 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U |
| 2-Nitrophenol | | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 10 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U |
| 4.6-Dintiro-2-Methylphenol | | 50 U | 24 U | 50 U | 24 U | 50 U | 26 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U |
| 4-Chloro-3-Methylphenol | | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 10 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U |
| 4-Methylphenol | | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 10 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9.4 U |
| 4-Nitrophenoi | | 50 U | 24 U | 50 U | 24 U | 50 U | 26 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U |
| Pentachlorophenol | | 50 U | 24 U | 50 U | 24 U | 50 U | 26 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U | 50 U | 24 U |
| Phenol | I | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 10 U | 10 U | 9.4 U | 10 U | 9.4 U | 10 U | 9,4 U | 10 U | 9.4 U | 10 U | 9.4 U |
| Total Metals, milligrams per liter | | | | | | | | | | 1 | | | | | | | |
| Arsenic | 0.025 | 0.01 11 | 0.01.0 | 0.0248 | 0.0283.11 | 0.01.11 | 0.0483.1 | 0.0192 | 0.01.01 | 0.01.11 | 0.01.11 | 0.01.11 | 0.025 1/ | 0.01 U | 0.01 U | 0.0483 J | 0.01 U |
| Calcium | | 18.8 | 28.3 | 212 | 252 | 49.9 | 59.8 | 206 | 211 | 562 | 586 | 21.6 | 54 | 27.5 | 45.2 | 586 | 18.8 |
| Chromium | 0.05 | 0.01 U | 0.0113 | 0.0524 | 0.0551 | 0.0134 | 0.0492 | 0.133 | 0.088 | 0.01.0 | 0.01 U | 0.01 U | 0.01 UJ | 0.01 U | 0.0155 | 0.133 | 0.01 U |
| Hexavalent Chromium | 0.05 | 0.01 U | (0.01 U) R | 40 U | 0.0592 1 | 0.01 U | (0.01 U) R | (0.01 U) R | (0.01 U) R | 0.01 U | (0.01 U) R | 0.01 U | 0.0225 J | 0.01 U | (0.01 U) R | 40 U | 0.01 U |
| Iron | 0.3 | 13.5 | 16.1 J | 0.146 | 0.115 | 8.45 | 70 | 1.98 | 4.81 | 66.9 | 71.4 | 0.378 | 1.81 | 6.92 | 8.4 | 71.4 | 0.115 |
| Magnesium | 35* | 6.81 | 8.3 | 104 | 107 | 15.9 | 22.5 | 89.4 | 75.2 | 36 | 35.4 | 5.84 | 15.7 | 9.05 | 2.6 | 107 | 2.6 |
| Potassium | | 2.59 | 2.66 | 24.3 | 25.2 | 7.69 | 13.9 | 23.7 | 20.8 | 3.43 | 3.76 | 3.33 | 4.69 | 4.24 | 5.28 | 25.2 | 2.59 |
| Sodium | 20 | 154 | 144 | 295 | 297 | 950 | 1030 | 197 | 185 | 21.2 | 27 | 384 | 347 | 163 | 109 | 1030 | 21.2 |
| Zinc | 2• | 0.042 | 0.0652 | 0.14 | 0.03 U | 0.118 | 0.416 | 0.02 U | 0.0451 | 0.0348 | 0.02 U | 0.02 U | 0.03 U | 0.0655 | 0.561 | 0.561 | 0.02 U |
| Soluble Matels ⁴ millioneme see liter | | | | | | | | | | | | | | | | | |
| Arrente overalis , minigrants per mer | | 0.01 | | | | | | | | | | | | | 0.01.11 | 0.036.11 | 0.01.11 |
| Calcium | 0.025 | 0.01 0 | 0.01 U | NA NA | | NA | 0.025 0 | 0.0152 | 0.025 U | NA. | NA NA | NA | NA NA | | 0.01.0 | 200 | 14.0 |
| Chromium | | 0.01.11 | 24.5 | NA | | NA | NA | 209 | NA 0.0P21 | <u>NA</u> | NA | NA | | NA NA | | 0.134 | 0.01.11 |
| Landard Chromium | 0.05 | | | NA | | NA | 0.0114 | 0.134 | 0.0821 | NA | NA | NA | NA | NA | | 0.1.14 | 0.010 |
| tron | 0.03 | 0.01 0 | | | NA | NA | 0.0103 1 | 0.01 0 | | <u>NA</u> | NA | NA | NA NA | NA NA | 0.0(18.) | 0.0118.1 | 0.010 |
| Magnesium | | 4 76 | 6.99 | NA NA | | NA NA | | 0.920 | | | | | | | NA | 90 R | 4 76 |
| Potassium | .,, | 711 | 2.00 | NA NA | | NA | | 24.4 | NA NA | NA | | | jNA NA | | | 24.4 | 211 |
| Sodium | 20 | 154 | 140 | NA I | NA | NA | | 29.4 | NA NA | NA | | | | NA NA | NA | 203 | 140 |
| Zinc | 2. | 0.02 [] | 0.0236 | NA | NA | NA | 0.0784 | 0.02.11 | 0.03.11 | NA | NA | NA | NA | NA | 0.02 U | 0.0784 | 0.02 U |

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TABLE 4-9

Page 2of 2

ANALYTICAL RESULTS FOR BEDROCK GROUNDWATER SAMPLES FROM THE INACTIVE LANDFILL AREA

Peter Cooper Site Gowanda, New York

| <u></u> | | Sample Location and Date Collected | | | | | | | | | | | | | | | |
|--|-------------|------------------------------------|-----------|-----------|-----------|-----------|-----------|------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|---------|---------|
| | | МИ | Y-1D | MV | V-2D | MW- | 4D2 | MW- | 4D(R) | M | ¥-5D | MW | 7D | М₩ | -8D | | |
| | Groundwater | 111000119 | 050101124 | 110800107 | 050401148 | 110900115 | 050301146 | 111000118 | 050301145 | 110900111 | 050301141 | 110700105 | 050401149 | 110900114 | 040301122 | Maximum | Minimum |
| Constituent ² | Criteria ' | 11/10/2000 | 5/1/2001 | 11/8/2000 | 5/4/2001 | 11/9/2000 | 5/3/2001 | 11/10/2000 | 5/3/2001 | 11/9/2000 | 5/3/2001 | 11/7/2000 | 5/4/2001 | 11/9/2000 | 4/30/2001 | Conc. | Conc. |
| Other Geochemical Data, milligrams | | | | | | | | | | | | | | | | | |
| per liter | | | | | | | | | | | | | | | | | I |
| Ammonia | 2 | 0.826 | 0.8 | 353 | 349 | 9.35 | 8.99 | 241 | 186 | 10.4 | 10.5 | 1.31 | 1.8 | 0.762 | 0.716 | 353 | 0.716 |
| Bicarbonate Alkalinity | | 274 | 260 | 1980 | 1980 | 1100 | 2000 | 2010 | 1550 | 289 | 275 | 902 | 620 | 350 | 4.67 | 2010 | 4.67 |
| Carbonate Alkalinity | | 2 U | 2 U | 2 U | 2 U | 2 U | 2 U | 2 U | 2 U | 2 U | 2 U | 2 U | 2 U | 2 U | 85.3 | 85.3 | 2 U |
| Chloride | 250 | 111 | 98.5 | 177 | 148 | 579 | 914 | 62.5 | 44.6 | 4.2 | 11 | 249 | 464 | 87.1 | 148 | 914 | |
| Ferrous Iron | | NA | NA | NA | NA | NA | NA | 0.524 | NA | NA | 22 | NA | NA | NA | NA | 22 | 0.524 |
| Nitrate Nitrogen | 10 | 0.5 U | 0.5 U | 0.5 U | 0.0548 | 0.5 U | 0.715 | 0.5 U | 0.05 U | 0.5 UJ | 0.484 | 0.5 U | 0.0753 | 0.5 UJ | 0.548 | 0.715 | 0.05 U |
| Soluble Organic Carbon | | 1.43 U | 7.87 J | 42.1 | NA | 12.275 | 12.7 1 | 40.925 | 39.625 J | 3.5325 U | NA | 3.835 U | NA | 2.5925 U | 16.15 | 42.1 | 1.43 U |
| Sulfate | 250 | 2.07 | 10.4 | 715 | 745 | 13.2 | 3.4 | 162 | 266 | 1620 | 1460 | 30.5 | 50.8 | 17.4 | 30.7 | 1620 | 2.07 |
| Total Alkalinity | | 274 | 260 | 1980 | 1980 | 1100 | 2000 | 2010 | 1550 | 289 | 275 | 902 | 620 | 350 | 90 | 2010 | 90 |
| Total Dissolved Solids | | 451 | NA | 1930 | NA | 1980 | NA | 1170 | NA | 2460 | NA | 1070 | NA | 533 | NA | 2460 | 451 |
| Total Kjeldahl Nitrogen | | 1.37 | 1.37 | 336 | 351 | 11.3 | 10.4 | 2.38 | 181 | 10.1 | 10.2 | 2.06 | 2.73 | 1.29 | 2.05 | 351 | 1.29 |
| Total Organic Carbon | | 1.06 U | 3.165 J | 38.9 | 31.45 | 11.675 | 10.6 J | 41.775 | 31.625 J | 3.6975 U | 5.31 | 3.775 U | 5.495 | 1.61 | 15.45 | 41.775 | 1.06 U |
| Total Sulfide | 0.05* | ΙU | 1.2 | 9.7 | 6.4 J | 10 | 2 UJ | 7.6 | 6.8 | ΙU | 2 U | 10_ | 1.2 J | 1.0 | 2 U | 9.7 | 10 |
| Field Measured Parameters ⁵ | | | | | | | | | | | | | | | | | l |
| pH (pH units) | | 7.77 | 5.98 | 6.59 | 6.54 | 7.2 | 6.12 | 6.73 | 6.47 | 6.19 | 4.83 | 7.23 | 6.28 | 7.9 | 10.68 | 10.68 | 4.83 |
| Conductivity (uS/cm) | | 528 | 826 | 4802 | 1595 | 2454 | 5006 | 2672 | 3214 | 1920 | 2538 | 1689 | 1642 | 994 | 711 | 5006 | 528 |
| Temperature (°C) | | 11.37 | 15.48 | 13.51 | 13.59 | 14.82 | 12.14 | 13.01 | 19.75 | 13.4 | 12.23 | 12.05 | 12.97 | 14.94 | 17.15 | 19.75 | 11.37 |
| Turbidity (NTU) | | >1000 | 579 | 2 | 9.6 | 51.8 | 276.5 | 72 | 321 | 3.34 | 16.1 | 2.89 | 35.8 | 33 | 1049 | 1049 | 2 |
| Oxidation Reduction Potential (mV) | | -191.6 | -149 | -283.3 | -112.2 | -92.8 | -46.9 | -330.5 | -266.6 | -94.9 | 6.6 | -146.5 | -9.1 | -35.9 | 202.5 | 202.5 | -330.5 |
| Dissofved Oxygen (ppm) | | 2.94 | 0.77 | 0.63 | 7.32 | 3.1 | 7.9 | 8.31 | 1.26 | 1.73 | 0.45 | 0.91 | 0.91 | 4.49 | 2.82 | 8.31 | 0.45 |
| Ferrous Iron (mg/l) | | 0.6 | 0 | 0 | 0 | 0 | 0.6 | 0.2 | 0 | 4.5 | 6.4 | 0.2 | ō | 0 | 0 | 6.4 | 0 |

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

 Groundwater criteria for Class GA groundwater as provided in Division of Water Technical and Operational Series (1.1.1). Ambient Water Quality Standards and Guidance Values and Groundwater Effluent Limitations, October 22, 1993, reissued June 1998.

Values are guidance values.

4. Samples collected for soluble metals analysis were field filtered.

5. The YSI 600XL was used in the November and May sampling events for temperature, pH, specific electrical conductance, dissolved oxygen, and redox potential measurements. Ferrous iron was field measured with the HACH18-R field kit (for QC, 10% were sent to analytical laboratory). The turbidity measurements on the YSI 600 XL were not accurate during the May sampling event and as such, the LaMotte turbidity meter was used to measure turbidity. Turbidity measurements were collected with the TURB2020 meter during the November sampling event.

NA = not analyzed

 $\sim \approx$ indicates value does not exist.

mg/3 = milligrams per liter

NTU = Nephelometric Turbidity Unit

uS/cm = microsiemens per centimeter at 25°C.

ppm = parts per million

mV = millivolts

- J = indicates an estimated value. U = indicates compound was not detected. Re indicates data rejected by data validator. (value) = indicates value reported before data validation. indicates exceedance of groundwater criteria. UJ = indicates compound was not detected above the listed detection firmit. However, the reported quantitation limit is approximate and may or may
- not represent the actual limit of quantitiation necessary to accurately

and precisely measure the compound in the sample.



ANALYTICAL RESULTS FOR OVERBURDEN GROUNDWATER FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Site Gowanda, New York

| | [| Sample | Location, Identifi | cation and Date | Collected | ſ | |
|---------------------------------------|-----------------------|-----------|--------------------|-----------------|--------------|-------------|-------------|
| | Į | MW | FP-2S | MV | VFP-3S | t | 1 |
| | Groundwater | 110700106 | 050301140 | 110700088 | 050201128 | Maximum | Minimum |
| Constituent ² | Criteria ³ | 11/7/2000 | 5/3/2001 | 11/7/2000 | 5/2/2001 | Conc | Conc |
| Volatile Organic Compounds micrograms | | 11///2000 | 5/5/2001 | 11,72000 | 57272001 | | Conc. |
| ner liter |] | | | | | | 1 |
| Acetope | 50* | 27 | NA | 10.11 | NA | | 10.11 |
| Benzene | | 10 U | 10 U | 10 U | 10 U | 10 U | 101 |
| Bromodichloromethane | 50* | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Bromoform | 50* | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Bromomethane | 5 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2-Butanone (MEK) | 50* | 10 U | NA | 10 U | NA | 10 U | 100 |
| Methyl tert-Butyl Ether | | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Carbon Disulfide | | 10 U | . NA | 10 U | NA | 10 U | <u>10 U</u> |
| Carbon Tetrachloride | 5 | 10 U | 10 U | 10 U | <u>10 U</u> | 10 U | 10 U |
| Chlorobenzene | 5 | 10 U | NA | <u> </u> | NA | 10 U | <u>10 U</u> |
| Chloroethane | 5 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Chloroform | 7 | 10 U | 10 U | <u>10 U</u> | <u>.10 U</u> | 10 U | 10 U |
| Chloromethane | | 10 U | NA | <u>10 U</u> | NA | 10 U | 10 U |
| 1.2-Dibromo-3-chloropropane | 0.04 | 10 0 | NA | 10 U | | 101 | 10 U |
| | - | 11 | NA | 10 0 | NA | 11 | 101 |
| 1.2. Dibromoethane | | 10 0 | | 10 0 | | 10.0 | 100 |
| 1.2-Dichlorobenzene | | 10 11 | NA NA | 10 0 | NA NA | 101 | 1010 |
| 1 4-Dichlorobenzene | 3 | 10 U | NA | 10 11 | NA | 1011 | 1017 |
| 1.3-Dichlorobenzene | 3 | 10 U | NA | 10 U | NA | 10 U | 100 |
| Dichlorodifluoromethane | 5 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 1.1-Dichloroethane | 5 | 10 U | NA | 2 J | NA | 10 U | 2 J |
| 1,2-Dichloroethane | 0.6 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 1.1-Dichloroethene | 5 | 10 Ü | NA | 10 U | NA | 10 U | 10 U |
| trans-1,2-Dichloroethene | 5 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| cis-1,2-Dichloroethene | 5 | 10 U | NA | 5 J | NA | 10 U | 51 |
| 1.2-Dichloropropane | 1 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| trans-1,3-Dichloropropene | 0.4 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| cis-1.3-Dichloropropene | 0.4 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Ethylbenzene | 5 | 10 U | NA | 10 U | NA | 10 U | 100 |
| 2-Hexanone | 50* | 10 U | <u>NA</u> | <u>10 U</u> | NA | 10 U | 10 U |
| Isopropylbenzene | 5 | 10 U | NA | 10 U | NA | 10 U | <u>10 U</u> |
| Methyl Acetate | | 10 0 | | 10 U | NA | 10 U | 100 |
| Methylepe Chloride | | 10 | NA NA | 10 U | NA | 10 | 100 |
| A.Methyl, 2. Pentanons | | 10 U | NA NA | 10 0 | NA | 100 | 101 |
| Styrene | | 10 U | NA | 10 U | NA NA | 10.0 | 100 |
| 1 1 2 2. Tetrachloroethane | | 10 1 | NA | 10 11 | NA | 101 | 100 |
| Tetrachloroethene | 5 | 10 U | 10.11 | 551 | 311 | 101 | 311 |
| Toluene | 5 | 10 U | NA | 10 U | NA | 10 U | 10.0 |
| 1.2.4-Trichlorobenzene | 5 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 1,1,1-Trichloroethane | 5 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 1,1,2-Trichloroethane | 1 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Trichloroethene | 5 | 10 U | 10 U | 2.9 J | 3.6 J | 10 U | 2.9 J |
| Trichlorofluoromethane | 5 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 1,1,2-Trichloro-1,2,2-Trifluoroethane | 5 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Vinyl Chloride | 2 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| m-/p-Xylene | 5 | 4.6 | NA | 10 U | NA | 10 U | 4.6 |
| o-Xylene | 5 | 1.9 | <u>NA</u> | <u>10 U</u> | NA | <u>10 U</u> | 1.9 |
| Semi-Volatile Organic Compounds, | | | | | | | |
| micrograms per liter | | | | | | | |
| Acenaphthene | 20* | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Acenaphthylene | | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Acetophenone | | 10 U | NA | <u>10 U</u> | NA | 100 | 10 U |
| Anthracene | 50* | 10 U | NA | 10 U | NA | | 10 0 |
| Atrazine | 1.5 | 10 0 | NA | 10 U | NA | - 10 0 | |
| Benzaldehyde | | 10 U | NA | 10 U | NA | 100 | - 10 U |
| Benzo(a)aninracene | 0.002 | 10 0 | 9.50 | 10 0 | | | 9.4 U |
| Benzo(a)pyrene | | 10 0 | 9.5 U | 10 0 | 9.4 U | 100 | 9,40 |
| Benzo(a h i)pervlene | <u> </u> | 10 0 | 9.3 U | 10 0 | 9.4 U N'A | 100 | 9.40 |
| Benzo(E)fluoranthene | 0.002* | 10 0 | NA NA | 10 0 | | 101 | 10.11 |
| Dereo (Katuoranniene | 0.00~ | 100 | 111 | 10 0 | | 100 | 100 |

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TABLE 4-10

ANALYTICAL RESULTS FOR OVERBURDEN GROUNDWATER FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Site Gowanda, New York

| | Sample Location, Identification and Date Collected ¹ MWFP-3S MWFP-3S | | | | | | |
|------------------------------------|--|-------------|------------|-------------|------------|---------|-------------|
| | | MW | FP-2S | MV | WFP-3S |] | |
| | Groundwater | 110700106 | 050301140 | 110700088 | 050201128 | Maximum | Minimum |
| Constituent ² | Criteria 3 | 11/7/2000 | 5/3/2001 | 11/7/2000 | 5/2/2001 | Conc. | Conc. |
| 1.1-Biphenyl | 5 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Butyl Benzyl Phthalate | 50* | 10 U | NA | 10 U | NA | 10 U | 10 U |
| di-N-Butylphthalate | - | 10 U | NA | 1.1 J | NA | 10 U | 1.1 J |
| Caprolactam | | 290 D | NA | <u>10 U</u> | NA | 290 D | <u>10 U</u> |
| Carbazole | | 10 U | NA | <u>10 U</u> | NA | 10 U | 10 U |
| Indeno(1,2.3-cd)pyrene | 0.002* | 10 U | 9.3 U | 10 0 | 9.4 U | 100 | 9.4 0 |
| +-Chloroethory)methane | | 10 U | | 10 0 | | 100 | 100 |
| bis(2-chloroethy))ether | | 10 1 | NA | 10 1 | NA | 10 U | 100 |
| 2-Chloronaphthaiene | 10* | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2-Chlorophenol | | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2,2-oxybis(1-chloropropane) | - | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Chrysene | 0.002* | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Dibenzo(a,h)anthracene | | 10 U | 9.5 U | 10 U | 9.4 U | 10 U | 9.4 Ŭ |
| Dibenzofuran | | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 3.3-Dichlorobenzidine | | 10 U | NA | 10 U | NA | 10 U | <u>10 U</u> |
| 2.4-Dichlorophenoi | 5 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Dimethyl Phthalate | 50* | 10 U | | 10 0 | NA NA | 100 | 100 |
| 2 4.Dimethylohenol | 50* | 10 0 | | 10 0 | NA NA | 100 | 1011 |
| 2.4-Dinitrophenol | 10* | 25 11 | NA | 25 11 | NA | 25 U | 2511 |
| 2.4-Dinitrotoluene | 50 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2.6-Dinitrotoluene | 5 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| bis(2-Ethylhexyl)phthalate | - | 4 J | NA | 10 U | NA | 10 U | <u>4 J</u> |
| Fluoranthene | 50* | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Fluorene | 50* | 10 U | NA | 10 U | NA | 10 U | <u>10 U</u> |
| Hexachlorobenzene | 0.04 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Hexachlorobutadiene | 0.5 | 10 U | NA | <u>10 U</u> | NA | 10 U | 10 U |
| Hexachiorocyclopentadiene | | 10 U | NA | 10 U | NA | 100 | 100 |
| Isophorope | 50+ | 10 U | NA | 10 0 | NA | 10 U | 100 |
| 2-Methylnaphthalene | | 10 U | NA | 10 11 | NA | 10 U | 100 |
| 4.6-Dinitro-2-Methylphenol | | 25 U | NA | 25 U | NA | 25 U | 25 U |
| 4-Chloro-3-Methylphenol | | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2-Methylpheno) | | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 4-Methylphenol | | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Naphthalene | 10* | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2-Nitroaniline | 5 | 25 U | NA | 25 U | NA | 25 U | 25 U |
| 3-Nitroaniline | 5 | 25 U | NA | 25 U | NA | 25 U | 25 U |
| 4-Nitroaniline | 5 | 25 U | NA | 25 U | NA | 25 U | 25 U |
| Nitrobenzene | 0.4 | <u>10 U</u> | NA | <u>10 U</u> | NA | 10 U | <u>10 U</u> |
| 2-Nitrophenol | | 10 U | NA | 10 0 | NA | 100 | 10 U |
| 4-Nitropheno/ | 50* | 25 U | NA | 25 U | NA | 250 | 25 0 |
| di n. Octul Phthalate | | 10 U | NA | 10 0 | NA | 100 | 10 U |
| Pentachloronhenol | 1 | 25 U | NA | 25 1 | NA | 2511 | 25 U |
| Phenanthrene | 50* | 10 U | NA | 10 U | NA | 1011 | 1011 |
| Phenol | 1 | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 4-Bromophenyl-Phenylether | | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 4-Chlorophenyl-Phenylether | _ | 10 U | NA | 10 U | NA | 10 U | 10 U |
| n-Nitroso-di-n-Propylamine | | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Рутеле | 50* | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2.4,6-Trichlorophenol | | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2.4,5-Trichlorophenol | | 25 U | NA | 25 U | NA | 25 U | 25 U |
| Total Metals, milligrams per liter | | | | | | | |
| Aluminum | | 0.331 | NA | 0.406 | NA | 0.406 | 0.331 |
| Antimony | 0.003 | 0.06 U | NA | 0.06 U | NA | 0.06 U | 0.06 U |
| Arsenic | 0.025 | 0.01 U | NA | 0.01 U | NA | 0.01 U | 0.01 0 |
| Banum | 1.0 | 0.112 | NA | 0.103 | <u>NA</u> | 0.00511 | 0.00511 |
| Cadmium | 0.005 | 0.005 0 | | 0.005 0 | | 0.005.0 | 0.005.0 |
| Calcium | 5.005 | 313 | 337 | 360 | 312 | 360 | 312 |
| Chromium | 0.050 | 0.0114 | 14 | 0.01 U | 0.01 Ú | 14 | 0.01 U |
| Hexavalent Chromium | 0.050 | 0.01 U | (0.02 U) R | 0.01 U | (0.01 U) R | 0.01 U | (0.01 U) R |

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ANALYTICAL RESULTS FOR OVERBURDEN GROUNDWATER FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Site Gowanda, New York

| | | Collected ' | | | | | |
|--|-----------------------|-------------|-----------|-----------|-----------|----------|--------------------|
| | | MW | FP-2S | М₩ | FP-3S | 1 | |
| | Groundwater | 110700106 | 050301140 | 110700088 | 050201128 | Maximum | Minimum |
| Constituent ² | Criteria ³ | 11/7/2000 | 5/3/2001 | 11/7/2000 | 5/2/2001 | Conc. | Conc. |
| Cobalt | | 0.05 U | NA | 0.05 U | NA | 0.05 U | 0.05 U |
| Copper | 0.200 | 0.02 U | NA | 0.02 U | NA | 0.02 U | 0.02 U |
| Iron | 0.300 | 0.535 | 4.21 | 16 | 5.51 | 16 | 0.535 |
| Lead | 0.025 | 0.005 U | 0.005 U | 0.005 U | 0.005 U | 0.005 U | 0.005 U |
| Magnesium | 35* | 32.8 | 26.4 | 17.5 | 17 | 32.8 | 17 |
| Manganese | 0.300 | 0.43 | 0.68 | 2.08 | :1.49 | 2.08 | 0.43 |
| Mercury | 0.0007 | 0.0003 U | NA | 0.0003 U | NA | 0.0003 U | 0.0003 U |
| Nickel | 0.100 | 0.04 U | NA | 0.04 U | NA | 0.04 U | 0.0 4 U |
| Potassium | 1 | 10.7 | 6.41 | 6.6 | 4.63 | 10.7 | 4.63 |
| Selenium | 0.010 | 0.005 U | NA | 0.0061 | NA | 0.0061 | 0.005 U |
| Silver | 0.050 | 0.01 U | NA | 0.01 U | NA | 0.01 U | 0.01 U |
| Sodium | 20 | 18.7 | 9.98 | 122 | 45.9 | 122 | 9.98 |
| Thallium | 0.0005* | 0.01 U | NA | 0.01 U | NA | 0.01 U | 0.01 Ū |
| Vanadium | | 0.05 U | NA | 0.05 U | NA | 0.05 U | 0.05 U |
| Zinc | 2* | 0.124 | NA | 0.0551 | NA | 0.124 | 0.0551 |
| Other Geochemical Parameters, milligrams | | | | | | | |
| per liter | | | | | | | |
| Total Organic Carbon | 1 | 8.6375 U | 4.77 U | 5.19 Ū | 3.15 U | 8.6375 U | 3.15 U |
| Bicarbonate Alkalinity | _ | 700 | 680 | 558 | 435 | 700 | 435 |
| Carbonate Alkalinity | - | 2 U | 2 U | 2 U | 2 U | 2 U | 2 U |
| Chloride | 250 | NA | 10 | NA | 63.5 | 63.5 | 10 |
| Ferrous Iron | 1 | NA | 4.95 | NA | NA | 4.95 | 4.95 |
| Solubie Organic Carbons | - | 9.22 U | NA | 5.405 U | NA | 9.22 U | 5.405 U |
| Sulfate | 250 | : 346 | 301 | 651 | 448 | 651 | 301 |
| Total Alkalinity | | 700 | 680 | 558 | 435 | 700 | 435 |
| Total Dissolved Solids | - | 1190 | 1170 | 1570 | 1180 | 1570 | 1170 |
| Total Sulfide | 0.05* | 1 U | 2 UJ | 1.1 U | 2 UJ | 2 UJ | 10 |
| Field Measured Parameters ⁴ | | | | | | | |
| Conductivity (uS/cm) | - | 1588 | 1559 | 2136 | 1513 | 2136 | 1513 |
| Dissolved Oxygen (ppm) | - | 3.1 | 4.81 | 0.35 | 0.42 | 4.81 | 0.42 |
| Ferrous Iron (mg/l) | - | 0 | 3.0 | 6.50 | 5.2 | 5.2 | 0 |
| Oxidation Reduction Potential (mV) | - | 82 | -31.6 | -89.7 | -17 | 82 | -31.6 |
| pH (pH units) | - | 7.01 | 6.16 | 6.70 | 5.9 | 7.01 | 5.9 |
| Temperature (°C) | | 12.83 | 13.9 | 14.27 | 12 | 14.27 | 12 |
| Turbidity (NTU) | - | NA | 59 | 35.1 | 191 | 191 | 59 |

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

 Groundwater criteria for Class GA groundwater as provided in Division of Water Technical and Operational Series (1.1.1), Ambient Water Quality Standards and Guidance Values and Groundwater Effluent Limitations, October 22, 1993, reissued June 1998.

* Values are guidance values.

4. The YSI 600XL was used in the November and May sampling events for temperature, pH, specific electrical conductance, dissolved oxygen, and redox potential measurements. Ferrous iron was field measured with the HACH18-R field kit (for QC, 10% were sent to analytical laboratory). The turbidity measurements on the YSI 600 XL were not accurate during the May sampling event and as such, the LaMotte turbidity meter was used to measure Turbidity. Turbidity measurements were collected with the TURB 2020 meter during the November sampling events.

mgA = milligrams per liter NA = not analyzed NTU = Nephelometric Turbidity Unit uS/cm = microsiemens per centimeter at 25°C. ppm = parts per million mV = millivolts (values) = laboratory reported value prior to data validation J = an estimated concentration.

U = compound was not detected at or above the listed detection limit.

 \mathbf{R} = value was rejected by data validator.

D = indicates spike diluted out.

- = indicates value does not exist.

indicates exceedance of groundwater criteria. UJ = indicates compound was not detected above the listed detection limit.

However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitation necessary to accurately and precisely measure the compound in the sample.



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TABLE 4-11

ANALYTICAL RESULTS FOR BEDROCK GROUNDWATER FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Site Gowanda, New York

| | Sample Location, Identification, and Date Collected' MWFP-1D MWFP-2D MWFP-1D | | | | | | | | |
|--|---|-------------|-------------------|-----------|-------------------|-------------|-------------------|---------|---------|
| | | MWI | P-1D | MWFI | P-2D | MW | FP-JD | | |
| | Groundwater | 110600086 | 0 50101125 | 110600087 | 0 50201135 | 110700090 | 0 50101126 | Maximum | Minimum |
| Compound | Criteria | 11/6/2000 | 5/1/2001 | 11/6/2000 | 5/2/2001 | 11/7/2000 | 5/1/2001 | Conc. | Conc. |
| Volatile Organic Compounds, micrograms per | | | | | | | | | |
| Actions | 50* | 10 1/ | NA | 80 | NA | 671 | NA | 80 | 671 |
| Benzene | 1 | 10 U | 10 U | 3.6 J | . 2.4] | 10 U | 1.2 J | 10 U | 1.2 J |
| Bromodichloromethane | 50* | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Bromoform | 50* | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Bromomethane | 50* | 10 U | NA NA | 10 U | NA | 10 U | NA NA | 100 | 10 (|
| Methyl tert-Buryl Ether | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 100 |
| Carbon Disulfide | | 10 U | NA | 1.3 J | NA | 10 U | NA | _10 U | 1.3 J |
| Carbon Tetrachloride | 5 | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 Ü | 10 U |
| Chlorobenzene | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Chloroform | 7 | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 100 | 1010 |
| Chloromethane | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 1,2-Dibromo-3-chkoropropane | 0.04 | 10 U | NA | 10 U | NA | 10 Ŭ | NA | 10 U | 10 U |
| Cyclohexane | | 10 U | NA | 14 | NA | 8.8 J | NA | 14 | 8.8 J |
| Dibromochloromethane | 50* | 10 U | NA | 10 U | NA NA | 10 U | NA | 10 U | 10 U |
| 1.2-Dibromoethane | | 10 U | NA | 10 0 | NA | 10 0 | NA | 1010 | 10 11 |
| 1.4-Dichlorobenzene | 3 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 1,3-Dichlorobenzene | 3 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Dichlorodifluoromethane | 5 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 1.1-Dichloroethane | - 5 | 10 U | NA | 10 U | NA | 2.3 J | NA | 10 U | 2.3 J |
| 1,2-Dichloroethane | <u> </u> | 10 0 | NA NA | 10 U | NA NA | 10 U | NA NA | 10 0 | 10 U |
| rans-12-Dichlamethene | | 10 U | NA | 10.0 | NA | 10 U | NA | 1010 | 1010 |
| cis-1.2-Dichloroethene | 5 | 10 U | NA | 10 U | NA | 8.2 J | NA | 10 U | 8.2 J |
| 1.2-Dichloropropane | 1 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| trans-1,3-Dichloropropene | 0.4 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| cis-1.3-Dichloropropene | 0.4 | <u>10 U</u> | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Ethylbenzene | | | <u>NA</u> | 10 U | <u>NA</u> | 10 0 | NA NA | 10 U | 10 U |
| Isopmyybenzene | 5 | 10 U | NA | 10 U | NA | 10 10 | NA | 10 U | 10 U |
| Methyl Acetate | - | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Methylcyclohexane | | 10 U | NA | 15 | NA | 4.9 J | NA | 15 | 4.9 J |
| Methylene Chloride | 5 | <u>10 U</u> | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 4-Methyl-2-Pentanone | | <u>10 U</u> | NA NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 1 1 2 2 Tetrachlomethane | | 10 U | NA | 10 U | NA | 10 U | NA | 1010 | 100 |
| Tetrachloroethene | 5 | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U |
| Toluene | 5 | 10 U | NA | 6.8 J | NA | 10 U | NA | 10 U | 6.8 J |
| 1.2.4-Trichlorobenzene | 5 | 10 U | NA | 10 U | NA | <u>10 U</u> | NA | 10 U | 10 U |
| 1.1.1-Trichloroethane | 5 | <u>10 U</u> | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 1.1.2-1 richioroethane | | <u>10 U</u> | <u>NA</u> | 10 U | NA | 10 U | | 10 U | 10 U |
| Trichloroftuoromethane | | 10 0 | NA | 10 U | NA | 10 U | NA | 1010 | 100 |
| 1,1,2-Trichloro-1.2 2-Trifluoroethane | 5 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Vinyl Chloride | 2 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| m-/p-Xylene | 5 | 10 U | NA | 6.4 J | NA | 10 U | NA | 10 U | 6.4 J |
| o-Xylene | 5 | 10 U | <u>NA</u> | 3.7 J | NA | 10 U | <u>NA</u> | 10 U | 3.7 J |
| Semi-Volatile Organic Compounds, microgram | | | | | | | | | |
| Acchaphthene | 20* | 10 U | NA | 10.11 | NA | 10.11 | NA | 10 U | 10 U |
| Acenaphthylene | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Acetophenone | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Anthracene | 50* | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Atrazine | 7.5 | 10 U | NA NA | 10 U | NA NA | 10 U | NA NA | 100 | 10 U |
| Benzeisenver | 0.002 | 10 U | 941 | 10 0 | 9511 | 10 0 | 9411 | 100 | 9411 |
| Benzo(a)pyrene | | 10 U | 9.4 U | 10 U | 9.5 U | 10 U | 9.4 U | 100 | 9.4 U |
| Benzo(b)fluoranthene | 0.002* | 10 U | 9.4 U | 10 U | 9.5 U | 10 U | 9.4 U | 10 U | 9.4 U |
| Benzo(g.h.i)perylene | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Benzo(k)fluoranthene | 0.002* | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 1,1-Biphenyl | 5 | 10 U | NA | 10 U | NA | 10 U | NA | 100 | 10 U |
| Duiyi Denzyi munalare | | 10 0 | | 10 U | NA NA | 10 U | <u>NA</u> | 1011 | 1010 |
| Caprolactam | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Carbazole | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Indeno(1.2.3-cd)pyrene | 0.002* | 10 U | 9.4 U | 10 U | 9.5 U | 10 U | 9.4 U | 10 U | 9.4 U |
| 4-Chloroaniline | 5 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| bis(2-chloroethoxy)methane | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| bisi 2-chloroethyl)ether | | 10 U | NA | 10 U | <u>NA</u> | 10 U | NA | 10 U | 100 |
| 2-Chlorophenol | | 10 0 | | 10 U | NA | 10 0 | NA NA | 1011 | 10 11 |
| 2,2-oxybis(1-chloropropane) | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Chrysene | 0.002* | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 100 |

1 Project004771 PRP Group Deer Cooper NPLR1 reportFTNAL REPORT (November 2003 Seberginal (Tables (Final)(Table 4-11 Man. plant water DEEP FINAL



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TABLE 4-11

ANALYTICAL RESULTS FOR BEDROCK GROUNDWATER FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Site Gowanda, New York

| | | MW | FP-ID | MWF | P-2D | MW | FP-3D |] | |
|--|-------------|-----------------|------------|-----------|------------|-------------|------------|-------------|--------------|
| , | Groundwater | 110600086 | 050101125 | 110600087 | 050201135 | 110700090 | 050101126 | Maximum | Minimum |
| Compound ¹ | Criteria' | 11/6/2000 | 5/1/2001 | 11/6/2000 | 5/2/2001 | 11/7/2000 | 5/1/2001 | Conc. | Conc. |
| Dibenzo(a,h)anthracene | · · · · | 10 U | 9.4 U | 10 U | 9.5 U | 10 U | 9.4 U | 10 U | 940 |
| Dibenzofuran | | 10 U | NA | 10 U | NA | 10 U | NA NA | 10 U | 10 (|
| 3.3-Dichlorobenzidine | | 10 U | NA | 10 U | <u>NA</u> | 10 U | <u>NA</u> | 101 | 10 U |
| 2.4-Dichlorophenol | 50. | 10 U | NA | 10 U | NA | | NA | 100 | 101 |
| Dimethyl Phthalatc | 50* | 10 U | NA | 100 | NA | 10 U | NA | 10 U | 10 U |
| 2,4-Dimethylphenol | 50* | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2.4-Dinitrophenol | 10* | 25 U | NA | 25 U | NA | 25 U | NA | 25 U | 25 U |
| 2.4-Dinitrotohuene | 50 | <u>10 U</u> | NA | 10 U | NA | 10 U | NA | 10 U | 100 |
| 2.6-Dinitrotoluene | 5 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| bis(2-Ethylhexyl)phthalate | | 10 U | <u>NA</u> | 3.7 J | NA | 10 U | NA | 10 U | 3.71 |
| Fuorantiene | 50* | 10 U | NA | 100 | NA NA | 10 0 | NA NA | 100 | 10.0 |
| Hesachioropenzene | 0.04 | 10 U | NA | 10 U | NA NA | 10 11 | NA | 100 | 100 |
| Hexachlorobutadiene | 0.5 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Hexachlorocyclopentadiene | 5 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Hexachloroethanc | 5 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Isophorone | 50* | 10 U | NA | 10 U | NA | <u>10 U</u> | NA | 10 U | 10 U |
| 2-Methylnaphthalene | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 100 |
| 4.6-Dinitro-2-Methylphenol | | 25 0 | NA NA | 25 U | NA | 25 U | NA | 25 0 | 25 U |
| 2-Methylphenol | | 10 0 | NA | 10.11 | NA NA | 10 0 | NA NA | 10 U | 10.17 |
| 4-Methylphenol | | 10 U | NA | 10 U | NA | 10 U | NA | 100 | 10 U |
| Naphthalene | 10* | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2-Nitroaniline | 5 | 25 U | NA | 25 U | NA | 25 U | NA | 25 U | 25 U |
| 3-Nitroaniline | 5 | 25 U | NA | 25 U | NA | 25 U | NA | 25 U | 25 Ú |
| 4-Nitroaniline | 5 | 25 U | NA | 25 U | NA | 25 U | NA | 25 U | 25 U |
| Nitrobenzene | 0.4 | 10 U | NA NA | 10 U | NA | 10 U | NA | <u>10 U</u> | 10 0 |
| 2-Nitrophenol | | 25.11 | NA | 10 0 | NA | | NA NA | 10 C | 10 0 |
| N. Nitrosodiobenylamity | 50* | 10 U | NA | 10 11 | NA | 10 1 | NA | 1010 | 1011 |
| di-n-Octvl Phthalate | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Pentachlorophenol | 1 | 25 U | NA | 25 U | NA | 25 U | NA | 25 U | 25 U |
| Phenanthrene | 50* | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | IQU |
| Phenol | 1 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 4-Bromophenyl-Phenylether | | 10 U | NA | 10 U | NA | 10 U | NA | <u>10 U</u> | 10 U |
| a Nitmondia Providence | | 10 0 | | 10 U | NA NA | 10 U | NA NA | 10 0 | 100 |
| Pyrenc | 50* | 10 U | NA | 10 0 | NA | 10 1 | NA | 100 | 100 |
| 2.4.6-Trichkorophenoi | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2.4.5-Trichlorophenol | | 25 U | NA | 25 U | NA | 25 U | NA | 25 U | 25 U |
| Metals, milligrams per liter | | | | | | | | | |
| Aluminum | | 0.12 | NA | 0.641 | NA | 0.116 | NA | 0.641 | 0.116 |
| Antimony | 0.003 | 0.06 U | NA | 0.06 U | NA | 0.06 U | NA | 0.06 U | 0.06 U |
| Arsetuc | 0.025 | 0.01 U | NA | 0.01 U | NA | 0.01 U | NA | 0.01 U | 0.01 U |
| Benlinn | 0.003 | 0.275 | NA | 0.07/5 | NA NA | 0.0722 | NA | 0.275 | 0.0722 |
| Cadmium | 0.005 | 0.005 U | NA | 0.005 U | NA | 0.005 U | NA | 0.005 U | 0.005 U |
| Calcium | | 62 | 64.5 | 18.9 | 28.8 | 370 | 348 | 370 | 18.9 |
| Chromium | 0.05 | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U |
| Hexavalent Chromium | 0.05 | 0.01 U | (0.01 U) R | 0.01 U | (0.01 U) R | 0.01 U | (0.01 U) R | 0.01 U | 0.01 U |
| Cobalt | | 0.05 U | NA | 0.05 U | NA | 0.05 U | NA | 0.05 U | 0.05 U |
| Copper | 0.2 | 0.02 U | NA | 0.02 U | NA | 0.02 U | NA | 0.02 U | 0.02 U |
| tron | 0.3 | 0.417 | 0.211 | 1.89 | 0.348 | 21.5 | 17.7 | 21.5 | 0.211 |
| Magnetium | 35= | <u>0.005 L'</u> | 10.005 U | 4.25 | 0.005 0 | 0.005 0 | 0.005 0 | 0.005 0 | 0.005 0 |
| Manganese | 03 | 0 112 | 0.127 | 9.23 | 0.0579 | 2.06 | 106 | 7 04 | 0.0446 |
| Mercury | 0.0007 | 0.0003 U | NA | 0.0003 U | NA | 0.0003 U | NA | 0.0003 U | 0.0003 U |
| Nickel | 0.1 | 0.04 U | NA | 0.04 U | NA | 0.04 U | NA | 0.04 U | 0.04 U |
| Potassium | - | 2 U | 2 U | 3.72 | 3.04 | 7.04 | 5.68 | 7.04 | 2 U |
| Selenium | 0.01 | 0.005 U | NA | 0.005 U | NA | 0.005 U | NA | 0.005 U | 0.005 U |
| Silver | 0.05 | 0.01 U | NA | 0.01 U | NA | 0.01 U | NA | 0.01 U | 0.01 U |
| Sodium | 20 | 26.7 | 25 | 293 | 352 | 119 | 78.9 | 352 | 25 |
| Thallium | 0.005* | 0.01 U | NA | 0.01 U | NA | 0.01 U | <u>NA</u> | 0.01 U | 0.01 U |
| Vanadium Zine | | 0.05 U | NA NA | 0.05 U | NA NA | 0.05 U | NA NA | 0.05 U | 0.02.11 |
| Call Marcal Annalist | | 0.02 0 | | 0.02 0 | | 0.02 0 | | 0.02 0 | 0.02.0 |
| Soluble Metals', milligrams per liter | 0.05 | N1A | | | | | 0.01.17 | 0.0111 | 0.0111 |
| Liuonuum Hexavalent Chromium | 0.05 | NA NA | | NA | NA NA | NA NA | 0.01 U | (0.01 U | (0.01 11) 10 |
| Imp | 0.03 | NA | NA NA | NA NA | NA NA | NA NA | 16.4 | 164 | 164 |
| Lead | 0.025 | NA | NA | NA | NA | NA | 0.005 U | 0.005 U | 0.005 U |
| Manganese | 0.3 | NA | NA | NA | NA | NA | 1.89 | 1.89 | 1.89 |
| Other Geochemical Parameters, milligrams per | | | | | | | | | |
| liter | | | | | | | | | |
| Total Organic Carbon | | 1.29 U | 1.78 U | 3.40 U | 6.41 U | 5.02 U | 4.36 U | 6.41 U | 1.2875 U |
| Bicarbonate Alkalinity | •• | 200 | 187 | 288 | 355 | 575 | 480 | 575 | 187 |
| Carbonate Alkalinity | | 2 U | 2 U | 2 U | 2 U | 20 | 2 U | 20 | 2 U |



ANALYTICAL RESULTS FOR BEDROCK GROUNDWATER FROM THE FORMER MANUFACTURING PLANT AREA

Peter Cooper Site Gowanda, New York

| | | MW | FP-ID | MWFP-2D | | MWFP-3D | | 1 | |
|--|-------------|-----------|-----------|-----------|-----------|-----------|-----------|----------|----------|
| | Groundwater | 110600086 | 050101125 | 110600087 | 050201135 | 110700090 | 050101126 | Maximum | Minimum |
| Compound ² | Criteria ' | 11/6/2900 | 5/1/2001 | 11/6/2000 | 5/2/2001 | 11/7/2000 | 5/1/2001 | Conr. | Conc. |
| Chloride | 250 | NA | 22.5 | NA | 166 | NA | 77.7 | 166 | 22.5 |
| Soluble Organic Carbons | | 9.11 U | NA | 3.31 U | NA | 6.02 U | 4.92 | 9.1075 U | 3.3125 U |
| Sulfate | 250 | 45.5 | 47.2 | 56.7 | 241 | 695 | 544 | 695 | 45.5 |
| Total Alkalinity | | 200 | 187 | 288 | 355 | 575 | 480 | 575 | 187 |
| Total Dissolved Solids | - | 290 | 293 | 917 | 1000 | 1660 | 1350 | 1660 | 290 |
| Total Sulfide | 0.05* | 1.1 U | _2 UJ | 1.1 U | 2 ປປ | 1.1 U | 2 UJ | 2UJ | 1.1 U |
| Turbidity, NTU | | NA | 0.8 | NA | NA | NA | 79,7 | 79.7 | 0.8 |
| Field Measured Parameters ⁵ | | | | | | | | | |
| Conductivity (uS/cm) | | 495 | 503 | 1616 | 1595 | 2316 | 2159 | 2316 | 495 |
| Dissolved Oxygen (ppm) | | 2.70 | 0.96 | 0.29 | 0.6 | 0.95 | 1.70 | 2.7 | 0.29 |
| Ferrous Iron (mg/l) | | 0.1 | 0 | 0 | 0 | 10 | 6.6 | 10 | 0 |
| Oxidation Reduction Potential (mV) | | -219.5 | -3.2 | -223.5 | -112.2 | -94.9 | -68.2 | -3.2 | -223.5 |
| pH (pH units) | | 7.01 | 5.59 | 8.62 | 6.54 | 6.7 | 6.61 | 8.62 | 5.59 |
| Temperature (°C) | | 13.75 | 17.47 | 13.06 | 13.59 | 14.35 | 11.46 | 17.47 | 11.46 |
| Turbidity (NTU) | | 3.5 | 12.2 | 3.6 | 9.6 | 8.55 | 24 | 24 | 3.5 |

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

3 Groundwater criteria for Class GA groundwater as provided in Division of Water Technical and Operational Series (1.1.1). Ambient Water Quality Standards and Guidance Values and Groundwater Effluent Limitations, October 22, 1993, reissued June 1998.

Values are guidance values.
 A. Samples collected for soluble metals analysis were field filtered.

5. The YSI 600XL was used in the November and May sampling events for temperature, pH, specific electrical conductance, dissolved oxygen, and redox potential measurements. Ferrous iron was field measured with the HACH18-R field kit (for QC, 10% were sent to analysical informatory). The turbidity measurements on the YSI 600 XL were not accurate during the May sampling event and as such, the LaMotte turbidity meter was used to measure turbidity Turbidity measurements were collected with the TURB 2020 meter during the November sampling events.

mg/l = miligrams per liter

NA = not analyzed

NTU = Nepheiometric Turbidity Unit

uS/cm = microsiemens per centimeter at 25°C.

ppm = parts per miliion mV = millivolts

(values) = laboratory reported value prior to data valudation

J = an estimated concentration

 $U\approx$ compound was not detected as or above the listed detection limit.

R = value was resorted by data validator

-- = indicates value does not exist. indicates exceedance of groundwater criteria.

UI = indicates compound was not detected above the listed detection limit.

However, the reported quantitation firmit is approximate and may or may not represent the actual limit of quantitation necessary to accurately and precisely measure the compound in the sample.



ANALYTICAL RESULTS FOR SEEP SAMPLES FROM THE INACTIVE LANDFILL AREA

Peter Cooper Site Gowanda, New York

| 1 | | See | p#1 | See | p #2 | Se | ep #3 | | |
|--|-----------------------|-------------|------------|-------------|-------------|-------------|------------|--------------|------------|
| | Surface Water | 110800102 | 052001137 | 110800103 | 052001138 | 110800104 | 052001139 | Maximum | Minimum |
| Constituent ² | Criteria ³ | 11/8/2000 | 5/20/2001 | 11/8/2000 | 5/20/2001 | 11/8/2000 | 5/20/2001 | Conc. | Conc. |
| Volatile Organic Compounds, | | | | | { | | | | |
| micrograms per liter | | | | 1 | | i | | l _ | |
| Benzene | 210* | 10 U | 10 U | 10 U | 10 UJ | 10 U | 10 UJ | 10 UJ | 10 UJ |
| Chlorobenzene | 5 | 10 U | 10 U | 10 U | 10 UJ | 10 U | 10 UJ | <u>10 UJ</u> | 10 UJ |
| 1,2-Dichlorobenzene | 5 | 10 U | 10 U | 10 U | 10 UJ | 10 U | 10 UJ | 10 UJ | 10 UJ |
| 1.4-Dichlorobenzene | 5 | 10 U | 10 U | 10 U | 10 UJ | 10 U | 10 UJ | 10 UJ | 10 UJ |
| Ethylbenzene | 17* | 10 U | 10 U | 10 U | 10 UJ | 10 U | 10 UJ | 10 UJ | 10 UJ |
| Toluene | 100* | 3.1 J | 2.8 J | 2 J | 3.5 J | 10 U | 10 UJ | 10 UJ | 2 J |
| m/p-Xylene | 5 | 10 U | 10 U | 10 U | 10 UJ | 10 U | 10 UJ | 10 UJ | 10 UJ |
| o-Xylene | 5 | 10 U | 10 U | 10 U | 10 UJ | 10 U | 10 UJ | 10 UJ | 10 UJ |
| | | | | | | | | | |
| Semi-Volatile Organic Compounds, | | | 1 | |] | | 1 | | |
| micrograms per liter | | | | | | | | L | |
| 2-Chorophenol | | 10 U | 9.4 0 | <u>10 U</u> | <u>10 U</u> | 10 U | 16 U | 16 U | 9.4 U |
| 2,4-Dichlorophenol | | 10 U | 9.4 U | 10 U | 10 U | 10 U | 16 U | 160 | 9.4 U |
| 2,4-Dimethylphenol | | 10 0 | 9.4 0 | 10 0 | 10 0 | 10 U | <u> </u> | 160 | 9.4 U |
| 2,4-Dinitrophenol | | 50 U | 24 U | 50 U | 25 U | <u>50 U</u> | 40 0 | <u>50 U</u> | 24 U |
| 4,6-Dintiro-2-methylphenol | | <u>50 U</u> | 24 U | 50 U | 25 U | <u>50 U</u> | 40 U | 50 0 | 240 |
| 4-Chloro-3-Methylphenol | | 10 U | 9.4 0 | · 10 U | 10 U | 10 U | 16 U | 160 | 9.4 0 |
| 2-Methylphenol | | 10 U | 9.4 U | 10 0 | 10 U | 10 0 | 16 U | 16 U | 9.4 U |
| 4-Methylphenol | | 10 () | 9.4 U | 10 U | 10 U | 10 U | 16 U | 16 U | 9.4 U |
| 2-Nitrophenol | | 10 U | 9.4 U | 10 U | 10 U | 10 U | 16 U | 16 U | 9.4 U |
| 4-Nitrophenol | | 50 U | 24 U | <u>50 U</u> | 25 U | 50 U | 40 U | 50 0 | 24 U |
| ^o entachlorophenol ⁴ | 20.2 | 50 U | 24 U | 50 U | 25 U | 50 U | 40 U | 50 U | 24 U |
| rhenol | | 10 U | 9.4 U | 1.8 J | 10 U | 1.8 J | 16 U | 16 U | 1.8 J |
| 2,4,6-Trichlorophenol | | 10 U | 9.4 U | 10 U | 10 U | 10 U | 16 U | 16 U | 9.4 U |
| 2.4,5-Trichlorophenol | | 10 U | 24 U | 10 U | 25 U | 10 U | 40 U | 40 U | 10 U |
| | | | | | 1 | | | | |
| Total Metals, milligrams per liter | | | | | | | | | |
| Arsenic | 0.150 | 0.071 | 0.052 | 0.0520 | 0.038 | 0.062 | 0.0314 | 0.071 | 0.0314 |
| | | 156 | 171 | 150 | 156 | 116 | 170 | 1/1 | 116 |
| Chromium | 0.120 | 0.374 | 0.221 | 0.423 | 0.312 | 0.0949 | 0.129 | 0.423 | 0.0949 |
| Hexavalent Chromum | 0.011 | 0.04 U | (0.01 U) R | 0.04 U | (0.01 U) R | 0.01 U | (0.01 U) R | 0.04 0 | 0.01 U |
| | 0.300 | - 3.01 | .1.18 | 28.6 | 0.1 U | 0.39 | 0.123 | 28.0 | 0.10 |
| Magnesium | | 190 | 102 | 163 | 123 | 82.9 | 90.5 | 190 | 82.9 |
| Potassium | | 10.9 | 7.71 | 8.79 | 6.19 | 3.56 | 4.12 | 10.9 | 3.56 |
| Sodium | | 26.8 | 18.1 | 19.7 | 18.3 | 17.5 | 18 | 26.8 | 17.5 |
| Zinc | 0.200 | 0.02 U | 0.02 U | 0.0747 | 0.02 U | 0.02 U | 0.02 U | 0.0747 | 0.02 U |
| , | | | | | | | | | |
| Soluble Metals [°] , milligrams per liter | | | | | | | | | |
| Arsenic | 0.15 | 0.0665 | NA | 0.0528 | NA | 0.0599 | NA | 0.0665 | 0.0528 |
| Calcium | | 155 | NA | 132 | NA | 113 | NA | 155 | 113 |
| Chromium | 0.120 | 0.369 | NA | 0.325 | NA | 0.0969 | NA | 0.369 | 0.0969 |
| Hexavalent Chromium | 0.011 | 0.04 U | (0.01 U) R | 0.04 U | (0.01 U) R | 0.04 U | (0.01 U) R | 0.04 U | (0.01 U) R |
| Iron | 0.3 | 4.78 | NA | 0.914 | NA | 0.107 | NA | 4.78 | 0.107 |
| Magnesium | | 184 | NA | 144 | NA | 84.1 | NA | 184 | 84.1 |
| Potassium | | 10.5 | NA | 6.4 | NA | 3.7 | NA | 10.5 | 3.7 |
| Sodium | | 26 | NA | 19.6 | NA | 17 | NA | 26 | 17 |
| Zinc | 0.200 | 0.02 U | NA | 0.02 U | NA | 0.02 U | NA | 0.02 U | 0.02 U |

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ANALYTICAL RESULTS FOR SEEP SAMPLES FROM THE INACTIVE LANDFILL AREA

Peter Cooper Site Gowanda, New York

| | | See | p #1 | See | p #2 | Seep #3 | |] | |
|--|-----------------------|----------------|----------------|----------------|----------------|-----------|-----------|---------|---------|
| | Surface Water | 110800102 | 052001137 | 110800103 | 052001138 | 110800104 | 052001139 | Maximum | Minimum |
| Constituent ² | Criteria ³ | 11/8/2000 | 5/20/2001 | 11/8/2000 | 5/20/2001 | 11/8/2000 | 5/20/2001 | Conc. | Conc. |
| Other Geochemical Data, milligrams | | | | | | | | | |
| per liter | | | | | | | | | |
| Ammonia | 1.1 Nov./1.3 Apr.6 | 891 | 627 | 734 | 678 | - 381 | 393 | 891 | 381 |
| Bicarbonate Alkalinity | | 4000 | 2800 | 3150 | 3100 | 1340 | 1550 | 4000 | 1340 |
| Carbonate Alkalinity | | 2 U | 2 U | 2 U | 2 U | 2 U | 2 U | 2 U | 2 U |
| Chloride | | 33.9 | 17.3 | 29.9 | 20.6 | 17.5 | 20.3 | 33.9 | 17.3 |
| Nitrate Nitrogen | | 2.35 | 0.545 | 0.746 | 0.05 U | 2.84 | 1.74 | 2.84 | 0.05 U |
| Soluble Organic Carbon | | 97.875 | NA | 81.925 | NA | 31.025 | NA | 97.875 | 31.025 |
| Sulfate | + | 241 | 242 | 157 | 150 | 595 | 632 | 632 | 150 |
| Total Alkalinity | | 4000 | 2800 | 3150 | 3100 | 1340 | 1550 | 4000 | 1340 |
| Total Dissolved Solids | | 1060 | NA | 1030 | NA | 855 | NA | 1060 | 855 |
| Total Hardness | - | 1100 | NA | 800 | NA | 608 | NA | 1100 | 608 |
| Total Kjeldahl Nitrogen | •- | 836 | 602 | 721 | 667 | 380 | 392 | 836 | 380 |
| Total Organic Carbon | | 100.675 | 55.525 | 81.425 | 64.875 | NA | 38.425 | 100.675 | 38.425 |
| Total Sulfide | 2 | 9.00 | 5.9 | 3.70 | 5.2 | 1 U | 2 U | 9 | 1 U |
| Turbidity, NTU | | NA | 120 | NA | 137 | NA | 4.38 | 137 | 4.38 |
| Field Measured Parameters ⁷ | | | | | | | | | |
| Conductivity (uS/cm) | | >1990 | >1990 | >1990 | >1990 | >1990 | >1990 | >1990 | >1990 |
| Dissolved Oxygen (ppm) | | 7.11 | NA | 8.48 | NA | 8.53 | NA | 8.53 | 7.11 |
| Oxidation Reduction Potential (mV) | | <-50 and >1050 | <-50 and >1050 | <-50 and >1050 | <-50 and >1050 | 75 | -40 | >1050 | <-50 |
| pH (pH units) | | 7.92 | 7.88 | 8.21 | 7.9 | 8.25 | 8.2 | 8.25 | 7.88 |
| Temperature (°C) | •• | 11.1 | 12.8 | 14.3 | 20 | 14.3 | 18.3 | 20 | 11.1 |
| urbidity (NTU) | | 212 | NA | 110 | NA | 5.8 | NA | 212 | 5.8 |

Notes:

1. Sample locations provided on Plate 1.

- 2. Data qualifications reflect 100% data validation performed by Data Validation Services.
- Surface water criteria for Class A. A-S, AA, AA-S, B, C fresh water fish propogation as provided in Division of Water Technical and Operational Series (1.1.1), Ambient Water Quality Standards and Guidance Values and Groundwater Effluent Limitations, October 22, 1993, reissued June 1998.
 Values are guidance values.
- 4. pH dependent criteria; pH = 8.1 was used to calculate Pentachlorophenol guidance value.

5. Samples collected for soluble metals analysis were field filtered.

- 6. Total Alamonia calculated with the (T) or (TS) Specifications (most conservative) using an average pH of 8.1 (Nov) and 8.0 (Apr) and average temp of 13.2 °C (Nov) and 17.0°C (Apr).
- 7. The YSI 600XL was used in the November and May sampling events for temperature. pH, specific electrical conductance, dissolved oxygen, and redox potential measurements.

Ferrous iron was field measured with the HACH18-R field kit (for QC, 10% were sent to analytical laboratory). Turbidity measurements were collected with the TURB2020 meter during the November sampling events.

NA = not analyzed -- = indicates value does not exist. NTU = Nephelometric Turbidity Unit uS/cm = microsiemens per centimeter at 25°C. ppm = parts per million mV = millivolts J = indicates an estimated value.

- U = indicates compound was not detected.
- R= indicates value was rejected by data validator.

UJ = indicates compound was not detected above the listed detection limit. However, the reported quantitation limit is approximate and may or may

not represent the actual limit of quantitiation necessary to accurately

- and precisely measure the compound in the sample.
 - indicates exceedance of surface water criteria.

(values) = laboratory reported value prior to data validation rejection.



ANALYTICAL RESULTS FOR SURFACE WATER SAMPLES FROM CATTARAUGUS CREEK

Peter Cooper Site Gowanda, New York

| | | | | Sample Le | scation, Identificat | ion, and Date Colle | ected ' | | | Ľ | 1 |
|---------------------------------------|---------------------------------------|-----------|-----------|-----------|----------------------|---------------------|---------------|---------------|------------|-----------------|---|
| | Surface Water Criteria | Creek | Nater #1 | Creek V | Nater #2 | Creek Wi | ater #3 | _ Creek | Water #4 | 1 | 1 |
| Constituent ' | | 110700101 | 050201134 | 110700100 | 050201130 | 110700098 | 050201131 | 110700097 | 050201132 | Maximum Conc | Minimum Conc. |
| | | 11/7/2000 | 5/2/2001 | 11/7/2000 | 5/2/2001 | 11/7/2000 | 5/2/2001 | 11/7/2000 | 5/2/2001 | | |
| Volatile Organic Compounds, | | | | | | | | | | 1 | |
| micrograms per liter | 1 | | 1 | | } ' | 1 | | 1 | | | |
| Acetone | | 3.5 J | NA | 10 U | NA NA | 10 U | NA | 3.2.1 | NA | 10 1 | 321 |
| Benzene | 210* | 1 10 0 | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 1 10 U | 10 U |
| Bromodichloromethane | | 1 100 | + NA | 10 1 | NA NA | 10 U | NA | 10 U | + | 1 1011 | 101 |
| Bromoform | | 1 100 | +NA | 10 U | NA | 10U | NA | 10 U | NA | 100 | + 10U |
| Romomethane | | 1 10 Ŭ | | 10 U | | 10 U | NA | 10 0 | NA NA | 100 | 101 |
| 7 Buranone (MFK) | <u> </u> | | | 10 U | NA NA | 100 | | 10.1 | | 101 | 1-101-1 |
| 4-duil ton Burd Ether | ↓ | 101 | + | 10 1 | + | 10 1 | - NA | 10 1 | - NA | 1-100- | + |
| Methyl ten-Dutyl Enter | | 100 | | 10 0 | | 100 | | 10.0 | | 100 | |
| Carbon Disulnoe | | 100 | | 10 0 | | 100 | | 100 | I INA | 10 0 | 100 |
| Carbon Tetrachionde | <u> </u> | 10 0 | 100 | 100 | 100 | 10 0 | 100 | 100 | 100 | 100 | 100 |
| Chlorobenzene | ,, | 10 0 | 100 | 10 U | 100 | 10 U | 10 U | 100 | 10 U | 10 U | 10 U |
| Chioroethane | | 10 U | NA NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Chioroform | <u> </u> | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U |
| Chloromethane | <u> </u> | 10 U | NA / | 10 U | NA | 10 U | NA 1 | 10 U | NA | 10 U | 10 U |
| 1.2-Dibromo-3-chloropropane | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 100 |
| Cyclohexane | · · · · · · · · · · · · · · · · · · · | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Dibromochloromethane | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 100 | 10 U |
| 1.2-Dibromoethane | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 100 |
| 1.2-Dichlorobenzene | 5 | 100 | 10 U | 10 U | 100 | 10 U | 100 | 10 U | 10 U | 100 | 10 U |
| 1 4-Dichlorobenzenc | 5 | 10 U | 10 U | 10 U | 10 U | 10 U | 100 | 10 U | 10 U | 100 | 10 0 |
| 1.3.Dichlombenzeng | 5 | 1 100 | NA | 10 U | NA NA | 10 U | NA | 10 11 | | 101 | 101 |
| Dichlandifugmmethane | <u> </u> | 10 1 | NA | 101 | NA | 101 | NA | 10 U | - NA | 100 | 1011 |
| 1.1 Dichlemethane | L | 101 | - NA | 10 1 | NA NA | 1011 | | 101 | - NA | 1011 | 1011 |
| 1.1-Dichloroethane | | 100 | | 10 0 | | | NA | 1010 | - NA | | 100 |
| 1.2-Dichioroethane | <u> </u> | 100 | | 100 | | 100 | | 100 | NA | 100 | 100 |
| 1,1-Dichloroethene | I | 10 0 | | 10 0 | | 10 0 | 1 <u>nn</u> | 10 0 | NA | 100 | 100 |
| trans-1.2-Dichloroethene | ļ | 100 | | 10 0 | NA P | 10 0 | | 100 | NA NA | 100 | 100 |
| cis-1,2-Dichloroethene | | 2.7) | NA | 10 U | NA | 10 U | NA | 10 0 | NA | 100 | 2.71 |
| 1.2-Dichloropropane | <u> </u> | 10 U | NA NA | 10 U | NA | 10 U | NA I | 10 U | NA | 100 | 10 U |
| trans-1.3-Dichloropropene | · · · · · · · · · · · · · · · · · · · | 10 U | | 10 U | NA | 10 U | | 10 U | NA | 100 | 10 U |
| cis-1,3-Dichloropropene | | 10 U | NA | 10 U | NA | 10 U | | 10 U | NA | 10 U | 10 U |
| Ethylbenzene | 17* | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U |
| 2-Hexanone | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Isopropylbenzene | 2.6* | 100 | NA | 10 U | NA | 10 U | NA NA | 10 U | NA NA | 10 U | 10 U |
| Methyl Acetate | | 10 U | NA | 10 U | NA | 10 U | I NA | 10 U | NA NA | 10 U | 10 U |
| Methylovelabezane | <u> </u> | 101 | NA NA | 10.11 | NA NA | 101 | - NA | 10 1 | - NA | 100 | 101 |
| Melliya yenia Aur | L | + | | 1011 | NA NA | 10 1 | | 1011 | NA NA | 1 100 | 101 |
| Methylene Cibonae | | 100 | | 100 | NA NA | 10 0 | | 100 | NA | 100 | 100 |
| 4-Methyl-2-Pentanone | | 100 | | 100 | | 10 0 | <u> </u> | 10.0 | | | 100 |
| Styrene | · | 10 0 | | 10 0 | | 10 U | | 10 U | | 400 | 10 0 |
| 1,1,2.2-Tetrachloroethane | <u> </u> | 10 U | NA J | 100 | NA NA | 10 U | NA J | 100 | NA NA | 100 | 100 |
| Tetrachloroethene | | 10 U | 100 | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 100 |
| Toluene | 100* | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U |
| 1.2.4-Trichlorobenzene | 5 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 101 |
| 1.1.1-Trichloroethane | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 1 1.2-Trichloroethane | | 10 U | NA I | 10 U 7 | NA | 10 U | NA / | 10 U | NA | 100 | 10 U |
| Trichlomethene | | 10 U | 1007 | 10 U | 100 | 10 U | 1 100 1 | 10 U | 1 10 U | iou | 10 U |
| Trichlorofluoromethate | L | 10 U | NA 7 | 10 0 | NA I | 10 U | 1 NAT | 10 U | NA NA | inu | 101 |
| 1 1.2 Tricklorn 1.7 2-Trifkinroethate | l | 101 | NA 1 | 1 1011 | NA 1 | 1011 | NA | | NA NA | 1 inu | 101 |
| 1.1.2-110000-1.2.2-11000000000 | <u>↓</u> | + | | | | 100 | + | 1011 | | 100 | 100 |
| Vinyl Chionde | | 100 | 1011 | 100 | <u></u> | 10 0 | | 100 | 101 | | - 10 0 |
| m-/p-X yiene | 3 | 100 | | 100 | | 100 | | 100 | 1 100 | | 100 |
| o-Xylene | , | 100 | <u></u> | 10 0 | 100 | 10 U | 100. | 100 | 100 | 100 | 100 |
| | 1 | 1 | 1 1 | 1 ' | l / | 1 | 1 1 | 1 1 | 1 ' | L 1 | í I |
| Semi-Volatile Organic | l | 1 | 1 1 | 1 ' | 1 1 | 1 | 1 1 | i ' | ' | 1 1 | 1 |
| Compounds, micrograms per liter | l | | L/ | L′ | LJ | I | <i>1</i> | L | L | L | í |
| Acenaphthene | 5.3* | 10 U | | 10 U | NA | 10 U | | <u>10 U /</u> | NA NA | 10 U | 100 |
| Acenaphthylene | | 100 | NA | 10 U | NA T | 10 U | NA | 10 U | NA | 10 U | 100 |
| Acetophenotic | | 100 | NA 7 | 10 U | NA | 10 U | NA | 10 U | NA | 100 | 100 |
| Anthracene | 3.8* | 10 U | 1 NA 7 | 10 U | NA T | 10.0 | 1 | 10 U | NA | 10 U | 10 U |
| A months | | 1 1011 | NA I | 1011 | NA NA | 101 | - NA | 10.11 | NA NA | 1-1011-1 | 1011 |
| All dzur | | 1 101 | | | | 100 | + | 101 | NA | | 100 |
| Benzaidenvic | | | + | | | 100 | + | | 1- <u></u> | 100 | |
| Benzo(a)anthracene | 0.03 | 10 0 | | 10 0 | 100 | 10 0 | - <u>10 U</u> | | 4 | 100 | 9.50 |
| Benzo, a)pyrene | | 10 U | 100 | 10 U | 10 U | 100 | 100 | 100 | 9.5 0 | 100 | 9.50 |
| Benzo(b)fluoranthene | | 10 U | 10 U | 10 U | <u>10 U</u> | 10 U | 10 U | <u>10 U /</u> | 9.5 U_1 | 100 1 | 9.50 |
| Benzo(g.h.i)perylene | | 10 U | NA | 10 U | NA | 10 U | NA NA | 10 U | | 100 | 10 U |
| Benzo(k)fluoranthene | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA 1 | 10 U | 100 |
| 1.1-Biphenyl | | 100 | NA 7 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 100 |
| Ruryl Renzyl Phthalate | l | 1 iou | NA T | 10 U | | 100 | + | 10 0 | NA | 100 | 100 |
| J. M. Dunulahthalata | f | + 10 T | | 1011 | | 10 11 | + | 1011 | NA NA | 1011 | 1-1011 |
| di-N-Butyiphinaizie | | 100 | + | 100 | | 10 0 | + | | | 100 | |
| Caprolactam | <u> </u> | 100 | | 10.0 | | 10 0 | 1 <u>No</u> 1 | 100 | <u> </u> | 100 | 100 |
| Carbazole | <u> </u> | 10 U | NA J | 10 U | NA | 10 U | NA I | 10 0 | NA I | 100 | 100 |
| Indeno(1,2,3-cd)pyrene | | 10 U | 100 | 10 U | 100 | 10 U | 100 | 100 | 9.5 U | 100 | 9.5 U |
| 4-Chloroaniline | · · · | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA I | 10 U | 100 |
| bis(2-chloroethoxy)methane | | 10 U | NA / | 10 U | NA | 10 U | NA | <u>10 U</u> | NA | 10 U | 10 U |
| hist2-chloroethyliethet | | 100 | NA / | 10 U | NA | 10 U | NA | 10 U | NA 7 | 10 U | 10 U |
| 2-Chloronaphthalene | <u> </u> | 100 | NA NA | 10 U | NA | 10 U | NA 1 | 10 U | NA | 100 | 100 |
| 2-Chioronberol | | 10 1 | 101 | 10.11 | 10.11 | 10.11 | 1011 | 10 11 | 9511 | 101 | 950 |

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TABLE 4-13

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ANALYTICAL RESULTS FOR SURFACE WATER SAMPLES FROM CATTARAUGUS CREEK

Peter Cooper Site Gowanda, New York

| Sample Location, Identification, and Date Collec | | | | | | | | ned' | | | |
|--|----------------------|-----------|------------|-----------|------------|-----------|-------------|-----------|------------|----------|------------|
| | | Creek | Vater #1 | Creek 1 | Water #2 | Creek Wa | uer #3 | Creek | Water #4 | 1 | |
| | Surface Water | 110700101 | 050201134 | 110700100 | 050201130 | 110700098 | 050201131 | 110700097 | 050201132 | Maximum | Minimum |
| Constituent ¹ | Criteria ' | 11/7/2000 | 5/2/2001 | 11/7/2000 | 5/2/2001 | 11/7/2000 | 5/2/2001 | 11/7/2000 | 5/2/2001 | Conc | Conc. |
| 2.2-oxybis(1-chloropropane) | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 101 |
| Chrysene | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 101 | 1011 |
| Dibenzo(a,h)anthracene | | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 9.5 U | 10 U | 9.5 U |
| Dibenzofuran | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 101 |
| 3 3-Dichlorobenzidine | - | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 101 |
| 2.4-Dichlorophenol | | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 9.5 U | 101 | 9.5 U |
| Diethylphthalate | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 0 | 100 |
| Dimethyl Phthalate | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2,4-Dimethylphenol | | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 9.5 U | 10 U | 9.5 U |
| 2,4-Dinitrophenol | | 25 U | 26 U | 25 U | 25 U | 25 U | 25 U | 25 U | 24 U | 26 U | 24 U |
| 2.4-Dinitrotohuene | - | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2.6-Dinitrotoluene | | 10 U | NA | 10 U | NA NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| bis(2-Ethylhexyl)phthalate | - | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Fluoranthene | - | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA NA | 10 Ü | 10 U |
| Fluorenc | 0.54* | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Hexachlorobenzene | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | iou |
| Hexachlorobutadiene | 1 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Hexachlorocyclopentadiene | 0.45 | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Hexachloroethane | | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| Lophorone | <u> </u> | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | 10 U |
| 2-Methylnaphthaiene | 4.7* | 10 U | NA | 10 U | NA | 10 U | NA | 10 U | NA NA | 10 U | _ 10 U |
| 4.6-Dinitro-2-Methylphenol | - | 25 U | 26 U | 25 U | 25 U | 25 U | 25 U | 25 U | 24 U | 26 U | 24 U |
| 4-Chloro-3-Methylphenol | | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 9.5 U | 10 U | 9.5 U |
| 2-Methylphenol | | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 9.5 U | 10 U | 9.5 U |
| 4-Methylphenol | | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 10 U | 9.5 U | 100 | 9.5 U |
| Naphthalene | 13* | 10 U | NA | 10 U | NA | 10 U | NA | 10 0 | NA | 10 U | 10 U |
| 2-Nitroaniline | | 25 U | NA NA | 25 0 | NA NA | 25 0 | | 25.0 | NA NA | 250 | 250 |
| 3-Nitroaniine | | 25 0 | NA | 25.0 | NA | 25 0 | NA | 25.0 | NA | 25 0 | 25 0 |
| 4-Nitroaniine | | 25 U | NA | 25 0 | NA | 25 0 | NA | 250 | <u>NA</u> | 250 | 250 |
| Nirobenzene | | 10 U | 10.11 | 10 0 | | 10 0 | | 100 | | 100 | 0.61 |
| 2-Nitrophenol | | 25.11 | 26 11 | 25.11 | 25.1/ | 75.11 | 25.11 | 26.11 | 9.5 0 | 2611 | 9.50 |
| A Nitror eduption latering | | 10.1 | 200 | 10.11 | 250 | 250 | 250 | 10 11 | NA | 1011 | 1011 |
| din Ortyl Phthalate | | 10 0 | NA | 100 | NA NA | 100 | NA | 10 0 | NA NA | 1011 | 100 |
| Provide Fision | | 100 | 24.11 | 26.11 | | | 26.11 | 100 | 24.11 | 26.11 | 100 |
| Peniachiorophenoi | 24.7 | 25 0 | 26 0 | 25 0 | 25 0 | 25.0 | 250 | 250 | 24 0 | 26 0 | 24 0 |
| Phenaminrene | 5.0* | 10 0 | | 10.0 | | 100 | | 100 | | 100 | 100 |
| Preno: | | 10 0 | 100 | 10 U | 100 | 100 | 100 | 10 0 | 9.5 0 | 100 | 9.50 |
| 4 Bromophenyl-Phenylether | | 10 0 | NA | 10 0 | | 100 | | 10 0 | | 100 | 100 |
| 4-Chlorophenyl-Phenylether | ···· | 10 0 | NA | 10 0 | NA NA | 10 0 | | 10 0 | NA NA | 100 | 100 |
| n-Nitroso-di-n-Propylanune | | 100 | NA NA | 100 | NA NA | 100 | <u>. NA</u> | 100 | | 100 | 100 |
| 2 1 6 Trichlemanhanol | 4.0 | 10 0 | 10.11 | 10 0 | 10 | 10 0 | 10 | 100 | 0.5.11 | 100 | 0.5.11 |
| 2,4,0- Thendrophenol | | 25.11 | 26 11 | 10 0 | 10 | 10 0 | 10 | 100 | 9.30 | 2611 | 9.50 |
| 2,4,5-1161010010010010 | | 450 | 200 | 250 | 23 | 23 0 | | 250 | 24 U | 200 | 240 |
| | | | i i | | | | | | | | |
| A huminum | 0.1 | 0111 | NA | 0.1.1/ | NA | 0.1.1/ | NA | 011 | N'A | 0111 | 0111 |
| Antimonu | 0.1 | 0.10 | · NA | 0.7 0 | NA NA | 0.10 | NA NA | 0.10 | | 0.10 | 0.10 |
| Aranio | 0.16 | 0.00 0 | | 0.00 0 | 0.01.11 | 0.00 0 | 001.0 | 0.00 0 | 0.011 | 0.000 | 0.06 0 |
| A Iselia | 0.15 | 0.010 | 0.010 | 0.01 0 | 0.010 | 0.010 | 0.01 U | 0.010 | 0.01 0 | 0.010 | 0.010 |
| Barmin | | 0.0041 | | 0.0047 | NA | 0.0018 | | 0.0093 | | 0.0075 | 0.0018 |
| Berymum | 1.5 | 0.005 0 | NA | 0.005 0 | NA | 0.005 0 | NA NA | 0.005 0 | NA | 0.005.0 | 0.005 U |
| Cadmium' | 0.0035 | 0.005 U | NA | 0.005 U | NA | 0.005 U | NA NA | 0.005 U | NA | 0.005 U | 0.005 U |
| Calcium | | 57.8 | 51.8 | 59.6 | 51.9 | 58.3 | 53.4 | 59.1 | 56.6 | 59.6 | 51.8 |
| Chromum ³ | 0.1200 | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U |
| Hexavalent Chromium | 0.0110 | 0.04 U | (0.01 U) R | 0.01 U | (0.01 U) R | 0.01 U | (0.01 U) R | 0.01 U | (0.01 U) R | 0.04 U | (0.01 U) R |
| Cobah | 0.0050 | 0.05 U | NA | 0.05 U | NA | 0.05 U | NA | 0.05 U | NA | 0.05 U | 0.05 U |
| Copper ⁵ | 0.0158 | 0.02 U | NA | 0.02 U | NA | 0.02 U | NA | 0.02 U | NA | 0.02 U | 0.02 U |
| Iron | 0.3000 | 0.129 | 0.39 | 0.126 | 0.403 | 0.143 | 0.47 | 0.151 | 0.344 | 0.47 | 0.126 |
| Lead ⁵ | 0.0078 | 0.005 12 | 0.005.11 | 0.005 11 | 0.005.11 | 0.005.11 | 0.005 U | 0.005.11 | 0.005.11 | 0.005.11 | 0.00512 |
| Magnesium | 0.0010 | 10.3 | 9.25 | 10.3 | 945 | 9.88 | 9.21 | 10.8 | 9.99 | 10.8 | 9.21 |
| Manganese | | 0.0115 | 0.0161 | 0.0138 | 0.0149 | 0.0129 | 0.0216 | 0.0184 | 0.0206 | 0.0216 | 0.0115 |
| Memury | 0.0008 | 0.0003.11 | NA | 0.0003 11 | NA NA | 0.0003.11 | N.A. | 0.0003.11 | NA NA | 0.000311 | 0.0003.11 |
| without y | 0.0000 | 0.00050 | | 0.0005 0 | | 0.0005 0 | | 0.0005 0 | NA | 0.00050 | 0.000.7 0 |
| Njckel | 0.0915 | 0.04 U | NA NA | 0.04 0 | NA | 0.04 U | NA | 0.04 0 | NA | 0.04 0 | 0.04 0 |
| Polassium | | 20 | 20 | 20 | 20 | 20 | 20 | 2.0 | 20 | 20 | 20 |
| Selenium | 0.0046 | 0.005 0 | NA | 0.005 U | NA | 0.005 U | NA | 0.005 0 | NA NA | 0.005 0 | 0.005 U |
| Silver | 0.0001 | 0.01 U | NA | 0.01 U | NA NA | 0.01 0 | NA NA | 0.01 U | NA | 0.01 0 | 0.01 0 |
| Sodaum | | 13.7 | NA | 13.9 | NA | 13.4 | NA | 16.2 | NA | 16.2 | 15.4 |
| 1 hamm | 0.008 | 0.010 | NA NA | 0.01.0 | NA | 0.01 U | NA NA | 0.01 D | NA | 0.010 | 0.01 0 |
| Zinc | 0.0094 | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U |
| Vanadium | 0.0140 | 0.05 U | NA | 0.05 U | NA | 0.05 U | NA | 0.05 U | NA | 0.05 U | 0.05 U |
| Other Geochemical Data, | 1 | | | | | | | | 1 7 | | |
| milligrams per liter | | | | | | | | | | | |
| Anunonia | 0.58 Nov./0.44 Apr.6 | 0.05 U | 0.05 U | 0.05 U | 0.05 U | 0.234 | 0.306 | 0.17 | 0.442 | 0.442 | 0.05 U |
| Bicarbonate Alkalinity | - | 167 | 270 J | 166 | 133 | 164 | 135 | 169 | 140 | 270 J | 133 |
| Carbonate Alkalinity | | 2 U | 2.0 | 2 U | 2 UJ | 2 U | 20 | 2 U | 2 U | 2 UJ | 2 UJ |

1 WrojactWR2731 PRP Group Pater Couper NPLVR) reportFDNAL REPORT (November 2001 Sobujita/MTables (Final/MTable 4-13 surface + dos FDNAL

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TABLE 4-13

ANALYTICAL RESULTS FOR SURFACE WATER SAMPLES FROM CATTARAUGUS CREEK

Peter Cooper Site Gowanda, New York

| | Sample Location, Identification, and Date Collected | | | | | | | | | | |
|--|---|----------------|-----------|-----------|-----------|-----------|-----------|----------------|-----------|---------|---------|
| | | Creek Water #1 | | Creek V | Vater #2 | Creek Wa | ter #3 | Creek Water #4 | | 1 | |
| | Surface Water | 110700101 | 050201134 | 110700100 | 050201130 | 110700098 | 050201131 | 110700097 | 050201132 | Maximum | Minimum |
| Constituent 4 | Criteria ' | 11/7/2000 | 5/2/2001 | 11/7/2000 | 5/2/2001 | 11/7/2000 | 5/2/2001 | 11/7/2000 | 5/2/2001 | Conc | Conc. |
| Chloride | | 24 | 26.4 | 22.9 | 27 | 23.4 | 27.1 | 28.7 | 46.9 | 46.9 | 22.9 |
| Ferrous Iron | - | NA | NA | NA | 0.1 U | NA | NA | NA | NA | 0.1 U | 0.1 U |
| Nitrate Nitrogen | | 1.78 | 1.07 | 1.81 | 1.11 | 1.81 | 1.07 | 1.9 | 1.12 | 1.9 | 1.07 |
| Sulfate | | 28.8 | 24.8 | 27.6 | 25.9 | 27.5 | 24.9 | 28.5 | 28 | 28.8 | 24.8 |
| Total Alkalinity | | 167 | 270 J | 166 | 133 | 164 | 135 | 169 | 140 | 270 J | 133 |
| Total Dissolved Solids | | 254 | 216 | 250 | 221 | 249 | 216 | 255 | 264 | 264 | 216 |
| Total Hardness | | 191 | 166 | 198 | 164 | 195 | 161 | 200 | 175 | 200 | 161 |
| Total Kjeklahl Nitrogen | - | 0.308 | 0.345 | 0.412 | 0.2 U | 0.417 | 0.445 | 0.344 | 0.648 | 0.648 | 0.2 |
| Total Organic Carbon | | 1.975 U | 1.665 | 1.8875 | 1.6525 | 2.135 | 1.675 | 1.9875 | 1.7225 | 2.135 | 1.6525 |
| Total Sulfide | 2 | 10 | 2 UJ | 10 | 2 | 10 | 2 UJ | 10 | 2 UJ | 2 UJ | 10 |
| Total Suspended Solids | | 1.3 | 6.6 | 1.6 | 7.1 | 1.3 J | 8.2 | 1.9 | 4.9 | 8.2 | 1.3 J |
| Field Measured Parameters ⁷ | | | | | | | | | | | |
| Conductivity (uS/cm) | - | 440 | 350 | 390 | 340 | 320 | 340 | 340 | 390 | 440 | 320 |
| Dissolved Oxygen (ppm) | •• | NA | NA | 9.8 | NA | 8.65 | NA | 13.6 | NA | 13.6 | 8.65 |
| Ferrous Iron (mg/l) | - | NA | NA | NA | 0 | NA | 0 | NĂ | 0 | 0 | 0 |
| Oxidation Reduction Potential (mV) | | 30 | -40 | 35 | -60 | 35 | -60 | -5 | -45 | 35 | -60 |
| pH (pH units) | | 8.52 | 8.5 | 8.3 | 8.42 | 8.37 | 8.4 | 8.36 | 8.5 | 8.52 | 8.3 |
| Temperature (°C) | | 16.5 | 14.4 | 7.9 | 14.4 | 7.8 | 14.4 | 5.3 | 14.4 | 16.5 | 5.3 |
| Turbidity (NTU) | | 2.43 | NA | NA | NA | 5.18 | NA | 4.14 | NA | 5.18 | 2.43 |

Notes

1. Sample tocations provided on Plate 1

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

- 3 Surface water criteria for Class A. A S. AA. AA-S. B. C fresh water fish propogation as provided in Division of Water Technical and Operational Series (1.1.1); Ausbient Water Quality Standards and Guidance Values and Groundwater Effluent Limitations, October 22, 1993, reissand June 1998.
- * Values are guidance values.
- 4. pH dependent criteria: pH = 8.3 was used to calculate Pentachlorophenol guidance value.

- 5. Hardness dependent criteria: Hardness value of 181 prus was used. 6. Total Animonia calculated with the (T) or (TS) Specifications (most conservative) using an average pH of 8.4 (Nov) and 8.5 (Apr) and average temp of 9.4 °C (Nov) and 14.4 °C (Apr).
- 7. The YSI 600XL was used in the November and May sampling events for temperature. pH. specific electrical conductance, dissolved oxygen.and redox potential measurements.

Ferrous iron was field measured with the HACH18-R field kit (for QC. 10% were seat to analytical laboratory) Turbidity successreteness were collected with the TURB2020 meser during the November sampling events.

EVERYTHERTE PRP Group Peter Cooper NPL/RE-reportFD-AL REPORT (November 2018 Schmithel/Tables (Final/Eable 4-13 writer water FD-AL

NA = not analyzed

(values) = laboratory reported value prior to data validation ang/l = milligrams per liter NTU = Nephelometric Tarbidity Unit

#S/cm = microsiemess per centimeter at 25°C. - = indicates guidance value does not exist.

ppen = parts per million

ns∨ = millivohs

U = indicates compound was not detected. R= indicates value was rejected by data validator UI = indicates compound was not detected above the listed detection limit.

J = indicates an estimated value

- However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitiation necessary to accurately
- and precisely measure the compound in the sample.

- a indicates value does not exist.

indicates exceedance of surface water criteria

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TABLE 4-14

ANALYTICAL RESULTS FOR WETLAND SEDIMENT SAMPLES

Peter Cooper Site Gowanda, New York

| | | Sediment/So | il Criteria ³ | | Sample Location, Identification, and Date Collected | | | | | | في والتقديم مريكم | | | | | |
|---------------------------------|----------|-------------|--------------------------|----------|---|------------|------------|------------|------------|------------|-------------------|------------|------------|------------|-----------|-----------|
| | NYSDEC | | | | WSS-1 | WSS-2 | WSS-3 | WSS-4 | WSS-5 | WSS-6 | WSS-7 | WSS-8 | WSS-9 | WSS-10 | • | |
| | Sediment | | | Region 9 | 101000047 | 101000048 | 101000049 | 101000050 | 101000051 | 101000052 | 101000054 | 101000055 | 101000056 | 10100057 | Maximum | Minimum |
| Constituent 4 | Criteria | TA | GM | PRGs | 10/10/2000 | 10/10/2000 | 10/10/2000 | 10/10/2000 | 10/10/2000 | 10/10/2000 | 10/10/2000 | 10/10/2000 | 10/10/2000 | 10/10/2000 | Conc. | Conc. |
| Volatile Organic Compounds, | | | | | | | | | | | | | | | | |
| milligrams per kilogram | | | | | | | | | | | | | | | | |
| Benzene | | 0. | 06 | 1.5 | 0.0065 J | 0.0085 J | 0.0037 J | 0.0058 J | 0.005 J | 0.004 J | 0.0068 J | 0.0082 J | 0.0035 J | 0.0026 J | 0.0085 J | 0.0026 J |
| Chlorobenzene | | 1 | .7 | 54 | 0.0120 U | 0.023 UJ | 0.0063 UJ | 0.013 UJ | 0.014 UJ | 0.018 UJ | 0.012 U | 0.014 UJ | 0.023 UJ | 0.017 UJ | 0.023 UJ | 0.0063 UJ |
| 1,2-dichlorobenzene | | 7 | .9 | 370 | 0.0120 UJ | 0.023 UJ | 0.0063 UJ | 0.013 UJ | 0.014 UJ | 0.018 UJ | 0.012 U | 0.014 UJ | 0.023 UJ | 0.017 UJ | 0.023 UJ | 0.0063 UJ |
| 1,4-dichlorobenzene | | 8 | .5 | 8.1 | 0.0120 UJ | 0.023 UJ | 0.0063 UJ | 0.013 UJ | 0.014 UJ | 0.018 UJ | 0.012 U | 0.014 UJ | 0.023 UJ | 0.017 UJ | 0.023 UJ | 0.0063 UJ |
| Ethylbenzene | | 5 | .5 | 230 | 0.0015 J | 0.0034 J | 0.0009 J | 0.013 UJ | 0.014 UJ | 0.018 UJ | 0.0014 J | 0.0021 J | 0.023 UJ | 0.0033 J | _0.023 UJ | 0.00094 J |
| m/p-Xylene | | | .2 | | 0.0082 J | 0.015 J | 0.0044 J | 0.0058 J | 0.006 J | 0.0053 J | 0.0083 J | 0.011 J | 0.023 UJ | 0.017 UJ | 0.023 UJ | 0.0044 J |
| o-Xylene | | 1 | .2 | | 0.0027 J | 0.0044 J | 0.0013 J | 0.0017 J | 0.0019 J | 0.018 U | 0.0023 J | 0.0033 J | 0.023 UJ | 0.017 UJ | 0.023 UJ | 0.0013 J |
| Toluene | | 1 | .5 | 520 | 0.0120 | 0.018 J | 0.0066 J | 0.011 J | 0.0082 J | 0.0082 J | 0.011 J | 0.015 J | 0.016 J | 0.0041 J | 0.018 UI | 0.0041 J |
| | | Fastern | Rec Soil | | | | | | | | | | | | | |
| Metals, milligrams per kilogram | | USA | Objective | | | | | | | | | | | | | |
| Arsenic | 6 | 3-12 | 7.5 or SB | 2.7 | 7.4 | 16.3 | 8.7 | 8.5 | 9.4 | 10.7 | 5.2 | 5.6 | 9.9 | 8.6 | 16.3 | 5.2 |
| Chromium | 26 | 1.5-40** | 10 or SB | 450 | 6.5 | 44.9 | 11.8 | 28.4 | 30.6 | 31.2 | 8.9 | 13.7 | 17.2 | 55.9 | 55.3 | 6.5 |
| Hexavalent Chromium | | | | 64 | 5.07 U | 7.12 U | 5.35 U | 5.29 U | 5.43 U | 5.87 U | 4.68 U | 5.55 U | 6.34 U | 5.81 U | 7.12 U | 4.68 U |
| Zinc | 120 | 9-50 | 0.2 | 100,000 | 45.7 | 227 | 69.8 | 80.5 | 74.9 | 92.5 | 58.8 | 65.6 | 290 | 110 | 290 | 45.7 |
| | | | | | | | | | | | | | | | | |
| Other | | | L | | | | | | | [| | | | | | L |
| Percent Solids, % | | | | | 78.9 | 56.2 | 74.8 | 75.6 | 73.6 | 68.2 | 85.5 | 72.1 | 63.1 | 68.8 | 85.5 | 56.2 |
| pH | | | | | 8.17 | 7.56 | 7.68 | 7.76 | 7.48 | 7.74 | 7.91 | 7.47 | 7.30 | 6.92 | 8.17 | 6.92 |
| Total Organic Carbon, % | | | | | 0.29 | 3.4 | 1.50 | 1.70 | 1.90 | 2.70 | 0.290 | 1.50 | 3.80 | 4.40 | 4.4 | 0.29 |

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

3. Soil criteria from NYSDEC Division of Technical and Administrative Guidance Memorandum #4046 (TAGM) and U.S.EPA, Region 9 Preliminary Remediation Golas (PRGs) for Industrial Soil (October 2002).

Sediment criteria from NYSDEC Technical Guidance for Screening Contaminated Sediments, Division of Fish and Wildlife.

** A New York State Background value

 $\mathbf{J}=\mathbf{indicates}$ a laboratory estimated value or estimated as a result of data validation.

U = indicates compound was not detected at or above the listed detection limit. UI = indicates compound was not detected above the listed detection limit.

(i) = indicates compound was not detected above the listed detection limit.

However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitation necessary to accurately and precisely

measure the compound in the sample.

SB ≈ Site Background -- = indicates value does not exist.

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indicates exceedance of upper range of US Eastern Soils.



Page 1 of 3

TABLE 4-15

ANALYTICAL RESULTS FOR CATTARAUGUS CREEK SEDIMENTS

Peter Cooper Site Gowanda, New York

| | | Sample Location, Identification, and Date Collected | | | | | |
|---------------------------------------|-----------------------|---|---------------|---------------|---------------|----------|----------|
| | | Creek Sed. #1 | Creek Sed. #2 | Creek Sed. #3 | Creek Sed. #4 | | |
| | Sediment | 110700096 | 110700095 | 110700093 | 110700092 | Maximum | Minimum |
| Constituents ² | Criteria ³ | 11/7/2000 | 11/7/2000 | 11/7/2000 | 11/7/2000 | Сопс. | Сопс. |
| Volatile Organic Compounds, | | | | | | | |
| milligrams per kilogram | | | | | | | |
| Acetone | 1 | 0.024 | 0.078 | 0.019 | 0.022 | 0.078 | 0.019 |
| Benzene | | 0.017 U | 0.0025 J | 0.0015 J | 0.0014 J | 0.017 U | 0.0014 |
| Bromodichloromethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| Bromoform | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| Bromomethane | 1 | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| 2-Butanone (MEK) | | 0.017 U | 0.0095 J | 0.011 U | 0.011 U | 0.017 U | 0.0095 J |
| Methyl tert-Butyl Ether | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| Carbon Disulfide | | 0.01 J | 0.025 | 0.019 | 0.02 | 0.025 | 0.01 |
| Carbon Tetrachloride | I | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| Chlorobenzene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| Chloroethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| Chloroform | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| Chloromethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| 1.2-Dibromo-3-Chloropropane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| Cyclohexane | | 0.017 U | 0.0045 J | 0.0022 J | 0.0022 J | 0.017 U | 0.0022 |
| Dibromochloromethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| 1,2-Dibromoethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| 1,2-Dichlorobenzene | | 0 017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| 1.4-Dichlorobenzene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| 1,3-Dichlorobenzene | [| 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| Dichlorodifluoromethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| 1.1-Dichloroethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| 1,2-Dichloroethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| 1,1-Dichloroethene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| trans-1.2-Dichloroethene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.911 |
| cis-1.2-Dichloroethene | | 0.017 U | 0.0035 J | 0.011 U | 0.011 U | 0.017 U | 0.0035 |
| 1.2-Dichloropropane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| trans-1.3-Dichloropropene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| cis-1,3-Dichloropropene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| Ethylbenzene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| 2-Hexanone | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| lsopropylbenzene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| Methyl Acetate | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| Methylcyclohexane | | 0.017 U | 0.0072 J | 0.0033 J | 0.0034 J | 0.017 U | 0.0033 |
| Methylene Chloride | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| 4-Methyl-2-Pentanone | | 0.017 U | 0.0025 J | 0.011 U | 0.011 U | 0.017 U | 0.0025 |
| Styrene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| 1.1.2.2-Tetrachloroethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| Tetrachloroethene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| Toluene | | 0.0059 J | 0.0068 J | 0.0045 J | 0.0041 J | 0.0068 J | 0.0041 |
| 1.2.4-Trichlorobenzene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| 1.1.1-Trichloroethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| 1.1.2-Trichloroethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| Trichloroethene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| Trichlorofluoromethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| 1.1.2-Trichloro-1.2.2-Trifluoroethane | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 U |
| Vinyl Chloride | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |
| m-/n-Xylene | | 0.017 U | 0.0027 1 | 0.0015 J | 0.00151 | 0.017 U | 0.0015 |
| o-Xylene | | 0.017 U | 0.012 U | 0.011 U | 0.011 U | 0.017 U | 0.011 |



TABLE 4-15

ANALYTICAL RESULTS FOR CATTARAUGUS CREEK SEDIMENTS

Peter Cooper Site Gowanda, New York

| | | Sample | Γ | | | | |
|------------------------------------|-----------------------|---------------|---------------|---------------|---------------|---------|---------|
| 1 | | Creek Sed. #1 | Creek Sed. #2 | Creek Sed. #3 | Creek Sed. #4 | 1 | |
| 1 | Sediment | 110700096 | 110700095 | 110700093 | 110700092 | Maximum | Minimum |
| Constituents ² | Criteria ³ | 11/7/2000 | 11/7/2000 | 11/7/2000 | 11/7/2000 | Conc | Conc |
| | | | | | | I | conc. |
| Semi-Volatile Organic Constituents | | | | | | | |
| milligrams per kilogram | | | | | | | |
| A cepaphthene | | 0411 | 0.42 11 | 0411 | 0.41.11 | 0.42 | 0.4 |
| A cenanbibylene | | 0.4 U | 0.42 0 | 0.4 U | 0.41 U | 0.42 | 0.4 |
| Acetophenone | | 04 U | 0.42 U | 0.4 0 | 0.41 U | 0.42 | 0.4 |
| Anthracene | | 0.4 0 | 0.42.0 | 0.40 | 0.41 U | 0.42 | 0.4 |
| Atrazine | | 0.4 U | 0.42 0 | 0.4 0 | 0.41 U | 0.42 | 0.4 |
| Renzoldabyde | | 0.4 U | 0.42.0 | 0.4 U | 0.41 U | 0.42 | 0.4 |
| Benzo(a)anthracene | | 0.4 U | 0.42 05 | 0.4 05 | 0.41 U | 0.42 | 0.4 |
| Benzo(a)purene | | 0.4 U | 0.42.0 | 0.4 U | 0.41 U | 0.42 | 0.4 |
| Benzo(b)fluoranthene | | 0.4 U | 0.42.0 | 041 | 0.41 U | 0.42 | 0.4 |
| Benzo(g h i)pervlene | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| Benzo(k)fluoranthene | | 0.4 U | 0.42.0 | 041 | 0.41 U | 0.42 | 0.4 |
| 1 1-Binbery | | 04U | 0.42 U | 0.4 U | 041 U | 0.42 | 0.4 |
| Butyl Benzyl Phthalate | | 041 | 0.42 U | 0411 | 0.41 U | 0.42 | 0.4 |
| di-N-Butylphthalate | | 0.4 U | 0.42.11 | 04 U | 0.41 U | 0.42 | 0.4 |
| Caprolactam | | 0411 | 0.42 11 | 0.4 11 | 0.41 11 | 0.42 | 0.4 |
| Carthazole | | 041 | 0.42 U | 041 | 041 U | 0.4? | 0.4 |
| Indepo(1.2.3-cd)pyrene | | 04U | 0.42 U | 04 U | 0.41 U | 0.42 | 0.4 |
| A-Chloroaniline | | 0.4 [] | 0.42 U | 0.4 1 | 0.41 U | 0.42 | 0.4 |
| his(2-chloroethoxy)methane | | 041 | 0.42 11 | 0.4 0 | 0.41 U | 0.42 | 0.4 |
| bis(2-chloroethyl)ether | | 041 | 0.42 U | 041 | 041 U | 0.42 | 0.4 |
| 2-Chloronaphthalene | | 0.4 U | 0.42 11 | 0411 | 0.41 U | 0.42 | 0.4 |
| 2-Chlorophenol | | 0.4 U | 0.42 U | 041 | 0.41 U | 0.42 | 0.4 |
| 2 2-oxybis(1-chloropronane) | | 041 | 0.42 0 | 041 | 0.41 U | 0.42 | 0.4 |
| Chrysene | | 041 | 0.42.11 | 041 | 0.41 U | 0.42 | 0.4 |
| Dibenzo(a h)anthracene | | 04U | 0.42 U | 041 | 0.41 U | 0.42 | 0.4 |
| Dibenzofuran | | 04 U | 0.42 U | 04 U | 0.41 U | 0.42 | 0.4 |
| 3 3-Dichlorgbenzidine | | 0411 | 0.42 U | 041 | 0.41 U | 0.42 | 0.4 |
| 2 4-Dichlorophenol | | 0.4 0 | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| Diethylphthalate | | 0.4 U | 0.42 U | 041 | 0.41 U | 0.42 | 0.4 |
| Dimethyl Phthalate | | 041 | 0.42 U | 0.4 0 | 0.41 U | 0.42 | 0.4 |
| 2 4-Dimethylphenol | | 0.4 U | 0.42.0 | 0.4 U | 0.41 U | 0.42 | 0.4 |
| 2 4-Dinitrophenol | | 111 | 1 11 | 1 11 | 1 11 | 1 | 1 |
| 2 4-Dinitrotaluene | | 041 | 0.42 U | 0411 | 0.41 11 | 0.42 | 0.4 |
| 2.6 Disitrotoluene | | 0.4 U | 0.42 U | 0.40 | 0.41 U | 0.42 | 0.4 |
| bis(2 Ethylbaxid)abthalate | | 0.4 0 | 0.42 U | 0.4 0 | 0.41 U | 0.42 | 0.4 |
| Elucemethene | | 0.4 U | 0.42 U | 0.4 0 | 0.41 U | 0.42 | 0.4 |
| Fluorantitiene | | 0.40 | 0.42 U | 0.4 0 | 0.41 U | 0.42 | 0.4 |
| Variable abore | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| | | 0.4 U | 0.42 U | 0.4 0 | 0.41 0 | 0.42 | 0.4 |
| Hexachiorodulaciene | | 0.4 0 | 0.42 0 | 0.4 U | 0.41 U | 0.42 | 0.4 |
| Hexachiorocyclopentadiene | | 0.4 01 | 0.42 0J | 0.4 UJ | 0.41 UJ | 0.42 | 0.4 |
| Hexachioroethane | | 0.4 0 | 0.42 0 | 0.4 U | 0.41 U | 0.42 | 0.4 |
| Isophorone | | 0.4 U | 0.42 U | 0.4 0 | 0.41 U | 0.42 | 0.4 |
| 2-Methylnaphthalene | | U.4 U | U.42 U | 0.4 U | 0.41 0 | 0.42 | 0.4 |
| 4,6-Dinitro-2-Methylphenol | | 10 | 10 | 10 | 10 | 1 | |
| 4-Chloro-3-Methylphenol | | 0.4 U | 0.42 U | 0.4 U | <u>U.41 U</u> | 0.42 | 0.4 |
| 2-Methylphenol | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| 4-Methylphenol | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| Naphthalene | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| 2-Nitroaniline | | 10 | <u> </u> | 10 | 1 U | 1 | 1 |
| 3-Nitroaniline | | 10 | 10 | 10 | <u>1</u> U | 1 | 1 |
| 4-Nitroaniline | | 10 | 10 | 10 | 1 U | 1 | 1 |





TABLE 4-15

ANALYTICAL RESULTS FOR CATTARAUGUS CREEK SEDIMENTS

Peter Cooper Site Gowanda, New York

| | T | Sample | | | | | |
|--|-----------------------|---------------|---------------|---------------|---------------|---------|-------------|
| | | Creek Sed. #1 | Creek Sed. #2 | Creek Sed. #3 | Creek Sed. #4 | 1 | |
| | Sediment | 110700096 | 110700095 | 110700093 | 110700092 | Maximum | Minimum |
| Constituents ² | Criteria ³ | 11/7/2000 | 11/7/2000 | 11/7/2000 | 11/7/2000 | Conc. | Conc. |
| Nitrobenzene | Ť | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| 2-Nitrophenol | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| 4-Nitrophenol | | 1 U | 1 U | 1 U | 1 U | 1 | 1 |
| n-Nitrosodiphenylamine | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| di-n-Octyl Phthalate | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| Pentachlorophenol | | 10 | 10 | 10 | 10 | 1 | 1 |
| Phenanthrene | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| Phenol | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| 4-Bromophenyl-Phenylether | T | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| 4-Chlorophenyl-Phenylether | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| n-Nitroso-di-n-Propylamine | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| Рутепе | | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| 2,4,6-Trichlorophenol | 1 | 0.4 U | 0.42 U | 0.4 U | 0.41 U | 0.42 | 0.4 |
| 2,4,5-Trichlorophenol | | 1 U | 1 U | 1 U | 1 U | 1 | 1 |
| Metals, milligrams per kilogram | | | | | | | |
| Aluminum | 1 | 4820 | 4960 | 5730 | 6160 | 6160 | 4820 |
| Antimony | 1 | 6.9 UJ | 7.5 UJ | 7 UJ | 7.04 UJ | 7.5 | 6.9 |
| Arsenic | 6 | 7.2 J | 6.7 J | 7.1 J | 9.6 J | 9.6 | 6.7 |
| Barium | + | 31.5 | 36.1 | 38.6 | 41.4 | 41.4 | 31.5 |
| Beryllium | + | 0.57 U | 0.63 U | 0.58 U | 0.59 U | 0.63 | 0.57 |
| Cadmium | + | 0.57 U | 0.63 U | 0.58 U | 0.59 U | 0.63 | 0.57 |
| Calcium | + | 7490 | 10500 | 11700 | 5080 | 11700 | 5080 |
| Chromium | 26 | 6.3 | 6.5 | 7.1 | 8.6 | 8.6 | 6.3 |
| Cobalt | 1 | 5.7 U | 6.25 U | 6.7 | 7.5 | 7.5 | 5.7 |
| Copper | 16 | 13.7 | 11.3 | 13.9 | 14.8 | 14.8 | 11.3 |
| Hexavalent Chromium | 1 | 4.8 U | 5.05 U | 4.85 U | 4.93 U | 5.05 | 4.8 |
| lron | 20000 | 14400 | 18100 | 16900 | 18400 | 18400 | 14400 |
| Lead | 31 | 7.9 | 9.2 | 8.8 | 9.8 | 9.8 | 7.9 |
| Magnesium | + | 3290 | 3240 | 3160 | 3350 | 3350 | 3160 |
| Manganese | 460 | 250 | 356 | 401 | 246 | 401 | 246 |
| Mercury | 1 | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 | 0.06 |
| Nickel | 16 | 12.6 | 13.6 | 15.5 | 18.2 | 18.2 | 12.6 |
| Potassium | 1 | 525 | 591 | 617 | 786 | 786 | 525 |
| Selenium | 4+ | 1.1 | 0.71 | 0.58 U | 0.59 U | 1.1 | 0.58 |
| Silver | ++ | 1.1 UJ | 1.3 UJ | 1.17 UJ | 1.2 UJ | 1.3 | 1.1 |
| Sodium | ++ | 333 | 226 | 240 | 201 | 333 | 201 |
| Thallium | · 1 / | 1.1 U | 1.3 U | 12U | 1 17 U | 1.3 | 1.1 |
| Vanadium | 4 | 10.9 | 12.3 | 172 | 13.8 | 13.8 | 10.9 |
| 7 inc | 120 | 39.2 | 40.2 | 47.1 | 52.8 | 57.8 | 39.7 |
| Others | | 37.0 | | | J <u></u> | | |
| Descent Solids On | 4 | 833 | 79.7 | 875 | 91.7 | 83.3 | 79.2 |
| | -{ <i>!</i> | 96 | 17.2 | 0.10 | 01.2 | 96 | 9.19 |
| pH Totological de la companya de la comp | - / | 8.0 | 8.2 | 8.21 | 0.18 | 8.0 | ō.10 0.1 |
| Total Organic Carbon, % | | 0.10 | 0.10 | 0.10 | | 0.1 | 0.1 |

Notes:

1. Sample locations provided on Plate 1.

2. Data qualifications reflect 100% data validation performed by Data Validation Services.

3. Guidance values from NYSDEC Technical Guidance for Screening Contaminated Sediments, Division of Fish and Wildlife

J = indicates an estimated value.

U = compound was not detected at or above the listed detection limit.

UJ = indicates compound was not detected above the listed detection limit.

However, the reported quantitation limit is approximate and may or may not represent the actual limit of quantitiation necessary to accurately and precisely measure the compound in the sample.

PLATES

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P./Project/005771 PRP Group Peter Cooper NPL/Draft/Addendum May 2000/Site Location Map - Figure 1 dor



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Figure

OVERBURDEN GROUNDWATER CONTOUR MAP

INACTIVE LANDFILL AREA - NOVEMBER 2000 PETER COOPER LANDFILL SITE GOWANDA, NEW YORK

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Figure

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ECMATED

BEDROCK GROUNDWATER CONTOUR MAP INACTIVE LANDFILL AREA - APRIL 2001 PETER COOPER LANDFILL SITE

GOWANDA, NEW YORK

Project No.

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