## Supplemental Information

## Selection of the number of PMF factors

Figure $S 1$ presents the ratio between the summation of scaled residuals $(\mathrm{Q})$ and the expected value for this parameter ( $\mathrm{Q}_{\text {expected }}$ ) obtained for PMF models including 1 to 7 factors. The expected value of Q corresponds to the number of data points in the HR OA concentration matrix and represents a scenario in which the experimental errors equal the resulting model errors (Paatero, 2013). The $\mathrm{Q} / \mathrm{Q}_{\text {expected }}$ ratio under this ideal scenario would be 1, and thus, the relative value of this ratio can be used to evaluate deviations from this ideal condition. The summation of scaled residuals decreases as additional factors are incorporated into the model, and as illustrated in Figure S 1 , after a certain number of PMF factors, $\mathrm{Q} / \mathrm{Q}_{\text {expected }}$ exhibits only marginal reductions. According to Figure S1, major decreases in Q are observed when a second and third factor are considered in the model, while the retention of more than 5 factors does not lead to further evident reductions in the model residuals. Based on these observations, and considering the significant decrease in $\mathrm{Q} / \mathrm{Q}_{\text {expected }}$ after a third factor is incorporated, PMF models containing 3 to 5 factors were selected as potentially appropriate to explain the OA levels during the field campaign

The convergence of the 3 to 5-factor PMF models to a global minimum was examined by using twelve different initialization points (positions in a list of random values referred to as seeds; Paatero, 2013). The consistency of the model outputs utilizing twelve seeds was evaluated in terms of the variation in the $\mathrm{Q} / \mathrm{Q}_{\text {expected }}$ for each initialization point. As presented in Figure S2, at least two solutions with slightly differing $\mathrm{Q} / \mathrm{Q}_{\text {expected }}$ values were observed for the 4-factor model, while multiple outputs with close $\mathrm{Q} / \mathrm{Q}_{\text {expected }}$ levels resulted for the model containing 5 factors. In contrast, the 3-factor model exhibited $\mathrm{Q} / \mathrm{Q}_{\text {expected }}$ values resembling a single output likely corresponding to a global minimum.

The 3 to 5-factor PMF models were also evaluated in terms of their ability to reproduce the measured $\mathrm{PM}_{1} \mathrm{OA}^{\text {mass }}$ concentrations during the sampling period. Figure S3 compares the experimental and modeled OA concentrations based on these models. As illustrated in Figure S3, the reconstructed OA mass concentrations based on PMF models including 3, 4 and 5 factors closely resemble the measured OA levels, indicating the suitability of these modeling approaches. According to Figure S3, the inclusion of a fourth/fifth factor did not lead to improved modeling of the OA mass concentration and the additional factors were insufficient to reproduce a large spike in OA observed on 5/23/2015.

The degree of similarity between the mass spectra of the factors in the 3 to 5 -factor models was examined based on the spectral contrast angle $(\theta)$. As factor splitting may occur when additional factors are incorporated in the model, the $\theta$ between the factors provides insight on the presence of this phenomenon. As presented in Table S1, the factors in the 3-factor model closely resemble 3 of the factors in the models containing 4 and 5 factors. Similarly, the factors in the 4 -factor model were highly similar (or identical) to 4 of the factors in the model including 5 factors. The fact that the additional factors in the 4 and 5-factor models were distinct from those in the 3 and 4-factor models, respectively, discards the occurrence of factor splitting in these PMF models.

As reported in the manuscript, the mass spectra of the factors contained in the 3-factor model resembled multiple factors previously reported in the literature ( $\theta$ below $\sim 17^{\circ}$ ). In contrast, the mass signature of the additional factors in the 4 and 5 -factor models did not exhibit resemblance with any of the AMS factors reported in the UCB-AMS Spectral Database (Ulbrich et al.), preventing better evaluation of their physical meaningfulness.

Based on the larger stability of the PMF model containing 3 factors (as reflected by its repeated convergence to a single output), the marginal reduction in Q when a number of factors exceeding 3 was included in the PMF model, and the comparable OA mass concentrations reconstructed using 3 to 5 -factor models, the 3 -factor PMF model was selected as the most appropriate approximation to represent the observed OA levels during the field campaign.

Figure S1: Q/Qexpected for PMF models containing 1 and 7 factors.


Figure S2: Q/Qexpected for PMF models containing 3, 4 and 5 factors when different initialization points (seeds) are selected in PET v. 2.08D. Each seed number corresponds to a position in a list of randomly generated values (Paatero, 2013).


Figure S3: Measured and reconstructed OA mass concentration based on PMF models containing 3 (a), 4 (b) and 5 (c) factors.


Figure S4: Box-and-whiskers plots showing the diurnal profiles of (a) $\mathrm{PM}_{2.5}$, (b) nitrogen oxides, (c) ozone, (d) toluene, (e) isoprene, (f) UV radiation, (g) wind speed, and (h) ambient temperature at the Clinton Drive monitoring station in Houston, TX for 5-27 May 2015. The bottom whisker, bottom box line, top box line and top whisker indicate the $5^{\text {th }}, 25^{\text {th }}, 75^{\text {th }}$ and $95^{\text {th }}$ percentiles, respectively. Lines inside the boxes represent the hourly median and the continuous line represents the hourly mean. The hourly wind rose diagram (i) shows the frequency of occurrence (radial axis). All data were obtained from TCEQ (TCEQ, 2017). Daytime hours (7:00-18:00 LT) are highlighted yellow and correspond to the times of the daytime $\mathrm{PM}_{2.5}$ filter sample. Nighttime filter samples were collected 19:00- 6:00 LT.


Figure S5: Hourly $\mathrm{PM}_{2.5}$ concentrations $\left(\mu \mathrm{g} \mathrm{m}^{-3}\right)$ and precipitation (cm) at the Clinton Drive monitoring station in Houston, TX, during May 2015.


Figure S6: Hourly diurnal profiles $\left(25^{\text {th }}, 50^{\text {th }}\right.$, and $75^{\text {th }}$ whiskers) of OM:OC, O:C, and H:C of NR-PM ${ }_{1}$ OC at Clinton Drive measured by HR-ToF-AMS in May 2015.


Figure S7: Active fires (1 month - Terra/MODIS) in May 2015; adapted from the maps created by Jesse Allen and Reto Stockli, NASA Earth Observatory, using data courtesy of the MODIS Land Science Team at NASA Goddard Space Flight Center (https://neo.sci.gsfc.nasa.gov/analysis/index.php).


Figure S8: Three day back trajectories produced using the National Oceanic and Atmospheric Administration's (NOAA's) Hybrid Single-Particle Lagrangian Integrated Trajectory (HYSPLIT) model. Data were collected for air masses ending 500 m above the ground level at the Clinton Drive monitoring site in Houston, TX ( $29.733943^{\circ}$ N, $95.257684^{\circ} \mathrm{W}$, indicated by star); a) from 05-17 May 2015 and b) from 17-29 May 2015 at 1200 UTC (07:00 LT). A new trajectory was started for every 24 hours. Symbols in each trajectory indicate the location of air mass for every 24 hours for up to 3 previous days.


Figure S9: Isoprene emissions reported by industrial facilities in the Houston area in 2015 (https://www.epa.gov/toxics-release-inventory-tri-program/tri-basic-data-files-calendar-years-1987-2016)


Figure S10: Average source contributions to $\mathrm{PM}_{2.5}$ organic carbon ( $\mathrm{OC} ; \mu \mathrm{g} \mathrm{m} \mathrm{m}^{-3}$, squares) in at Clinton Drive in May 2015 obtained for the 9 factor MM-PMF solution. The circles represent the median OC from bootstrapping (BS) and the average OC from displacement (DISP) and BS-DISP error estimation methods. The error bars represent the concentration range of OC from DISP, BS and BS-DISP. The lower and the upper whiskers represent the $5^{\text {th }}$ and the $95^{\text {th }}$ percentiles of the OC concentrations from BS and BS-DISP and the minimum and maximum OC concentrations from DISP.


Table S1: Spectral contrast angle $\left(\theta^{\circ}\right)$ between the mass spectrum of factors in the 3 to 5 -factor AMS-PMF solutions. Numbers in bold indicate close resemblance between mass spectral signatures.

|  | 4-factor solution |  |  |  |  | 5-factor solution |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Factor | 1 | 2 | 3 | 4 | 1 | 2 | 3 | 4 | 5 |
|  | 1 | 2.6 | 49.1 | 32.1 | 75.3 | 44.0 | 2.5 | 31.5 | 51.0 | 76.2 |
| 3-factor model | 2 | 37.4 | 25.0 | 2.7 | 51.8 | 25.8 | 37.1 | 7.3 | 30.3 | 52.5 |
|  | 3 | 76.8 | 45.8 | 52.6 | 0 | 54.2 | 76.7 | 55.9 | 35.7 | 0 |
|  | 1 |  |  |  |  | 46.6 | 0 | 33.7 | 53.0 | 78.0 |
| 4 factor model | 2 |  |  |  |  | 12.0 | 51.7 | 30.0 | 27.7 | 47.8 |
| 4-factor model | 3 |  |  |  |  | 26.9 | 34.3 | 5.5 | 31.6 | 54.0 |
|  | 4 |  |  |  |  | 55.3 | 76.9 | 56.4 | 36.7 | 0 |

Table S2: The MM-PMF input data statistics for the modeled chemical compounds and base model diagnostics obtained for the 9-factor solution.

\left.|  | PMF input data statistics |  |  | Base model diagnostics obtained for the 9 factor |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| solution |  |  |  |  |  |$\right]$

${ }^{\text {a }}$ signal to noise ratio calculated by PMF (EPA, version 5); ${ }^{\text {b }}$ weak, if $\mathrm{S} / \mathrm{N}=0.5-1$, strong, if $\mathrm{S} / \mathrm{N}>1$ and the uncertainties of the compounds categorized as weak are automatically increased by a factor of 3 ; ${ }^{c}$ total variable, which by default categorized as
 difference between the number of non-weak data points in the input data matrix and the number of elements in the two matrices, source profiles and source contributions, taken together

Table S3: Summary of MM-PMF settings for base runs and error estimations

| Parameter | Description |
| :---: | :---: |
| Data type; sample collection/averaging time frame | $\mathrm{PM}_{2.5} ; 05-27$ May 2015 based on daytime (07:00-18:00 LT) and nighttime (19:00-06:00 LT) schedule |
| \# of species | 40 |
| Total variable | Organic carbon |
| \# of samples | 46 |
| \# of factors | 3 to 11 |
| Treatment of missing data | No missing data |
| Treatment of data below detection limit (BDL) | No modifications or censoring of BDL data |
| Treatment of concentrations equal to or less than zero | No modifications or censoring of data $\leq 0$ |
| Lower limit for normalized factor contributions $\mathrm{g}_{\mathrm{ik}}$ | -0.2 |
| Robust mode | Yes |
| Constraints | None |
| Seed value | Random |
| \# of base runs | 20 When developing a solution (3-11 factors) and 100 (5-9 factors) when determining a final solution |
| \# of bootstraps in BS | 100 |
| $\mathrm{R}^{2}$ for BS | 0.6 |
| BS block size | 1 |
| DISP dQ ${ }_{\text {max }}$ | 4, 8, 16, 32 |
| \# of DISP active species | 34 (only the species categorized as strong) |
| \# of bootstraps and $\mathrm{r}^{2}$ for BS in BS-DISP | 100 and 0.6 |
| BS-DISP active species | Elemental carbon, chrysene, benzo(b)fluoranthene, benzo(GHI)perylene, $17 \alpha(\mathrm{H})-21 \beta(\mathrm{H})$-hopane, cholesterol, levoglucosan, methylphthalic acid, 2methylglyceric acid, 2-methylthreitol, 2methylerythritol, 2,3-dihydroxy-4-oxopentanoic acid |
| BS-DISP dQ ${ }_{\text {max }}$ | 0.5. 1, 2, 4 |

Table S4. Elemental ratios, carbon oxidation state, $\mathrm{m} / \mathrm{z} 44$ and $\mathrm{m} / \mathrm{z} 43$ fractions of PMF factors identified in the $\mathrm{PM}_{1} \mathrm{OA}^{\mathrm{OA}}$ during the field campaign.

| PMF Factor | $\mathbf{O}: \mathbf{C}$ | $\mathbf{H}: \mathbf{C}$ | $\overline{\boldsymbol{O S}} \mathbf{c}$ | $\mathbf{f}_{43}$ | $\mathbf{f}_{44}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| HOA | 0.06 | 2.03 | -1.91 | 0.11 | 0.02 |
| CI-SV-OOA | 0.61 | 1.57 | -0.35 | 0.10 | 0.11 |
| LV-OOA | 1.24 | 1.21 | 1.27 | 0.03 | 0.26 |

Table S5: Summary of MM-PMF diagnostics and error estimation statistics obtained for 5 to 9 factor solutions.

| Diagnostic | 5 Factors | 6 Factors | 7 Factors | 8 Factors | 9 Factors |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\overline{\mathrm{Q}_{\text {expected }}}$ | 1164 | 1084 | 1004 | 924 | 844 |
| $\mathrm{Q}_{\text {robust }}$ (minimum) | 8361 | 7290 | 6406 | 5568 | 4841 |
| $\mathrm{Q}_{\text {true }}$ (corresponding to $\min \mathrm{Q}_{\text {robust }}$ ) | 11117 | 9569 | 7961 | 6770 | 5844 |
| $\mathrm{Q}_{\text {robust }} / \mathrm{Q}_{\text {expected }}$ | 7.2 | 6.7 | 6.4 | 6.0 | 5.7 |
| $\Delta\left(\mathrm{Q}_{\text {robust }} / \mathrm{Q}_{\text {expected }}\right)$ | - | 0.46 | 0.34 | 0.35 | 0.29 |
| DISP \%dQ | -9.57E-05 | 0 | 0 | 0 | 0 |
| DISP swaps at $\mathrm{dQ}_{\text {max }}$ $4,8,16$, and 32 | 0 | 0 | 0 | 0 | 0 |
| Factors with BS mapping < $80 \%$ | Mixed biomass burning, non-tailpipe vehicle emissions, and ship emissions (34\%) | Ship emissions (58\%) and mixed biomass burning and nontailpipe vehicle emissions (58\%) | Gasoline engines (73\%) and ship emissions (76\%) | High- $\mathrm{NO}_{x}$ anthropogenic SOA (56\%) and diesel engines (72\%) | High- $\mathrm{NO}_{x}$ anthropogenic SOA (67\%) and biomass burning (54\%) |

Note: When changing from 7 to 8 factor solution high- $\mathrm{NO}_{\mathrm{x}}$ anthropogenic SOA resolved from low- $\mathrm{NO}_{\mathrm{x}}$ anthropogenic SOA. When changing from 8 to 9 factor solution non-tailpipe vehicle emissions resolved from biomass burning. Also, for the 9 -factor solution levoglucosan is completely resolved from isoprene SOA.

## References

Paatero, P., 2013. User's Guide for Positive Matrix Factorization Programs PMF2 and PMF3, Part 1: Tutorial." University of Helsinki, Finland. http://helsinki.fi/~paatero/PMF/pmf2.zip

TCEQ: Texas Commission on Environmental Quality monitoring site information. www.tceq.texas.gov/cgibin/compliance/monops/agc_monthly_summary.pl?user_site=48_201_1035, (accessed May 09, 2017), 2017.

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