# An exploratory ecosystem model of the Bay of Bengal Large Marine Ecosystem 

## Sylvie Guénette

EcOceans, St. Andrews NB, Canada

prepared for the<br>Bay of Bengal Large Marine Ecosystem Project (www.boblme.org/)

December 2013
(revised 27 August 2014)


#### Abstract

The trend towards ecosystem-based management requires the development of tools to gain insights in ecosystem functioning and the impact of fishing on ecosystem structure. This study aims at developing one of such tools, an ecosystem model of the Bay of Bengal built with the Ecopath and Ecosim software. The Bay was divided in three sections considered relatively independent from each other while migrating species were assumed to occupy the entire study area. The available data and the methods to develop the model are described. The static model (Ecopath) was built to represent year 1978 and synthesise available population dynamics and fisheries data. A preliminary Ecosim model was set up to allow exploring interactions between functional groups and the impact of fishing. Time series of abundance, catches and effort were assembled for four functional groups for this purpose.

The Ecopath model is one image of the ecosystem that results from the data available and the choices that were made for each parameter at the input, stage and when balancing the model. Several crucial gaps in biomass estimates and level of exploitation were noted. Changing the assumptions would lead to different biomasses and $P / B$ values and possibly, the strength of the foodweb links. Given the lack of time series to fit the model to a larger number of functional groups, and the use of commercial CPUEs as abundance indices for large pelagics, the uncertainty of temporal simulations is large.

As it stands, the model is a great framework to articulate data, improve research questions, and determine what important piece of information would be useful to gather to answer the most crucial questions. It can be modified and expanded as data become available. The results from temporal simulations should not be used to give quantitative management advice. Instead, current Ecosim simulations could be used as a tool to explore food web dynamics and effects of fishing.


## Table of contents

Abstract .....  3
Introduction ..... 6
Methods ..... 6
Ecopath with Ecosim ..... 6
Study area and model structure .....  7
Catch, effort and CPUE ..... 9
Catch ..... 9
Effort ..... 9
CPUE and fleet ..... 10
Large pelagics effort and CPUE time series (contributed by Rishi Sharma, IOTC) ..... 10
Fish ..... 11
Diets ..... 11
Fish surveys ..... 11
Natural mortality ..... 12
$P / B$ and $P / Q$ ..... 13
Sharks and rays ..... 13
Tuna-like and scombrids ..... 14
Bathypelagic species ..... 15
Large piscivores (Lpisc) ..... 15
Carangids ..... 15
Small pelagics (Spel) ..... 15
Large piscivores of commercial interest (Lpisc comm) ..... 16
Other coastal fishes ..... 17
Hilsa ..... 17
Indian mackerel ..... 19
Benthic invertebrates ..... 19
Jellyfish ..... 19
Cephalopods ..... 20
Zooplankton ..... 20
Primary producers ..... 20
Birds and mammals ..... 21
Detritus. ..... 21
Balancing the model ..... 21
Uncertainty and pedigree ..... 28
Ecosim model ..... 30
Acknowledgements ..... 34
References ..... 34
Appendices ..... 39
A1 Functional groups ..... 39
A1.1 Allocation of catch species to functional groups ..... 39
A1.2 Area considered for each functional group ..... 44
A1.3 Functional groups composition ..... 45
A2 Catch, effort and CPUEs ..... 48
A2.1 Total catch (t/km2/year) per fleet and functional group as input in Ecopath for year 1978. ..... 48
A2.2. Effort in horse power days for the east coast of India (Bhathal 2013) and relative effort as used for region 2 in Ecosim ..... 49
A2.3. Biomass, catch, and fishing effort directed to hilsa compiled for Bangladesh, the extrapolation for 1978
1986, and the resulting fishing effort time series used in Ecosim (relative value). ..... 50
A2.4 Effort by large pelagic fleet contributed by Rishi Sharma, IOTC. ..... 51
A3 Diet compositions ..... 52
A3.1 Source of fish diets. ..... 52
A3.2 Original diet matrix ..... 56
A4. Calculation of natural mortality. ..... 59
A5 Marine and inland catches by country for hilsa. ..... 60
A6. Alternative fitting results with Ecosim ..... 62

## Introduction

The Bay of Bengal LME is an embayment of the northeastern Indian Ocean bordered by Sri Lanka, India, Bangladesh, Malaysia, Thailand, Myanmar, Indonesia and the Maldives (Figure 1). It is influenced by the second largest hydrologic region in the world, the Ganges-Brahmaputra-Meghna (GBM) Basin (Heileman et al. 2009). The region is also affected by monsoons, storm surges and tsunamis leading to strong seasonality in water characteristics and in productivity (Heileman et al. 2009).

The ecosystem is also under high pressure from fisheries including destructive practices and high bycatch levels (Heileman et al. 2009), and from habitat destruction and pollution (Holmgren 1994). For instance, the demand for Penaeus monodon larvae for aquaculture has resulted in large catches of other shrimp larvae and zooplankton that is probably having an impact on the ecosystem (Mahmood et al. 1994). Also, the damming of some rivers has destroyed some fisheries and reduced habitat for hilsa (Milton 2010).

The Bay of Bengal Large Marine Ecosystem (BOBLME) Project (http://www.boblme.org/) is a collaborative effort between the United Nations Food and Agriculture Organization (FAO) and the countries bordering the Bay of Bengal to improve regional management of fisheries and the marine environment. The aim of the BOBLME project is to identify threats to the marine ecosystem, improve the livelihoods of coastal communities and secure food resources of the Bay of Bengal. One of its goals is to improve the management of coastal and marine natural resources. Increasingly, fisheries management is moving towards Ecosystem Approaches to Fisheries in the hope of developing strategies providing the right incentive structure for stakeholders (Hilborn et al. 2005) and to protect the ecosystem structure and functioning (FAO 2001 ; Pikitch et al. 2004; Babcock et al. 2005; Gavaris 2009). This requires the development of tools to gain insights in ecosystem functioning and to explore management strategies. In the case of the Bay of Bengal, the first caveat in this process was the gaps in knowledge in the total amount of biomass extracted from the ecosystem. Recent work from the University of British Columbia (UBC) Sea Around Us Project (SAUP) team filled this gap by complementing the FAO database with estimates of illegal, unreported and unregulated catches (IUU) in the region (Zeller et al. 2013).

Several ecosystem models have been published in the study area in the last 20 years. Typically they cover a country coastal area such as the west coast of Peninsular Malaysia (Alias 2003), the southeast coast of India (Antony et al. 2010), and Bangladesh (Mustafa 2003; Rashed-Un-Nabi and Hadayet Ullah 2012; Ullah et al. 2012). Several models are available for the Gulf of Thailand (Christensen 1998; Vibunpant et al. 2003) and the South China Sea (Garces et al. 2003; Chen et al. 2008a; b; Hong et al. 2008; Chen et al. 2009). Each model was built with a specific goal in mind, to explore the effects of fishing in general, the effects of one specific fishery (e.g. Rashed-UnNabi and Hadayet Ullah 2012), or the role of a group of species (Antony et al. 2010).

The present report describes the construction of an ecosystem model for the Bay of Bengal in 1978, using the Ecopath with Ecosim (EwE) software (Walters et al. 1997; Christensen and Walters 2004), including all available data on population dynamics, biomass estimates, and diet compositions. This model was built to cover the entire Bay of Bengal divided in 3 geographic regions and focusses more specifically on pelagic fisheries: hilsa, bigeye tuna, yellowfin tuna, and blue and striped marlins. For these species, times series of biomass estimates or CPUEs, and fishing effort where included to provide a first trial in an Ecosim temporal simulation model (Christensen and Walters 2004; Christensen and Walters 2005) for the period 1978-2010. Other species, demersal and pelagics were grouped as a function of their habitat, and their economic importance for fisheries, preparing for future expansions of the model.

## Methods

## Ecopath with Ecosim

An Ecopath model describes the trophic interactions, synthesizing ecological and fisheries data of an ecosystem at a given time (Walters et al. 1997; Christensen and Walters 2004). These models account for the biomass of each functional group of species, their diet composition, production per unit of biomass ( $P / B$, per year), consumption
per unit of biomass ( $Q / B$, per year), mortality rate from natural causes $(M)$ and fishing ( $F$ ), accumulation of biomass and net migration rate (all rates are annual). The principle behind this ecosystem modelling approach is that, on a yearly basis, biomass and energy in an ecosystem are conserved.

The proportion of the mortality of each group that is accounted by the model (fishing, predation, biomass accumulation, migration) is called ecotrophic efficiency ( $E E$ ). Typically, when biomass estimates are absent, an EE value is provided and the biomass left to estimate by Ecopath. The $P / B$ is considered equal to total mortality under equilibrium condition (Allen 1971) since production and losses would be equal. The total mortality $(Z)$ is thus computed as the sum of fishing mortality (F) and natural mortality (M) which includes mortality by predation.

Ecosim is a tool for dynamic simulations based on the Ecopath model, an instantaneous image of the ecosystem at a given time (Christensen and Pauly 1992b; Walters et al. 1997; Christensen and Walters 2004). Ecosim uses a system of differential equations to describe changes in biomass and flows with the system by accounting for change in predation and fishing

$$
\begin{equation*}
\frac{D B_{i}}{d t}=g_{i} \sum_{j} Q_{j i}-\sum_{j} Q_{i j}+I_{i}-\left(M_{0, i}+F_{i}+e_{i}\right) B_{i} \tag{eq. 1}
\end{equation*}
$$

where $g_{i}$ is the net growth efficiency; $Q_{j i}$ and $Q_{i j}$ are the consumption of group $j$ by group $i$ and the consumption of group $i$ by group $j$ respectively; $I_{i}$ the immigration in $t / \mathrm{km}^{2} ; M_{o, i}$ the annual instantaneous rate of non-predatory natural mortality; $F_{i}$ the annual rate of fishing mortality; and $e_{i}$ the emigration rate. The estimation of consumption of prey $i$ by predator $j\left(Q_{i j}\right)$ at each time step is based on the foraging arena theory (Walters and Kitchell 2001; Christensen and Walters 2004) and calculated as:

$$
\begin{equation*}
Q_{i j}=\left(a_{i j} v_{i j} B_{i} B_{j}\right) /\left(2 v_{i j}+a_{i j} B_{j}\right) \tag{eq. 2}
\end{equation*}
$$

where $a_{i j}$ is the rate of effective search for prey $i, v$ (vulnerability) is the rate of exchange between the vulnerable and invulnerable prey biomass pools. The estimate of $a_{i j}$ is obtained by solving equation 2 using parameters from the Ecopath model and conditional on the value of $v$ (default value=2). Low vulnerability ( $1>v_{i j}<1.5$ ) implies a donor-control or type II functional response, while a large value implies that a change in biomass of the predator will cause a corresponding change in the mortality rate of its prey. Model fitting is achieved by estimating vulnerability values that minimizes the sum of squares of differences between model predictions and times series of biomass and catch.

## Study area and model structure

The study area is the Bay of Bengal defined as the LME Bay of Bengal extended to include the northern Sumatra and the Maldives (Figure 1). The study area covers $6,205,000 \mathrm{~km}^{2}$ of which $12 \%$ are on the continental shelf (<200 m , Table 1).

Based on 2 transects on the coast and the open seas in the western Bay of Bengal, the sea surface temperature varies from 17 to $29^{\circ} \mathrm{C}$ in the first 100 m , for an average of about $25^{\circ} \mathrm{C}$ while the first 50 m are mostly above $26^{\circ} \mathrm{C}$ (Prasanna Kumar et al. 2007, figure 2). Below 100 m , the temperature declines rapidly from $20^{\circ} \mathrm{C}$ to less than $12^{\circ} \mathrm{C}$ at 300 m (figure 2 in Prasanna Kumar et al. 2007). For modelling purposes, water temperature is assumed to be $26^{\circ} \mathrm{C}$ on the shelf and $12^{\circ} \mathrm{C}$ in deeper waters (the average temperature at about 250 m ). The oxygen minimum zone reaches as high as 50 m at some time of the year and can become a constraint for the biota but this factor is not included in the model.

The Bay of Bengal was divided in 3 regions: 1. the Maldives and open waters; 2. Sri Lanka, Indian coast, and Bengladesh; 3. the eastern coast of the Bay that covers the coast of Myanmar, Thailand, Malaysia, Sumatra and the Andaman and Nicobar Islands (Figure 1). In the Maldives, characterised by a very small reef area, the fisheries pursued mainly tuna and other large pelagics. Region 2 is characterised by a relatively small shelf under the influence of several rivers. Bangladesh and the West Bengal (India) are the most strongly influenced by the


Figure 1. Map of the study area, the Bay of Bengal LME and the Maldives. Source UBC SAUP.

Table 1. Shelf and EEZ surface area by country and region of the study area.

| Region | Entity | $\begin{gathered} \text { EEZ Area } \\ \left(\mathrm{km}^{2}\right) \\ \hline \end{gathered}$ | $\begin{gathered} \% \\ \text { total } \end{gathered}$ | Shelf Area $\left(\mathrm{km}^{2}\right)$ | $\%$ <br> shelf |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Maldives | 915,423 | 14.8 | 30,998 | 3 |
| 1 | High Seas | 1,929,874 | 31.1 | 0 | 0 |
| 2 | Bangladesh | 84,846 | 1.4 | 64,007 | 75 |
| 2 | India | 666,670 | 10.7 | 118,304 | 18 |
| 2 | Sri Lanka | 530,943 | 8.6 | 31,352 | 6 |
| 3 | India (Andaman \& Nicobar) | 659,573 | 10.6 | 219,820 | 33 |
| 3 | Indonesia (Western) | 719,333 | 11.6 | 133,939 | 19 |
| 3 | Malaysia (West Peninsula) | 68,317 | 1.1 | 67,717 | 99 |
| 3 | Myanmar | 511,356 | 8.2 | 219,820 | 43 |
| 3 | Thailand | 118,717 | 1.9 | 50,210 | 42 |
|  |  | 6,205,051 | 100 | 936,168 |  |

Bramaputra-Ganges river system. Region 3 has a larger shelf, but is less influenced from river discharge than region 2.

In each region, coastal fish and invertebrates, identified with the region number, are assumed to be relatively isolated from other regions. Some groups are assumed to occupy the whole study area (e.g. oceanic sharks and tuna-like) or straddle 2 regions or all three (e.g. hilsa, Indian mackerel, coastal sharks, coastal scombrids, jellyfish). The area used for each group is listed in Appendix A1.2.

## Catch, effort and CPUE

## Catch

We used the catches provided by the UBC Sea Around Us Project (SAUP) structured by country, sectors (subsistence, artisanal, industrial, tuna), and taxa, combining landings statistics and estimates of unreported catches (Zeller et al. 2013). Catches labeled miscellaneous fish, Perciformes and Pleuronectiformes were attributed to functional groups of the same region in equal proportions between fish groups (Appendix A1). Catches labeled as Scombrids and Scombroidei were attributed to functional groups tuna-like and coastal scombrids and Indian mackerel. Tuna catches labeled Thuninni were attributed to both tuna-like and coastal scombrids.

Landings (IUU + landings) and discards were summed by region defined as where the fish was caught regardless of the fishing country. Catches from unknown fishing grounds were assigned to the same region as that of the fishing country. The 1978 catches input into the Ecopath model are the sum of all landings and discards (Appendix A2.1) because the distinction is not necessary to balance the model and because in Ecosim, it is not yet possible to input separate times series of discards. Thus, the temporal simulations (Ecosim) will run with total catch time series.

Since catches were not reconstructed on the basis of effort (Zeller et al. 2013), it is difficult to link effort and fishing sectors which are defined differently among countries. In addition, the subsistence catch is derived from questionnaires, per capita consumption, and human population size, and thus not directly linked on the number of inshore boats. For the time being, artisanal and subsistence catches are grouped under small-scale and attributed to small vessels.

## Effort

Effort by type of boats has recently been estimated for the Indian fleet during the period 1950-2005, compiled by size of boats and engine power (horse power, hp) (Bhathal 2013). The small-scale fleet includes non-motorized traditional vessels and motorized traditional vessels with outboard motors of less than 50 hp (usually 7-9 hp). Industrial includes vessels using inboard motors of 50 hp and above and deep-sea fishing vessels using engines of 120 hp and above. The industrial effort (labeled large vessels, A2.2) includes that of the large trawlers, fishing for lobsters, prawns and catching demersal fishes. The total effort account for the horse power and the number of days fished in a year (hp days, Appendix A2.2).

The number of boats by type (inboard, outboard and non-motorised) was available for Sri Lanka (Joseph 1999; Samarayanke 2003) but it does not directly correspond to the catch statistics sectors (Zeller et al. 2013). The Sri Lanka boats with inboard engines ( 9 m long, 3.5 GT ) may correspond to the industrial fleet catches while the boats with or without outboard engines can be attributed to the artisanal sector. In absence of correspondence between the Indian (hp days) and Sri Lanka (number of boats) data on effort, and absence of data in Bangladesh, the Indian effort (small-scale and industrial fleets) was used as a representative trend for region 2.

For all other countries, it was not possible to obtain fishing effort. Most measurements of effort by type of boats (mechanised, non-mechanised, large vessels) cover only a few years and are known to be incomplete especially for small-scale fisheries. A reconstruction of effort time series is underway for several regions of the world at the UBC Fisheries Centre but there are still serious difficulties for the estimation for the small-scale and subsistence fishery (Krista Greer, UBC Fisheries Centre, pers. comm.).

Effort for hilsa consists of the number of boats in both marine and freshwater in Bangladesh (R. Sharma, IOTC, pers. comm.) (see Appendix A2.3). It was used as a proxy for the total effort on the species assuming that effort evolved the same way in other countries (and especially West Bengal). This assumption seems viable as Bangladesh was responsible for most of the catch (63-85\%) over the study period. However, the effort time series encompassed the period 1987-2006 only. It was extrapolated to 1978 by assuming that effort increased by only $10 \%$ from 1978 to 1987 because fishing has not changed dramatically before 1987, but it changed considerably in fleet and spatial structure in the early 1990s (R. Sharma, pers. comm.) (Figure 2). Between 2006 and 2010 effort was held constant at the 2006 level.


Figure 2. Nominal fishing effort for hilsa in number of boats as compiled for the stock assessment study (lines) and extrapolations from 1978 to 1986 (symbols)

## CPUE and fleet

Measures of catch per unit effort (CPUE) are available for Indian mackerel (1995-2009), skipjack (Katsuwonus pelamis in the Maldives, 1985-2011), kawakawa (Euthynnus affinis, 2004-2011), bigeye tuna (Thunnus obesus,1960-2012), yellowfin tuna (Thunnus albacares, 1963-2012), striped (Kajikia audax) and blue (Makaira nigricans) marlins (1971-2012), and 12 shark species in Sri Lanka (1972-2011) (Rishi Sharma, IOTC, Seychelles, pers. comm.). These could be used as time series of abundance but they still have to be paired with effort or fishing mortality to drive the model which is only possible for four of these species.

The fleet structure for the model consist of two general fleets, industrial and small-scale, in region 2 and only one general fleet each in regions 1 and 3, a fleet dedicated to hilsa, and three fleets dedicated for yellowfin tuna, bigeye tuna and marlins respectively (Appendix A2.1). For each fleet a time series of effort is required. In absence of data for regions 1 and 3 and given that it would be unrealistic to assume no increase in effort since 1978, effort was assumed to increase linearly 3-fold between 1978 and 2010, which is half of the increase of the Indian smallscale fleet. Ecosim uses relative time series of effort so that hp days or number of hooks were rescaled to start at 1 in 1978 (Appendix A2.4).

## Large pelagics effort and CPUE time series (contributed by Rishi Sharma, IOTC)

Effort and CPUE series were estimated for bigeye and yellowfin tuna based on data provided by Far Seas Research Agency, Japan (Dr. Matsumoto pers. comm.) on aggregated data of the Japanese longline fleet from 1960-2012 (52 years). Appendix A2.4 shows the data rescaled by fleets and areas of the Indian Ocean. For the yellowfin dataset, 5 areas are used to stratify the Indian Ocean and the NE Indian Ocean overlapping the BOB area
is used. Effort is measured as the total number of hooks deployed by the area/region and by year (and quarter). For the bigeye tuna, three areas are used to stratify the CPUE data, and the eastern Indian Ocean that overlays the BOB area is used as to display effort and CPUE trends over time (data provided by Dr. Matsumoto, Far Seas Research Agency, Japan).

For marlins, data aggregated for the entire Indian ocean is used to measure trends in CPUE and effort based on the Japanese and Taiwanese LL fleet from 1971 to 2012, though data across species of marlins were aggregated based on equal weights (for blue and striped marlin) as catches for these species are not well disaggregated in the data.

For little tuna, like skipjack, data used were based on boat days of the pole and line fleet operating in the western Indian Ocean off the Maldives atolls (data provided by Dr. S. Adam, Marine Research Center, Maldives; see Appendix A2.4). The same type of data set was used for mackerel tuna (kawakawa). In both cases, the effort in number of boat-days was stratified by month and quarter though finally aggregated on a yearly basis. Data was limited and only available from 2004-2011 (8 years) and for this reason, were not used with Ecosim.

## Fish

Of the 723 species listed for the Bay of Bengal, about 315 had growth, longevity or diet information. These species are listed in Appendix A1.3, grouped into functional groups. The name of fish functional groups is either based on the family name (e.g. Carangids) or the body size (Small, Medium, Large), the habitat (pelagic, oceanic, coastal), and their feeding habit: invertivore (inv) or piscivore (pisc). When assigned to a specific region, the name of a functional group starts with the region number (e.g. 2 L pisc).

## Diets

Diets for fish were mainly taken from Fishbase (Froese and Pauly 2013) using preferably local and regional studies (see sources in Appendix A3.1). Diets from the FAO area 57 account for $8 \%(N=12)$, other parts of the Indian Ocean $9 \%$, the South China Sea 15\%, and the Pacific (mainly western) 55\%. Most diets for large pelagics (sharks, tuna-like, scombrids) were obtained from Atlantic waters.

Diets were compiled in a spreadsheet following the breakdown found in each publication. Foreign diet items were assigned a functional group based on size, trophic level and habitat similarity. In a large number of cases, a part of the diet is not defined (bony fish, finfish, etc.). These items were allocated to functional groups as a function of habitat and size possibilities. Thus, there is large uncertainty in the diet of several groups. In addition, some species considered as separate functional groups (e.g. hilsa, Indian mackerel) may be under-represented in diets, especially for diets from outside the study area.

Coastal diet items consumed by large pelagics (sharks, tuna-like, scombrids) and other functional groups distributed in the whole area, were allocated by region in the same proportion as the region's shelf area ( $3.3 \%$, $22.8 \%$, and $73.9 \%$ for regions 1,2 and 3 respectively). This assumes that these species were distributed or obtained food equally on shelves as a function of the area available. This allocation works well for region 1 especially, the Maldives being a good area for sharks for instance, but with such a small shelf that the biomass of coastal fish is rather small compared to other regions. Being that there is more shelf area in region 3, most the predation on coastal fish for these groups would occur in region 3, an assumption that is debatable. The diets of hilsa and Indian mackerel were allocated to regions in proportion of the catch by region, assuming that catch reflects abundance.

## Fish surveys

Estimates of biomass present in the area was obtained from the summary of the Dr. Fridtjof Nansen Programme 1975-1993 (Sætersdal et al. 1999) that covered Sri Lanka and the eastern coast of the Bay of Bengal, from Bangladesh to northern Sumatra, in 1978-1980. The survey did not cover India and the results for the Maldives are exploratory only as the vessel was too big to maneuver efficiently among atolls (Strømme 1983). The survey was conducted using acoustics on the shelf (< 200 m depth) paired with trawl surveys conducted by local vessels. The report presented biomass estimates for pelagics and semi-demersals separately and then the trawl catch rates ( $\mathrm{kg} / \mathrm{hour}$ ) for each family, and by country, and portion of the coasts in some cases (Myanmar, Sri Lanka, and

Thailand). In some areas, biomasses varied greatly among seasons so semi-annual surveys were averaged to provide an annual biomass. The biomass of each family was calculated as the global tonnage for a category of fish (pelagic, semi-demersal) multiplied by the proportion of the weight they represent in the trawl survey. When indications of dominant species were given, they were used to allocate the biomass to functional groups. The highest biomass is estimated to be around Sri Lanka and Bangladesh (Table 2) which corresponds pretty well to the map of primary production from 1997-2010
(http://www.Ime.noaa.gov/index.php?option=com content\&view=category\&layout=blog\&id=50\&|temid=82).
In absence of biomass estimates for the east coast of India, estimates of densities for the Bangladesh coast were used for the West Bengal ( $14 \%$ of the Indian coast, Bhathal 2005) while the estimates for the Sri Lanka coast was used for the reminder of the Indian coast (This will be named rule 1 in the following sections). Ideally, estimates for the Indian coast would be added to the next version of the model. In a given region, the average biomass is weighted by the surface of each coast section.

The 1983 Dr. Fridtjof Nansen survey in the Maldives (Strømme 1983) recorded mainly deepwater species such as Peristedion adeni, Synagrops spp., Chlorophthalmus spp. and species from the Myctophidae family but also a few small pelagics (Cubiceps sp, Spratelloides gracilis), cephalopods, jellyfish, a few elasmobranch and caranx. It is difficult to attribute the estimated total biomass to any specific groups.

The 1986-1988 survey (Anderson et al. 1992) covers a few exploited groups, on and around the reefs (snapper, emperor, groupers, jacks, sharks),

Table 2. Resulting fish density ( $\mathrm{t} / \mathrm{km}^{2}$ ) in each country surveyed by the Fridtjof Nansen programme and area covered compared to shelf area estimated by the SAUP.

|  | Density $\left(\mathbf{t} / \mathbf{k m}^{\mathbf{2}}\right)$ |  | Area $\left(\mathbf{k m}^{\mathbf{2}}\right)$ |  |
| :--- | ---: | ---: | ---: | ---: |
| Country | Pelagic | Demersal | Nansen | SAUP |
| Sri Lanka | 4.42 | 8.89 | 23,010 | 31,352 |
| Myanmar | 2.98 | 2.27 | 29,067 | 219,820 |
| Bangladesh | 3.23 | 3.15 | 41,211 | 64,007 |
| Malaysia | 3.19 | 0.62 | 41,211 | 67,717 |
| Thailand | 1.83 | 0.74 | 41,211 | 50,210 |
| Sumatra | 1.88 | 2.67 | 85,857 | 133,939 | in 3 different habitats (atoll basins, outer atoll reef, and shallow reefs) throughout the Maldives. Reef fish were deemed not heavily exploited in the early 1990s as the Maldives is a tuna fishing nation (Anderson et al. 1992). The biomass of each group was obtained by using the average proportion in the survey catch and the total biomass collected in each habitat. Nowadays however, reef fish are heavily exploited to supply the tourist consumption in resorts (Hemmings et al. 2011).

As the area surveyed does not include the whole shelf (Table 2), the density of coastal species (Lpisc, Carangids, S pelagics, L pisc comm, SM inv, SM pisc, milkfish plus, hilsa and Indian mackerel) are assumed to be the same across the country continental shelf. For convenience and comparability species densities were first calculated for their habitat in each region. Typically, oceanic species (e.g. large tuna-like, oceanic sharks) are attributed to the entire study area, while coastal species are restricted to the EEZ of each region. Coastal scombrids are assumed to constitute only one stock travelling in all EEZ of the Bay of Bengal ( $69 \%$ of the area), while deep water species were assumed to be present in all deep waters ( $85 \%$ of the study area) (Table 1).

## Natural mortality

In absence of direct measurement, natural mortality $(M)$ is usually estimated using various empirical methods based on catch curves of unexploited populations, longevity, and other empirical relationships (see Kenchington 2013 for a review). None of these methods are valid for all species and all types of life history. I used two different empirical relationships: Pauly's (1980) equation based on growth and water temperature, and Hoenig's empirical relationship based on longevity (Tmax, Hoenig 1983). The temperature used in Pauly's equation is $26^{\circ} \mathrm{C}$ for most fish except demersal fish with distribution deeper than 50 m for which a value of $12^{\circ} \mathrm{C}$ was assumed.

Estimates based on Pauly's equation $M_{p}$ are sensitive to growth rate, increasing with von Bertalanffy's $k$. Growth parameters varying widely among populations and samples/studies, several values were extracted, mainly from Fishbase (Froese and Pauly 2013), to illustrate the potential level of variation. Unless only one valid growth
equation was available a high and a low value were used to calculate $M_{P}$, leading to estimates that can easily vary by $100 \%$ or more (Appendix A4: spreadsheet). Estimates based on Hoenig's equation ( $M_{H}$ ) are sensitive to estimates of longevity that can be seriously underestimated in heavily exploited populations. Thus, when several estimates of longevity were found, the highest estimate was chosen. For each functional group, a minimum and maximum estimate was calculated based on the average over all species for which one or several estimates were produced. The minimal estimate was preferred for large fish while the highest estimate was preferred for smaller species to account for predation (Table 3).

The estimates for tuna-like groups (tuna, sailfish, marlin, bonito) based on $M_{H}$ were often similar to the lowest of average $M_{p}$ estimate (Appendix 4). This corresponds very well with the large growth rate often observed for these species. A large growth rate is useful to avoid predation and decreased mortality rate in early stages of development and is followed by lower mortality for the rest of their life span. For instance, the growth rate estimated for Acanthocybium solandri is very high (Zischke et al. 2013) while longevity is close to 10 years, showing the discrepancy between some life histories and the principle behind some empirical equations.

## $P / B$ and $P / Q$

The production per unit of biomass $(P / B)$ was calculated as the sum of natural (M) and fishing (F) mortalities. Fishing mortality is the ratio catch/biomass often called exploitation rate. F was highly dependent of the biomass estimated from survey and was often found to be too high when biomass was underestimated.

As discussed in the natural mortality section, measurement of growth is highly variable and is likely to create large uncertainty in the calculation of the consumption per unit of biomass ( $O / B$ ). In addition, empirical equations generally used to calculate $Q / B$ from growth parameters, temperature and type of diet (Christensen and Pauly 1992a) or with the addition of

Table 3. Natural mortality used in the model compared to the average minimum and maximum values in each functional groups.

| Fish groups | M min | M max | M used |
| :--- | ---: | ---: | ---: |
| Oceanic sharks | 0.17 | 0.22 | 0.17 |
| Coastal elasmobranch | 0.35 | 0.43 | 0.35 |
| Tuna-like | 0.32 | 0.69 | 0.32 |
| Coastal scombrids | 0.56 | 0.97 | 0.56 |
| S bathy | 2.14 | 3.97 | $2.14 a$ |
| ML bathy |  |  | $0.37 a$ |
| L pisc | 0.42 | 0.60 | 0.42 |
| Carangids | 0.41 | 0.65 | 0.41 |
| S pelagics | 1.69 | 2.29 | 2.29 |
| L pisc comm | 0.39 | 0.66 | 0.39 |
| SM inv | 1.08 | 1.47 | 1.47 |
| SM pisc | 1.08 | 1.23 | 1.23 |
| Milkfish plus | 0.28 | 0.36 | 0.28 |
| Hilsa | 0.95 | 1.61 | 1.61 |
| Indian mackerel | 1.52 | 2.34 | 1.52 |
| Bigeye tuna | 0.20 | 0.40 | 0.20 |
| Yellowfin tuna | 0.21 | 0.63 | 0.21 |
| Marlins | 0.26 | 0.61 | 0.26 |

a based on 1 species
the aspect ratio of the caudal fin (Palomares and Pauly 1998) tend to overestimate $\mathrm{Q} / \mathrm{B}$ and result in very low gross efficiency $G E=P / Q$ (Guénette 2005). Instead, $P / Q$ was set at 0.15 for large fish (except for tuna and sharks), 0.2 for small and medium fish, and 0.25 for small pelagics.

## Sharks and rays

Sharks and rays were classified in oceanic and coastal species of which sharks are the largest component because of the information available and their presence in reported landings. Catches were labeled by species or by more generic names (family, order) that did not allow differentiating between oceanic and coastal sharks. In the Maldives, coastal sharks were traditionally fished and became the target of an important fishery in the 1970s because of higher demand. Oceanic sharks constituted roughly half the catch at the beginning but accounted for $70 \%$ of the catch in 1998 (Anderson and Waheed 1999). It was thus assumed that in region 1, the proportion of oceanic shark increased linearly from $50 \%$ to $70 \%$ between 1992 and 1998. There was no indication that this was the case in other regions except in Sri Lanka where the fishery expanded beyond the continental shelf with driftnets, increasing the catch of pelagic sharks starting in the 1970s (Joseph 1999). Thus, catches were assumed to contain half coastal sharks for the whole time series in region 2 and 3.

Oceanic sharks are assumed to be present in the whole study area. In region 1, the biomass for the Maldives (Anderson et al. 1992) was estimated at $0.6 \mathrm{t} / \mathrm{km}^{2}$ on the shelf, assuming equal biomasses for oceanic and coastal sharks (Table 4). Using this estimate and assuming a lower density for the open oceans, set at half that of the shelf (an arbitrary choice that could be re-examined by local experts), the average weighted by surface area results in a biomass of oceanic sharks of $0.21 \mathrm{t} / \mathrm{km}^{2}$, leading to $\mathrm{C} / \mathrm{B}=0.01$. In region 2 , the biomass for oceanic shark was estimated at $0.19 \mathrm{t} / \mathrm{km}^{2}$, assuming that the biomass for Bangladesh waters was representative of the whole region, leading to $C / B=0.31$. Biomasses from region 3 were ignored being smaller than the catch (Table 4). The weighted average biomass based on region 1 and 2 yielded a biomass for oceanic shark of $0.20 \mathrm{t} / \mathrm{km}^{2}$ for the Bay of Bengal and a fishing mortality of $0.08 /$ year (Table 4).

Table 4. Sharks biomass, catches and F by region as derived from surveys and catch statistics. See text for the derivation of resulting biomasses.

Coastal sharks are restricted to EEZs (69\% of the study area) which includes the continental shelves and some open waters but excludes the high seas. The weighted average results in a biomass of 0.118 $\mathrm{t} / \mathrm{km} 2$ and $\mathrm{C} / \mathrm{B}=0.31$, a relatively large estimate that may be plausible.

## Tuna-like and scombrids

Tuna-like species are large tuna, marlins, swordfish that are large-bodied with oceanic distribution, present in $100 \%$ of study area. Four of these species were considered separately because time series of CPUE and effort were available and it would be possible to try to fit the times series in Ecosim: yellowfin tuna, bigeye tuna, and marlins composed of 2 species, the blue and striped marlins. There are no estimates of biomass for these species while catches amount to $0.004,0.006$, and 0.0009 $\mathrm{t} / \mathrm{km}^{2} /$ year for yellowfin (YFT), bigeye (BET), and marlins respectively.

In the Maldives, based on surveys, the tuna-like biomass is estimated at $0.05 \mathrm{t} / \mathrm{km}^{2}$; assuming that this would be representative of the EEZ, and assuming half the density in open waters, the resulting biomass would amount to $0.035 \mathrm{t} / \mathrm{km}^{2}$ (Table 5). In Bangladesh, there is no record of tunalike species in catches or surveys. The biomass density for region 2 is thus based on estimates from

| Region | Biomass $\mathrm{t} / \mathrm{km}^{2}$ | Catch $\mathrm{t} / \mathrm{km}^{2} / \text { year }$ | $\begin{aligned} & \mathrm{F}=\mathrm{C} / \mathrm{B} \\ & \text { /year } \end{aligned}$ | $\begin{aligned} & \text { area } \\ & \mathbf{k m}^{2} b \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: |
| Oceanic sharks |  |  |  |  |
| 1 | 0.21 | 0.002 | 0.01 | 2,845,297 |
| 2 | 0.19 a | 0.06 | 0.31 | 1,282,459 |
| 3 | 0.003 | 0.01 | 3.45 | 2,077,295 |
| Result | 0.20 | 0.037 | 0.08 | 6,205,051 |
| Coastal sharks |  |  |  |  |
| 1 | 0.02 | 0.008 | 0.40 | 915,423 |
| 2 | 0.19 a | 0.086 | 0.46 | 1,282,459 |
| 3 | 0.02 | 0.019 | 1.02 | 2,077,295 |
| Result | 0.118 | 0.037 | 0.31 | 4,275,177 |

$a$ based on Bangladesh only
$b$ areas differ whether high seas are included or not in region 1

Table 5. Tuna-like and coastal scombrids biomasses, catches and F by region as derived from surveys and catch statistics. See text for the derivation of resulting biomasses.

| Region | Biomass <br> $\mathbf{t} / \mathbf{k m}^{\mathbf{2}}$ | Catch <br> $\mathbf{t} / \mathbf{k m}^{\mathbf{2}} /$ year | $\mathrm{F}=\mathbf{C} / \mathbf{B}$ <br> /year | Area <br> $\mathbf{k m}^{\mathbf{2}} \boldsymbol{b}$ |
| :--- | :---: | :---: | :---: | ---: |
| Tuna-like |  |  |  |  |
| 1 | 0.035 | 0.012 | 0.35 | $2,845,297$ |
| 2 | $0.124 a$ | 0.031 | 0.25 | $1,282,459$ |
| 3 | 0.008 c | 0.024 | 3.09 | $2,077,295$ |
| Result | 0.124 | 0.017 | 0.14 | $6,205,051$ |
| Coastal scombrids |  |  |  |  |
| 1 | - | 0.006 |  | 915,423 |
| 2 | 0.15 | 0.077 | 0.51 | $1,282,459$ |
| 3 | 0.01 | 0.036 | 3.79 | $2,077,295$ |
| Result | 0.15 | 0.042 | 0.28 | $4,275,177$ |
| based on Sri Lanka only |  |  |  |  | Sri Lanka, assuming that the Indian coast would hold

based on Sri Lanka only
$b$ areas differ whether high seas are included or not in region 1
$c$ based on Myanmar only
similar density as that of Sri Lanka (see rule 1 in the fish surveys section). The resulting biomass and F amount to $0.12 \mathrm{t} / \mathrm{km}^{2}$ and $0.25 /$ year respectively in region 2 . In region 3 , the tuna-like biomass estimate based on surveys from Myanmar alone is too low ( $0.008 \mathrm{t} / \mathrm{km}^{2}$ ), compared to catches $\left(0.02 \mathrm{t} / \mathrm{km}^{2}\right)$. I assumed that densities were similar to that of region 2 based on the estimate from Sri Lanka. The resulting biomass $0.124 \mathrm{t} / \mathrm{km}^{2}$ leads to an $F$ of 0.14 (Table 5).

Coastal scombrids are smaller species of the Scombridae family such as bonito Sarda orientalis, kawakawa Euthynnus affinis and Indo-Pacific king mackerel Scomberomorus guttatus. Their distribution covers 69\% of the study area, excluding the high seas. The biomass of coastal scombrids is based on Bangladesh and Sri Lanka estimates (region 2) that amounts to $0.15 \mathrm{t} / \mathrm{km}^{2}$ and $\mathrm{F}=0.28 /$ year.

## Bathypelagic species

Bathypelagic species were divided in small (S bathy) and medium and large (ML bathy) to account for differences in productivity. These species are important for large pelagics species in open oceans but they also play a role in the diet of other species that live on the shelf especially when the shelf is narrow and the slope steep. In the Maldives, for instance, the acoustic survey around the atolls showed more mesopelagics ( S bathy) than small pelagics (Strømme 1983). The biomass of mesopelagics ( $3.28 \mathrm{t} / \mathrm{km}^{2}$ ) was obtained from estimates provided by Lam and Pauly (2005) based on Gjøsaeter (1978). Bathypelagic species are assumed to be distributed in deeper waters, or $85 \%$ of the study area $\left(5,268,883 \mathrm{~km}^{2}\right)$. There is no estimate of biomass for ML bathy so the value of EE was set at 0.9.

## Large piscivores (Lpisc)

The group Lpisc is composed of barracudas (Sphyraenidae), hairtails (Trichiuridae), dolphinfish (Coryphaena hippurus), threadfins (Polynemidae), needlefish (Belonidae), and eel-like fish (Muraenidae, Congridae, Muraenesocidae).

Table 6. L pisc biomass, catches, and F by region as
In region 3, the biomass is estimated at $0.39 \mathrm{t} / \mathrm{km}^{2}$ on the shelf, leading to $F=0.68$ /year, which may be a very large exploitation rate for the late 1970s (Table 6). The resulting P/B would be very high (1.1/year) for such large-bodied species. The estimates are inexistent or too low compared to catches in regions 1 and 2 , so an EE of 0.9 was assumed for all three regions. Assuming a lower level of fishing in the 1970s, F was set at half that of region 3 in
obtained from surveys and catch statistics. Densities are presented for the shelf area.

| Region | Biomass <br> t/km | Catch <br> $\mathbf{t} / \mathbf{k m}^{2} /$ year | F=C/B <br> /year |
| :--- | ---: | ---: | ---: |
| 1 | - | 0.09 | - |
| 2 | 0.12 | 1.83 | 14.7 |
| 3 | 0.39 | 0.26 | 0.68 | all regions.

## Carangids

The carangids group is composed of trevally, scad, and pomfret, all members of the Carangidae family. Biomass estimates from surveys ( $0.35-1.77 \mathrm{t} / \mathrm{km}^{2}$ ) led to $F$ values of $0.35-0.82$ /year in the three regions (Table 7). In the Maldives (region 1), the biomass estimate is probably low since it is based on the 1986-88 survey Anderson et al. 1992) (see the calculation in the section L pisc comm). In region 3 , the exploitation rate is relatively high ( 0.82 ; $P / B=1.24$ ) for the 1970 s and could be revisited.

Table 7. Carangids biomasses, catches, and F by region as obtained from surveys and catch statistics. Densities are presented for the shelf area.
\(\left.$$
\begin{array}{lrrr}\hline \text { Region } & \begin{array}{c}\text { Biomass } \\
\text { t/km }\end{array} & \begin{array}{c}\text { Catch } \\
\text { t/km }\end{array} \text { /year }\end{array}
$$ \quad \begin{array}{c}F=C/B <br>

/year\end{array}\right]\)| 1 | 0.48 | 0.18 | 0.37 |
| :--- | ---: | ---: | ---: |
| 2 | 1.77 | 0.62 | 0.35 |
| 3 | 0.35 | 0.29 | 0.82 |

## Small pelagics (Spel)

The small pelagics include mainly sardines (Clupeidae), anchovies (Engraulidae), halfbeaks (Hemiramphidae) and flyingfish (Exocoetidae). Initially considered separately, they were grouped because there is little knowledge of their biomass and exploitation rate.

There was no biomass estimate for region 1, although there were indications that small pelagics were not very abundant in the Maldives compared to mesopelagics (Strømme 1983). The biomass was left to be estimated using $\mathrm{EE}=0.95$ and assuming a relatively low fishing mortality ( $\mathrm{F}=0.1$ ). In region 3 , this group was estimated at $0.64 \mathrm{t} / \mathrm{km}^{2}$, leading to $\mathrm{F}=0.65 /$ year. In region 2 the biomass was badly underestimated and was left to be estimated by Ecopath assuming the same fishing mortality as region 3 and $E E=0.95$.

## Large piscivores of commercial interest (Lpisc comm)

This functional group is composed of snappers (Lutjanidae), groupers (Serranidae), croakers (Sciaenidae), emperors (Lethrinidae), and lizardfish and Bombay duck (Synodontidae). All these families are targeted by fisheries and could be considered separately if there were better information on their exploitation.

## Region 1

In the Maldives, the 1986-1988 survey (Anderson et al. 1992) covers mainly this functional group, in 3 different habitats (atoll basins, outer atoll reef, and shallow reefs) throughout the EEZ. Reef fish were deemed not heavily exploited in the early 1990s as the Maldives (Anderson et al. 1992) but their exploitation increased since then (see Figure 3) to supply the tourist consumption (Hemmings et al. 2011). The biomass of each group was obtained by combining the average proportion in the survey catch and the total biomass (sharks, tuna, carangids, Lpisc comm, and others) collected in each habitat. The shallow reefs pose a problem because the total biomass was not estimated due to the large variability and uncertainty using handlines. A crude estimate was obtained by combining various sources of information and a method similar to that described in the most recent survey for groupers in the Maldives (Darwin Reef Fish Project 2011). Anderson et al. (1992) provided a crude estimate of MSY based on the Philippines reefs ( $1-2 \mathrm{t} / \mathrm{km}^{2}$, I used 1.5). Using the Cadima equation ( $\mathrm{MSY}=0.5^{*}(\mathrm{~F}+\mathrm{M}$ ) *avg biom, Garcia et al. 1989), $\mathrm{M}=0.22$ (the values estimated for groupers in the study area vary from 0.22-0.42, Appendix 4), $\mathrm{F}=0.05$ (a low estimate of mortality), the average biomass would amount to $38,889 \mathrm{t}$ for the entire functional group in the shallow reefs areas. Using the average percentage of groupers caught in night and day sampling (16\%, Anderson et al. 1992), groupers biomass is estimated at $6,178 \mathrm{t}$ or $1.8 \mathrm{t} / \mathrm{km}^{2}$, while the biomass estimates for groupers in atolls basins and in deep reef slopes amount to 0.45 and $2 \mathrm{t} / \mathrm{km}^{2}$ respectively. The value for the 19861988 survey in shallow areas is lower than that estimated using a more recent survey ( $3.4 \mathrm{t} / \mathrm{km}^{2}$, Table 9) (Darwin Reef Fish Project 2011).

The same calculation for the entire L pisc comm functional group (groupers, snappers, emperors) and all habitats leads to a biomass of $4.48 \mathrm{t} / \mathrm{km}^{2}$, catches of $0.15 \mathrm{t} / \mathrm{km}^{2}$ and $\mathrm{F}=\mathrm{C} / \mathrm{B}=0.03$ in 1986-88. I assumed a similar value of F for the period 1978-1980. Estimates for Carangids were obtained in the same manner in region 1, assuming they comprise $16 \%$ of the catch in shallow reefs.

The fishery for grouper started late in the area compared with snappers for instance (Figure 3) and remained below 1000 t during the whole time series. In 1986-88 the catch is estimated at 40 t which means a ratio $\mathrm{C} / \mathrm{B}$ of 0.006 compared with 0.029 in 2010 (Table 9). The Darwin Reef Fish Project (2011) estimated the catch for 20102011 at 950 t based on exports augmented with $20 \%$ in transport mortality, for a total of 1140 t , more than twice as much as the reconstructed catch from SAUP. This results in a larger C/B of 0.03, a low exploitation rate that would suggest a sustainable fishery. In contrast, there are reports of catch decline, rarity of some species, decrease in mean length, and increasing proportion of immature in the catch since 1987-1991 (Darwin Reef Fish Project 2011). The mixed signals may be due to the overestimation of biomass and underestimation of catch, and perhaps a lack of information by species.

Table 9. Estimate of biomass in the Maldives for the L pisc comm group (total biomass) and groupers (Anderson et al. 1992) compared to a more recent survey (Darwin Reef Fish Project 2011) in shallow reef areas.

| Study | Year | $\frac{\text { Area }}{\mathrm{km}^{2}}$ | Total biomass estimate |  |  |  | Groupers' biomass estimate |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Biomass ${ }^{\text {a }}$ |  | MSY used |  | Biomass |  | MSY |  | Catch SAUP | C/B |
|  |  |  | t | $\mathrm{t} / \mathrm{km}^{2}$ | $t$ | $\mathrm{t} / \mathrm{km}^{2}$ | t | $\mathrm{t} / \mathrm{km}^{2}$ | t | $\mathrm{t} / \mathrm{km}^{2}$ | t | /year |
| (Anderson et al. 1992) | 1986-88 | 3,500 | 38,889 | 11 | 5,250 | 1.5 | 6,178 ${ }^{\text {b }}$ | 1.77 |  |  | 40 | 0.006 |
| (Darwin Reef Fish |  |  |  |  |  |  |  |  |  |  |  |  |
| Project 2011) | 2010-11 | 4,513 |  |  |  |  | 15,486 | 3.43 | 2,118 | 0.47 | $443^{\text {c }}$ | 0.029 |

[^0]

Figure 3. Catches of three families composing the functional group $L$ pisc comm. in region 1 (SAUP compilation Zeller et al. 2013).

## Regions 2 and 3

In region 2, the biomass estimate from the survey ( $4.42 \mathrm{t} / \mathrm{km}^{2}$ ) leads to an $F$ value of 0.51 /year (Table 10). In region 3 the biomass is estimated at $0.24 \mathrm{t} / \mathrm{km}^{2}$, probably an underestimate, leading to $F=1.6 /$ year. The biomass for region 3 was left to be estimated by the model with a $P / B$ equal to that of region 2 , assumed to be more reasonable.

Table 10. L pisc comm biomass, catches, and F by region as obtained from surveys and catch statistics. Densities are presented for the shelf area.

| Region | Biomass <br> $\mathbf{t} / \mathbf{k m}^{2}$ | Catch <br> $\mathbf{t} / \mathbf{k m}^{2} /$ year | $\mathrm{F}=\mathrm{C} / \mathrm{B}$ <br> /year |
| :--- | ---: | ---: | ---: |
| 1 | 4.48 | 0.15 | 0.03 |
| 2 | 4.42 | 2.24 | 0.51 |
| 3 | 0.24 | 0.39 | 1.60 |

## Other coastal fishes

The other coastal fishes are demersal and benthopelagic species of small and medium body size divided into invertivores (SM inv), piscivores (SM pisc), and Milkfish plus. The invertivores include seabreams (Sparidae), spinefoot (Siganidae), mullet (Mugilidae), surgeonfish (Acanthuridae), parrotfish (Scaridae), goatfish (Mullidae), ponyfish (Leiognatidae) and several others (Appendix A4). The piscivores include grunts and sweetlips (Haemulidae), threadfins (Nemipteridae), turkeyfish (Scorpaenidae) and many other species. Milkfish plus are large invertivore fish including Chanos chanos, Pangasius pangasius, and Netuma thalassina. Very little is known of their exploitation rate and population dynamics. They were separated for ecological reasons (predator-prey relationships) more than for the necessities of the model dynamics. There are no biomass estimates except for SM inv in region 3 which is probably too low ( $0.63 \mathrm{t} / \mathrm{km}^{2}$ ). The exploitation rate was assumed to be relatively low for these species ( $\mathrm{F}=0.1$ ) and $\mathrm{EE}=0.95$.

## Hilsa

Currently, the main species of hilsa fishery of Bangladesh is Tenualosa ilisha that contributes more than $99 \%$ of the total hilsa catches (Milton 2010). For modelling purposes, hilsa is considered as a single stock that is straddling regions 2 and 3 in the Bay of Bengal (Milton 2010; BOBLME 2012).

Catches reconstructed by the Sea Around Us Team (Zeller et al. 2013) were compared with the compilation used for the stock assessment performed in 2012 (BOBLME 2012) for Bangladesh. Marine catches estimated by the SAUP team is $10 \%$ higher (Figure 4, Appendix A5), probably because of the effort in accounting for discards,
subsistence fishery, and unreported catches. The inland catch statistics from both sources (FAO ${ }^{1}$ and the assessment document (BOBLME 2012)) are very similar (Figure 4).

The earliest estimate of inland catches is for 1984 in Bangladesh (90,000 t) and 1992 in India (39,298 t). Assuming that inland catches were at least as large in 1978, the total catch amounts to 143,521 t in Bangladesh and 80,190 t in India. Malaysian catches are rather modest ( $4,569 \mathrm{t}$ ) and composed of Tenualosa macrura (Rudolf Hermes, FAO, pers. comm.) while in Myanmar, hilsa catches are undistinguishable from other clupeids and cannot be included here (A6). In 1978, the marine catch estimated for Bangladesh constitutes about $60 \%$ of that of the entire Bay of Bengal.

The recent stock assessment performed in 2012 provides a time series of biomass based on Bangladesh marine and freshwater catches and CPUE time series for the period 1984-2006, and CPUE time series compiled for Bangladesh fishery (BOBLME 2012). The biomass estimate was obtained using a surplus production model for the period 1987-20006, and projections were made until 2010 based on observed catches and projected effort reductions (Rishi Sharma, pers. comm.). According to the assessment, exploitation rate (=C/B) for this stock increased from 0.14 in 1987 to 0.29 in 2006. The biomass was scaled for the Bay of Bengal ( $\mathrm{B}_{\text {tot }}$ ) by using the 1987 C/B ratio (0.14/year) to divide the estimated Bay of Bengal catch (regions 2 and 3 ) for a minimum estimated biomass of $1,581,156 \mathrm{t}$ or $1.75 \mathrm{t} / \mathrm{km}^{2}$.


Figure 4. Hilsa catches in marine waters and inland in Bangladesh, as reconstructed by the Sea Around Us Project (SAUP), FAO, and the Bay of Bengal team (BOB).

Based on the assessment, it was concluded that the stock was below optimal yield in both Bangladesh and India and that current catch levels may not be sustainable (BOBLME 2012). In the 1960s, hilsa was composed mainly of 3 -year-old while $90 \%$ of the commercial catches was composed of fish of less than 1 year old in 1999, showing signs of overfishing (Milton 2010). Also, a larger portion of the habitat is exploited nowadays. The depletion may have been partially masked by the introduction of mechanised boats and nylon twine in the early 1980s, which allowed the fishery to move from rivers and coastal areas to increasingly wider areas of the Bay extending up to 200-250 km from the coastline (Milton 2010). Rivers constitute an important habitat as juveniles spend 7 months growing in rivers and then spend most of their lives in the ocean (Amin et al. 2008; Milton 2010). Thus, the damming of some rivers destroyed some fisheries and reduced habitat and hilsa production (Milton 2010). These aspects of the life history are not included in the model.

[^1]
## Indian mackerel

Indian mackerel (mainly Rastrelliger kanagurta) straddles region 2 and 3. Noble et al. (1992) present the earliest stock assessment for Indian mackerel along the Indian coast. Natural mortality was estimated at 1 /year based on maturity at 1 year old and Rikhter and Evanov (1976) method. F was estimated at 1.9/year on the west coast of India for the period 1984-1988 using length cohort analysis (Noble et al. 1992). The authors concluded that the stock was exploited at higher levels than

Table 11. Indian mackerel biomass, catches, and F by region as obtained from surveys and catch statistics. Densities are presented for the shelf area.

| Region | Biomass <br> $\mathbf{t} / \mathbf{k m}^{2}$ | Catch <br> $\mathbf{t} / \mathrm{km}^{2} /$ year | F=C/B <br> /year |
| :--- | ---: | ---: | ---: |
| 2 | 0.79 | 0.61 | 0.77 |
| 3 | 0.06 | 0.24 | 4.28 |
| Total | 0.23 | 0.33 | 1.44 | $\mathrm{F}_{\text {MSY }}$ and that fishing mortality should be decreased by $61 \%$. In contrast, based on catch curves, Abdussamad et al. (2010) estimated fishing mortality at 3.5 to 7 per year between 1997 and 2007 along the Tuticorin coast (Tamil Nadu).

The estimated biomass based on the survey amounts to $0.79 \mathrm{t} / \mathrm{km}^{2}$ in region 2 , leading to ratio $\mathrm{C} / \mathrm{B}=0.77 /$ year (Table 11). In region 3, the biomass estimate is too low, as the catch is 4 times higher. Assuming, $F=0.77 /$ year, the biomass would be at least of $0.43 \mathrm{t} / \mathrm{km}^{2}$ for regions 2 and 3 . Again, an exploitation rate of 0.77 /year seems high for the late 1970s.

## Benthic invertebrates

Benthic animals were first divided into crustaceans, macrobenthos, and meiobenthos. The separation between crustaceans (shrimps and crabs) and other species of macrobenthos was felt necessary for predator-prey relationships and the important commercial shrimp and crab/lobster fisheries especially in Bangladesh (Holmgren 1994). However, the available biomass estimates were given for both groups in aggregation (Holmgren 1994; Ansari et al. 2012). As a first trial, $25 \%$ of the estimated biomass of benthic invertebrates was arbitrarily assigned to crustaceans and $75 \%$ to the macrobenthos group (Table 12). In region 3, biomasses were based on Myanmar, Thailand and Andaman only. In region 2, data was only available for the Indian coast. In region 1, crustaceans and other macrobenthos were grouped for lack of information and the biomass was left to be estimated by Ecopath using an EE of 0.8. The $\mathrm{P} / \mathrm{B}$ ratio was assumed to be similar to the one calculated for the Mauritanian coast ( $0.15 /$ year, Guénette et al. 2014). The P/Q ratio (0.15) was taken from Jarre-Teichmann (1996).

These biomasses were estimated in coastal habitats and ignore deep-water crustaceans. For instance, Suman et al. (2006) sampled $7124 \mathrm{~km}^{2}$ at depth between 200 m and 1000 m and estimated the biomass of shrimps (e.g. Aristeus virilis, Acanthephyra armata, Heterocarpus sp.) and deep-sea scampi (Nephropsis stewarti and Puerulus angulatus) at $0.2 \mathrm{t} / \mathrm{km}^{2}$, a relatively small biomass compared to coastal areas. The authors state that the coastal fishery for shrimps in Indonesia is now taking 268\% of the maximum sustainable yield.

Meiobenthos biomass was obtained from Holmgren (1994). The P/B ratio (9/year) was taken from (Gerlach 1971). The P/Q ratio was set at the same value as that of the macrobenthos (0.15). The biomasses retained for the EEZ assumed that the biomass in deeper waters were half that of the shelf.

## Jellyfish

Table 12. Crustaceans and macrobenthos biomass, catches, and F by region as obtained from surveys and catch statistics. Densities are presented for the shelf area.

| Region | Group | $\begin{gathered} \text { Biomass } \\ \mathrm{t} / \mathrm{km}^{2} \end{gathered}$ | $\begin{gathered} \text { Catch } \\ \mathrm{t} / \mathrm{km}^{2} / \text { year } \end{gathered}$ | $\begin{aligned} & \mathrm{F}=\mathrm{C} / \mathrm{B} \\ & \text { /year } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: |
| 1 | macrobenthos | - | 2.1E-06 | - |
| 2 | crustaceans | 6.82 | 2.24 | 0.33 |
|  | macrobenthos | 13.65 | 4.47 | 0.03 |
|  | meiobenthos | 13.84 | - | - |
| 3 | crustaceans | 2.78 | 0.63 | 0.23 |
|  | macrobenthos | 8.35 | 0.09 | 0.01 |
|  | meiobenthos | 6.86 | - | - |

Jellyfishes can be important in ecosystems and swarm in coastal areas in large numbers for short periods of time (Mills 2001; Hay 2006; Lynam et al. 2006; Brotz et al. 2012). However, the level of knowledge is rather low in the study area. In absence of local biomass estimate, the value for FAO area 57 used for global modelling was kept
 (1996). The diet was also taken from Arai (1996).

## Cephalopods

There is not much information on cephalopods in the study area. In Bangladesh, Sepia sp, Loligo sp and Octopus vulgaris were observed in the 10-100 m depth zone (Hossain 2004).

Octopus mortality was estimated at $0.1 /$ month for a life span of about 1.5 year and adults dying massively after spawning (Morocco population, Robert et al. 2010). Thus, M was estimated at 1.2 /year. $\mathrm{Q} / \mathrm{B}$ was estimated by the model using a $P / Q$ value of 0.3 . Other cephalopods were assumed to have similar natural mortality and $P / Q$ ratio as octopus. The biomasses derived from trawl survey (in region 3 only) underestimated the biomass at levels below catches. Thus, the biomass was left to be estimated by EwE using EE=0.9. Diet compositions were derived from qualitative and quantitative studies for Octopus vulgaris (Gonçalves 1991), Loligo forbesi and L. vulgaris (Rost Martins 1982; Pierce et al. 1994; Hanlon and Messenger 1996), Loligo pealei (Vovk 1985), and Illex illecebrosus (Froerman 1984).

## Zooplankton

The biomass of zooplankton estimated based on 2 transects in the Bay of Bengal aggregated mesozooplankton (copepods, fish and invertebrate larvae, etc.) as well as larger and carnivorous zooplankton such as chaetognaths and euphausids. Thus, the group labeled zooplankton includes both herbivorous and carnivorous zooplankton.

In region 2, the biomass is of the estimate for coastal and open waters of India (Prasanna Kumar et al. 2007, table 2). The average was weighted by surface area, coastal waters representing $18 \%$ of the Indian EEZ (Table 13). The biomass estimate derived for Thailand was not very high ( $2 \mathrm{t} / \mathrm{km}^{2}$ ) probably not representative of region 3, and was not used in the model. Biomass of regions 1 and 3 was left to be estimated by Ecopath using $E E=0.6$. $P / B=24$ and $Q / B=112$ was based on Aydin et al. (2003).

Table 13. Estimates of zooplankton available in the study area. WW=wet weight

| Country | Location | Biomass | Unit | Source | Resulting biomass <br> $\mathbf{g} / \mathrm{m}^{2}$ |
| :--- | :--- | ---: | :--- | :--- | :---: |
| Thailand | coastal | 20 | $\mathrm{mg} \mathrm{WW} / \mathrm{m}^{3}$ | (Holmgren 1994) | $2 a$ |
| India | coastal | 16.93 | $\mathrm{~g} \mathrm{WW} / \mathrm{m}^{2}$ | (Prasanna Kumar et al. 2007, | 10.7 |
|  | open seas | 9.36 | $\mathrm{~g} \mathrm{WW} / \mathrm{m}^{2}$ | table2) |  |

$a$ assuming a mean depth of 100 m ; not used

## Primary producers

Current estimates of phytoplankton production were obtained from the Sea Around Us Project (www.seaaroundus.org/). The production in $\mathrm{mgC} / \mathrm{m}^{2} /$ day was transformed in wet weight per year (WW $\mathrm{t} / \mathrm{km}^{2} /$ year) by assuming a ratio C:WW of 1:9 (Pauly and Christensen 1995). Primary production estimated from other studies is provided for comparison (Table 14).

Assuming a $P / B$ ratio of 100/year, the standing biomass would be estimated at $12 \mathrm{t} / \mathrm{km}^{2}$ in region $1,29 \mathrm{t} / \mathrm{km}^{2}$ in region 2 and $21 \mathrm{t} / \mathrm{km}^{2}$ in region 3 . in comparison, the standing biomass was estimated at $9 \mathrm{t} / \mathrm{km}^{2}$ on the Indian coast and ( $7 \mathrm{t} / \mathrm{km}^{2}$ ) offshore of India (table 2 in Prasanna Kumar et al. 2007).

Benthic primary producers such as seaweeds are known to be important in some areas of the Bay of Bengal, but very little is known about their biomasses and exploitation. The group has been included as a place holder and for diet purposes. The P/B (4.1/year) was derived from detailed work performed in Mauritania (Vermaat et al. 1993). A low EE was assumed for this group.

Table 14. Estimates of primary productivity by country EEZ obtained from the Sea Around Us Project (SAUP) web page compared with other sources. WW=wet weight; $\mathrm{PP}=$ primary production

| Region | Entity | $\begin{gathered} \text { EEZ Area } \\ \left(\mathbf{k m}^{2}\right) \\ \hline \end{gathered}$ | PP g WW/m2/year |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | SAUP, current | Prasanna Kumar et al. $\begin{aligned} & \text { (2007), 2001- } \\ & 2006^{\text {a }} \end{aligned}$ | Holmgren (1994) $<1982^{\text {a }}$ |
| 1 | Maldives | 915,423 | 1261 |  |  |
| 1 | High Seas | 1,929,874 | 1179 | 757 |  |
| 2 | Bangladesh | 84,846 | 5614 |  |  |
| 2 | India | 666,670 | 3324 | $1065{ }^{\text {b }}$ | 1440 |
| 2 | Sri Lanka | 530,943 | 1994 |  |  |
| 3 | India (Andaman \& Nicobar) | 659,573 | 1501 |  |  |
| 3 | Indonesia (Western) | 719,333 | 1820 |  |  |
| 3 | Malaysia (West Peninsula) | 68,317 | 4293 |  |  |
| 3 | Myanmar | 511,356 | 3088 |  |  |
| 3 | Thailand | 118,717 | 2332 |  |  |
|  |  | 6,205,051 |  |  |  |
| ${ }^{\text {a }}$ Using the gC:WW=1:9 to convert PP values <br> b coastal area only |  |  |  |  |  |

## Birds and mammals

Only a few species of marine birds have been listed for the Bay of Bengal, and only partial population estimates mostly outdated, are available for the Bay of Bengal (Mondreti et al. 2013). From the estimates provided in the review document, the region counts a minimum of about 76,000 marine birds ( $\sim 20$ tonnes) for which very little is known. At this point the inclusion of marine birds in the model would act mainly as a place holder.

Marine mammals are also not very well known in terms of species composition and abundance. Some estimates are available for Orcaella brevirostris and Neophocaena phocaenoides in coastal waters of Bangladesh (Smith et al. 2008). Afsal et al. (2008) describe the distribution and number of sightings of 10 species compiled from 35 opportunistic surveys in Indian Seas (continental and Andaman \& Nicobar Islands). Another 16 species are known to occur in Indian waters but were not observed during the survey. No biomass estimates were derived from these observations.

## Detritus

Detritus biomass ( $D$, in $\mathrm{gC} / \mathrm{m} 2$ ) was estimated using an empirical equation [Pauly, 1993 \#3095] based on primary production and depth of the photic zone:
$\log _{10} D=-2.41+0.954 \log _{10}(P P)+0.863 \log _{10}(E)$
where PP is primary production in $\mathrm{gC} / \mathrm{m}^{2} /$ year, and $E$ the euphotic zone (set at 40 m ). Using PP of 1206,2925 and $2141 \mathrm{gWW} / \mathrm{m}^{2} /$ year for regions 1,2 and 3 respectively, and a conversion ratio C:WW of 1:9 (Pauly and Christensen 1995), the total detritus biomass was estimated at $137 \mathrm{t} / \mathrm{km}^{2}$.

## Balancing the model

Using the input values (see Table 15 for parameters and Appendix A3.2 for initial diet composition), Ecopath solves simultaneous linear equations and estimates the missing parameters, often the Ecotrophic Efficiency (EE) value. The balancing process is done manually by checking inconsistencies in data, adjusting biomasses, P/B ratio, and diet composition, starting with parameters that were deemed less reliable (see user guide for more detail on this). As such, diet compositions are often modified on account of the uncertainty caused by seasonal and individual
variation and sampling error. Overestimates of the proportion of rare prey in the diet of an abundant predator is a common source of excessive mortality. The biomasses of several functional groups were deemed uncertain and especially, in region 3 where they were systematically lower per unit area of shelf that in other regions. Thus biomasses were often modified to provide enough prey to predators and balance the model.

In all regions, consumption on carangids was too high, mainly because of biases in SM pisc diet composition in addition to over predation from coastal scombrids and tuna-like groups. The biomass of carangids was increased in all regions and especially in region 3 where the initial biomass was estimated at $0.35 \mathrm{t} / \mathrm{km}^{2}$; and needed to be increased 10 -fold (Table 16). Also, the problematic diets were modified (Table 17). In region 1, some of the problems with carangids were caused by a large biomass of Lpisc as estimated by Ecopath. To decrease the biomass, L pisc cannibalism was decreased and its importance in the diet of other fish (e.g. SM pisc, carangids) was decreased.

The biomass of benthic invertebrates was increased in all regions to accommodate predation pressure, and P/B was increased for SM inv in all regions. Cephalopods are a largely unknown component although they are an important predator in the system. To maintain their biomass at lower levels and reduce pressure on fish, cephalopod cannibalism was reduced. Finally, hilsa biomass was increased from 1.75 to $2.4 \mathrm{t} / \mathrm{km}^{2}$ and that of Indian mackerel from 0.43 to $1.9 \mathrm{t} / \mathrm{km}^{2}$.

Finally, after balancing, some groups seem to be not very well modelled. For instance, the group 1L pisc comm with its large biomass has an estimated EE of 0.4 which can indicate bias in diet or overestimate of their biomass or $P / B$. The same can be said for $S$ bathy as a large portion of their production is unexplained by the model. The biomass of oceanic shark does not seem to be too low resulting in an EE of 0.58, meaning that the mortality unexplained by the model is very high. In contrast, EE is very high for coastal elasmobranch, which is caused in part by the relatively high catch relative to biomass input in the model.

A word of caution: It is important to remember that the current Ecopath model is one image of the ecosystem given the data available and choices made for each parameter at the input stage and the modifications made to balance the model. Changing the assumptions would change biomasses and $P / B$ values and possibly the strength of the foodweb links.

Table 15. Summary of the input parameters into Ecopath

|  | Group name | Habitat area (fraction) | Biomass <br> in habitat <br> area <br> (t/km ${ }^{2}$ ) | P/B <br> (/year) | Q/B <br> (/year) | EE | P/Q | Unassimil. / consumption | Detritus import ( $\mathrm{t} / \mathrm{km}^{2} /$ year) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Oceanic sharks | 1 | 0.202 | 0.249 |  |  | 0.15 | 0.2 | 0 |
| 2 | Coastal elasmobranch | 0.689 | 0.118 | 0.66 |  |  | 0.15 | 0.2 | 0 |
| 3 | Tuna-like | 1 | 0.124 | 0.46 |  |  | 0.2 | 0.2 | 0 |
| 4 | Coastal scombrids | 0.689 | 0.151 | 0.84 |  |  | 0.2 | 0.2 | 0 |
| 5 | Jellyfish | 1 | 0.5 | 3 |  |  | 0.3 | 0.2 | 0 |
| 6 | Cephalopods | 1 |  | 1.2 |  | 0.9 | 0.3 | 0.2 | 0 |
| 7 | S bathy | 0.849 | 3.28 | 2.14 |  |  | 0.25 | 0.2 | 0 |
| 8 | ML bathy | 0.849 |  | 0.37 |  | 0.9 | 0.2 | 0.2 | 0 |
| 9 | 1 L pisc | 0.005 |  | 0.76 |  | 0.9 | 0.2 | 0.2 | 0 |
| 10 | 1 Carangids | 0.005 | 0.48 | 0.78 |  |  | 0.2 | 0.2 | 0 |
| 11 | 1 S pelagics | 0.005 |  | 2.39 |  | 0.95 | 0.25 | 0.2 | 0 |
| 12 | 1 L pisc comm | 0.005 | 4.48 | 0.42 |  |  | 0.2 | 0.2 | 0 |
| 13 | 1 SM inv | 0.005 |  | 1.57 |  | 0.95 | 0.25 | 0.2 | 0 |
| 14 | 1 SM pisc | 0.005 |  | 1.33 |  | 0.95 | 0.25 | 0.2 | 0 |
| 15 | 1 Macrobenthos | 0.459 |  | 2 |  | 0.8 | 0.15 | 0.2 | 0 |
| 16 | 1 Meiobenthos | 0.459 |  | 9 |  | 0.8 | 0.15 | 0.2 | 0 |
| 17 | 1 Zooplankton | 0.459 |  | 24 | 112 | 0.6 |  | 0.2 | 0 |
| 18 | 2 L pisc | 0.034 |  | 0.76 |  | 0.9 | 0.2 | 0.2 | 0 |
| 19 | 2 Carangids | 0.034 | 1.77 | 0.76 |  |  | 0.2 | 0.2 | 0 |
| 20 | 2 S pelagics | 0.034 |  | 2.94 |  | 0.95 | 0.25 | 0.2 | 0 |
| 21 | 2 L pisc comm | 0.034 | 4.42 | 0.89 |  |  | 0.2 | 0.2 | 0 |
| 22 | 2 Milkfish plus | 0.034 |  | 0.38 |  | 0.95 | 0.2 | 0.2 | 0 |
| 23 | 2 SM inv | 0.034 |  | 1.57 |  | 0.95 | 0.25 | 0.2 | 0 |
| 24 | 2 SM pisc | 0.034 |  | 1.33 |  | 0.95 | 0.25 | 0.2 | 0 |
| 25 | 2 Crustaceans | 0.034 | 3.98 | 2 |  |  | 0.15 | 0.2 | 0 |
| 26 | 2 Macrobenthos | 0.207 | 7.96 | 2 |  |  | 0.15 | 0.2 | 0 |
| 27 | 2 Meiobenthos | 0.207 | 8.07 | 9 |  |  | 0.15 | 0.2 | 0 |
| 28 | 2 Zooplankton | 0.207 | 11 | 24 | 112 |  |  | 0.2 | 0 |
| 29 | 2,3 Hilsa | 0.146 | 1.75 | 1.76 |  |  | 0.25 | 0.2 | 0 |
| 30 | 2,3 Indian mackerel | 0.146 | 0.43 | 2.29 |  |  | 0.2 | 0.2 | 0 |
| 31 | 3 L pisc | 0.111 |  | 0.76 |  | 0.9 | 0.2 | 0.2 | 0 |
| 32 | 3 Carangids | 0.111 | 0.348 | 1.24 |  |  | 0.2 | 0.2 | 0 |
| 33 | 3 S pelagics | 0.111 | 0.643 | 2.29 |  |  | 0.25 | 0.2 | 0 |
| 34 | 3 L pisc comm | 0.111 |  | 0.89 |  | 0.9 | 0.2 | 0.2 | 0 |
| 35 | 3 Milkfish plus | 0.111 |  | 0.38 |  | 0.9 | 0.2 | 0.2 | 0 |
| 36 | 3 SM inv | 0.111 | 0.625 | 1.57 |  |  | 0.25 | 0.2 | 0 |
| 37 | 3 SM pisc | 0.111 |  | 1.33 |  | 0.95 | 0.25 | 0.2 | 0 |
| 38 | 3 Crustaceans | 0.335 | 1.854 | 2 |  |  | 0.15 | 0.2 | 0 |
| 39 | 3 Macrobenthos | 0.335 | 5.563 | 2 |  |  | 0.15 | 0.2 | 0 |
| 40 | 3 Meiobenthos | 0.335 | 4.57 | 9 |  |  | 0.15 | 0.2 | 0 |
| 41 | 3 Zooplankton | 0.335 |  | 24 | 112 | 0.6 |  | 0.2 | 0 |
| 42 | 1 Phytoplankton | 0.459 | 12.1 | 100 |  |  |  | 0 | 0 |
| 43 | 2 Phytoplankton | 0.207 | 29.3 | 100 |  |  |  | 0 | 0 |
| 44 | 3 Phytoplankton | 0.335 | 21.4 | 100 |  |  |  | 0 | 0 |
| 45 | Benthic plants | 0.689 |  | 4.1 |  | 0.4 |  | 0 | 0 |
| 46 | Bigeye tuna | 1 |  | 0.33 |  | 0.9 | 0.2 | 0.2 | 0 |
| 47 | Yellowfin tuna | 1 |  | 0.34 |  | 0.9 | 0.2 | 0.2 | 0 |
| 48 | Marlins | 1 |  | 0.39 |  | 0.9 | 0.2 | 0.2 | 0 |
| 49 | Detritus | 1 | 137 |  |  | 0 |  | 0 | 0 |

Table 16. Parameters of the balanced model. The values in bold indicate parameters estimated by Ecopath.

|  | Group name | Trophic level | Fraction of habitat area | Biomass in habitat area ( $\mathrm{t} / \mathrm{km}^{2}$ ) | Biomass ( $\mathrm{t} / \mathrm{km}^{2}$ ) | P/B <br> (/year) | Q/B <br> (/year) | EE | P/Q |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Oceanic sharks | 4.22 | 1 | 0.202 | 0.202 | 0.249 | 1.66 | 0.582 | 0.15 |
| 2 | Coastal elasmobranch | 4.17 | 0.689 | 0.118 | 0.081 | 0.66 | 4.4 | 0.944 | 0.15 |
| 3 | Tuna-like | 4.55 | 1 | 0.15 | 0.150 | 0.46 | 2.3 | 0.894 | 0.2 |
| 4 | Coastal scombrids | 4.25 | 0.689 | 0.151 | 0.104 | 0.84 | 4.2 | 0.598 | 0.2 |
| 5 | Jellyfish | 3 | 1 | 0.5 | 0.500 | 3 | 10 | 0.066 | 0.3 |
| 6 | Cephalopods | 3.89 | 1 | 0.569 | 0.569 | 1.2 | 4 | 0.9 | 0.3 |
| 7 | S bathy | 3 | 0.849 | 3.28 | 2.785 | 2.14 | 8.56 | 0.132 | 0.25 |
| 8 | ML bathy | 3.51 | 0.849 | 0.270 | 0.230 | 0.37 | 1.85 | 0.9 | 0.2 |
| 9 | 1 L pisc | 4.08 | 0.005 | 1.780 | 0.009 | 0.76 | 3.8 | 0.9 | 0.2 |
| 10 | 1 Carangids | 3.96 | 0.005 | 2.2 | 0.011 | 0.78 | 3.9 | 0.968 | 0.2 |
| 11 | 1 S pelagics | 3.10 | 0.005 | 5.446 | 0.027 | 2.39 | 9.56 | 0.95 | 0.25 |
| 12 | 1 L pisc comm | 3.84 | 0.005 | 4.48 | 0.022 | 0.42 | 2.1 | 0.400 | 0.2 |
| 13 | 1 SM inv | 3.08 | 0.005 | 7.970 | 0.040 | 2 | 8 | 0.95 | 0.25 |
| 14 | 1 SM pisc | 3.60 | 0.005 | 5.723 | 0.029 | 1.33 | 5.32 | 0.95 | 0.25 |
| 15 | 1 Macrobenthos | 2.37 | 0.459 | 1.337 | 0.614 | 2 | 13.33 | 0.8 | 0.15 |
| 16 | 1 Meiobenthos | 2 | 0.459 | 0.502 | 0.230 | 9 | 60 | 0.8 | 0.15 |
| 17 | 1 Zooplankton | 2 | 0.459 | 1.822 | 0.836 | 24 | 112 | 0.6 | 0.21 |
| 18 | 2 L pisc | 4.10 | 0.034 | 4.431 | 0.151 | 0.76 | 3.8 | 0.9 | 0.2 |
| 19 | 2 Carangids | 3.98 | 0.034 | 5 | 0.170 | 0.76 | 3.8 | 0.920 | 0.2 |
| 20 | 2 S pelagics | 3.13 | 0.034 | 7.746 | 0.263 | 2.94 | 11.76 | 0.95 | 0.25 |
| 21 | 2 L pisc comm | 3.90 | 0.034 | 4.42 | 0.150 | 0.89 | 4.45 | 0.919 | 0.2 |
| 22 | 2 Milkfish plus | 3.50 | 0.034 | 4.104 | 0.140 | 0.38 | 1.9 | 0.95 | 0.2 |
| 23 | 2 SM inv | 3.11 | 0.034 | 12.127 | 0.412 | 2 | 8 | 0.95 | 0.25 |
| 24 | 2 SM pisc | 3.68 | 0.034 | 8.793 | 0.299 | 1.33 | 5.32 | 0.95 | 0.25 |
| 25 | 2 Crustaceans | 2.62 | 0.034 | 24 | 0.816 | 3 | 20 | 0.892 | 0.15 |
| 26 | 2 Macrobenthos | 2.34 | 0.207 | 12 | 2.484 | 2.5 | 16.67 | 0.847 | 0.15 |
| 27 | 2 Meiobenthos | 2 | 0.207 | 10 | 2.070 | 9 | 60 | 0.816 | 0.15 |
| 28 | 2 Zooplankton | 2 | 0.207 | 11 | 2.277 | 24 | 112 | 0.281 | 0.21 |
| 29 | 2,3 Hilsa | 2.36 | 0.146 | 2.4 | 0.350 | 1.76 | 7.04 | 0.935 | 0.25 |
| 30 | 2,3 Indian mackerel | 3.05 | 0.146 | 1.9 | 0.277 | 2.5 | 12.5 | 0.934 | 0.2 |
| 31 | 3 L pisc | 4.23 | 0.111 | 1.662 | 0.185 | 0.76 | 3.8 | 0.9 | 0.2 |
| 32 | 3 Carangids | 3.98 | 0.111 | 3.5 | 0.389 | 1.24 | 6.2 | 0.983 | 0.2 |
| 33 | 3 S pelagics | 3.15 | 0.111 | 5.6 | 0.622 | 2.29 | 9.16 | 0.925 | 0.25 |
| 34 | 3 L pisc comm | 3.94 | 0.111 | 1.472 | 0.163 | 0.89 | 4.45 | 0.9 | 0.2 |
| 35 | 3 Milkfish plus | 3.52 | 0.111 | 1.604 | 0.178 | 0.38 | 1.9 | 0.9 | 0.2 |
| 36 | 3 SM inv | 3.13 | 0.111 | 8 | 0.888 | 2 | 8 | 0.929 | 0.25 |
| 37 | 3 SM pisc | 3.75 | 0.111 | 6.120 | 0.679 | 1.33 | 5.32 | 0.95 | 0.25 |
| 38 | 3 Crustaceans | 2.63 | 0.335 | 9 | 3.015 | 3 | 20 | 0.948 | 0.15 |
| 39 | 3 Macrobenthos | 2.37 | 0.335 | 13 | 4.355 | 2 | 13.33 | 0.944 | 0.15 |
| 40 | 3 Meiobenthos | 2 | 0.335 | 13 | 4.355 | 9 | 60 | 0.928 | 0.15 |
| 41 | 3 Zooplankton | 2 | 0.335 | 6.971 | 2.335 | 24 | 112 | 0.6 | 0.21 |
| 42 | 1 Phytoplankton | 1 | 0.459 | 12.1 | 5.554 | 100 | 0 | 0.153 |  |
| 43 | 2 Phytoplankton | 1 | 0.207 | 29.3 | 6.065 | 100 | 0 | 0.389 |  |
| 44 | 3 Phytoplankton | 1 | 0.335 | 21.4 | 7.169 | 100 | 0 | 0.337 |  |
| 45 | Benthic plants | 1 | 0.689 | 1.139 | 0.784 | 4.1 | 0 | 0.4 |  |
| 46 | Bigeye tuna | 4.43 | 1 | 0.014 | 0.014 | 0.33 | 1.65 | 0.9 | 0.2 |
| 47 | Yellowfin tuna | 4.76 | 1 | 0.022 | 0.022 | 0.34 | 1.7 | 0.9 | 0.2 |
| 48 | Marlins | 4.67 | 1 | 0.003 | 0.003 | 0.39 | 1.95 | 0.9 | 0.2 |
| 49 | Detritus | 1 | 1 | 137 | 137 |  |  | 0.332 |  |

Table 17. Diet matrix from balanced model. Values in bold differ from the original diet by $0.1 \%$ or more.

| Prey | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 Oceanic shark | 0.0066 | 0.0295 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 Coastal shark ray | 0.0129 | 0.0581 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 Tuna-like | 0.0834 | 0.0018 | 0.0306 | 0 | 0 | 0 | 0 | 0.0019 | 0.0030 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 Coastal scombrids | 0.0111 | 0.0018 | 0.0490 | 0 | 0 | 0 | 0 | 0.0019 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 Jellyfish | 0.0100 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0125 | 0 | 0 | 0.0053 | 0 | 0 |
| 6 Cephalopods | 0.1605 | 0.0110 | 0.1124 | 0.0322 | 0 | 0.0100 | 0 | 0.0640 | 0.0666 | 0.0631 | 0.0002 | 0.0236 | 0.0026 | 0.0103 | 0 |
| 71 S bathy | 0.0379 | 0.0000 | 0.0561 | 0.0220 | 0 | 0.1100 | 0 | 0.0750 | 0.1260 | 0.0331 | 0 | 0.0874 | 0.0000 | 0.0650 | 0 |
| 81 ML bathy | 0.0658 | 0.0053 | 0.0246 | 0 | 0 | 0 | 0 | 0.0290 | 0.0628 | 0.0000 | 0 | 0.0370 | 0 | 0 | 0 |
| 91 L pisc | 0.0027 | 0.0032 | 0.0005 | 0.0003 | 0 | 0 | 0 | 0.0030 | 0.0150 | 0.0152 | 0 | 0.0121 | 0 | 0.0013 | 0 |
| 101 Carangids | 0.0007 | 0.0008 | 0.0020 | 0.002 | 0 | 0.0010 | 0 | 0 | 0.0200 | 0.0300 | 0.0001 | 0.02 | 0 | 0.0001 | 0 |
| 1115 pelagics | 0.0012 | 0.0010 | 0.0083 | 0.0093 | 0 | 0.0019 | 0 | 0.0066 | 0.2702 | 0.2290 | 0.0223 | 0.04 | 0.0240 | 0.0800 | 0 |
| 121 L pisc comm | 0.0006 | 0.0005 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0225 | 0.0110 | 0 | 0.012 | 0.0004 | 0.0045 | 0 |
| 131 SM inv | 0.0019 | 0.0038 | 0.0088 | 0.0360 | 0 | 0.0037 | 0 | 0.0053 | 0.1548 | 0.2371 | 0.0198 | 0.1603 | 0.0078 | 0.0869 | 0 |
| 141 SM pisc | 0.0030 | 0.0032 | 0.0019 | 0.0010 | 0 | 0.0037 | 0 | 0.0106 | 0.0496 | 0.0837 | 0.0123 | 0.1097 | 0.0010 | 0.0373 | 0 |
| 151 Macrobenthos | 0.0012 | 0.0000 | 0.0000 | 0.0038 | 0 | 0.0460 | 0 | 0.1961 | 0.1840 | 0.1737 | 0.1970 | 0.4634 | 0.5940 | 0.7013 | 0.05 |
| 161 Meiobenthos | 0.0004 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0004 | 0.0250 | 0 | 0.0503 | 0.0001 | 0.20 |
| 171 Zooplankton | 0.0024 | 0 | 0.0005 | 0.0004 | 0 | 0 | 0.4585 | 0.1509 | 0.0245 | 0.0900 | 0.6810 | 0.0345 | 0.1307 | 0.0123 | 0.10 |
| 182 L pisc | 0.0188 | 0.0222 | 0.0033 | 0.0023 | 0 | 0 | 0 | 0.0042 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 192 Carangids | 0.0050 | 0.0058 | 0.0582 | 0.0409 | 0 | 0.0010 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 202 S pelagics | 0.0082 | 0.0072 | 0.0569 | 0.0641 | 0 | 0.0200 | 0 | 0.0030 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 212 L pisc comm | 0.0041 | 0.0037 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 222 Mikfish plus | 0.0013 | 0.0010 | 0.0068 | 1.E-06 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 232 SM inv | 0.0128 | 0.0261 | 0.0060 | 0.0300 | 0 | 0.0304 | 0 | 0.002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 242 SM pisc | 0.0210 | 0.0224 | 0.0133 | 0.0071 | 0 | 0.0254 | 0 | 0.0048 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 252 Crustaceans | 5E-05 | 0.0646 | 0.0000 | 0.0198 | 0 | 0.0570 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 262 Macrobenthos | 0.0080 | 0.0233 | 0 | 0.0064 | 0 | 0.05 | 0 | 0.0884 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 272 Meiobenthos | 0.0029 | 0 | 0 | 0 | 0 | 0.0104 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 282 Zooplankton | 0.0166 | 0 | 0.0038 | 0.0002 | 0 | 0 | 0.2067 | 0.0680 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 29 2,3 Hilsa | 0.0413 | 0.0208 | 0.0226 | 0.0168 | 0 | 0.0100 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 30 2,3 Indian mackerel | 0.0290 | 0.0226 | 0.0327 | 0.0619 | 0 | 0.0251 | 0 | 0.0019 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 313 L pisc | 0.0607 | 0.0718 | 0.0106 | 0.0075 | 0 | 0 | 0 | 0.0068 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 323 Carangids | 0.0163 | 0.0188 | 0.1884 | 0.1324 | 0 | 0.0200 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 333 S pelagics | 0.0264 | 0.0232 | 0.1843 | 0.2075 | 0 | 0.0350 | 0 | 0.0049 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 343 L pisc comm | 0.0133 | 0.0121 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 353 Milkfish plus | 0.0041 | 0.0034 | 0.0221 | 5E-06 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 363 SM inv | 0.0415 | 0.0844 | 0.0403 | 0.1876 | 0 | 0.0821 | 0 | 0.0039 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 373 SM pisc | 0.0679 | 0.0725 | 0.0432 | 0.0229 | 0 | 0.0821 | 0 | 0.0078 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 383 Crustaceans | 1E-04 | 0.2091 | 0.0000 | 0.0640 | 0 | 0.3483 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 393 Macrobenthos | 0.0257 | 0.0755 | 0 | 0.0207 | 0 | 0.0268 | 0 | 0.1432 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 403 Meiobenthos | 0.0092 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 413 Zooplankton | 0.0538 | 0 | 0.0122 | 0.0007 | 1 | 0 | 0.3348 | 0.1101 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 421 Phytoplankton | 0.0021 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0117 | 0.0340 | 0 | 0.0345 | 0 | 0.10 |
| 432 Phytoplankton | 0.0144 | 0 | 0.0008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 443 Phytoplankton | 0.0466 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 45 Benthic plants | 0.0016 | 0 | 0.0002 | 0 | 0 | 0 | 0 | 0 | 0.0010 | 0 | 0 | 0.0003 | 0.1161 | 0.0001 | 0 |
| 46 Bigeye tuna | 0.0000 | 0 | 0.0000 | 0 | 0 | 0 | 0 | 0 | 0.0000 | 0 | 0 | 0 | 0.0000 | 0.0000 | 0 |
| 47 Yellowfin tuna | 0.0000 | 0 | 0.0000 | 0 | 0 | 0 | 0 | 0 | 0.0000 | 0 | 0 | 0 | 0.0000 | 0.0000 | 0 |
| 48 Marlins | 0.0000 | 0 | 0.0000 | 0 | 0 | 0 | 0 | 0 | 0.0000 | 0 | 0 | 0 | 0.0000 | 0.0000 | 0 |
| 49 Detritus | 0.0405 | 0.0086 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0092 | 0.0084 | 0 | 0.0327 | 0 | 0.55 |
| 50 Import | 0.0043 | 0.0798 | 0 | 0 | 0 | 0 | 0 | 0.0058 | 0 | 0 | 0.0001 | 0 | 0.0010 | 0.0002 | 0 |


| Prey \predator | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 Oceanic shark | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 Coastal shark ray | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 Tuna-like | 0 | 0 | 0.0030 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 Coastal scombrids | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 Jellyfish | 0 | 0 | 0 | 0.0125 | 0 | 0 | 0 | 0.0053 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 Cephalopods | 0 | 0 | 0.0666 | 0.0631 | 0.0002 | 0.0236 | 0.0206 | 0.0016 | 0.0103 | 0 | 0 | 0 | 0 | 0 |
| 71 S bathy | 0 | 0 | 0.0600 | 0.0340 | 0 | 0.0624 | 0 | 0 | 0.0290 | 0 | 0 | 0 | 0 | 0 |
| 81 ML bathy | 0 | 0 | 0.0190 | 0 | 0 | 0.0027 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 91 L pisc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 101 Carangids | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 111 S pelagics | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 121 L pisc comm | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 131 SM inv | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 141 SM pisc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 151 Macrobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 161 Meiobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 171 Zooplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 182 L pisc | 0 | 0 | 0.0150 | 0.0080 | 0 | 0.0121 | 0 | 0 | 0.0000 | 0 | 0 | 0 | 0 | 0 |
| 192 Carangids | 0 | 0 | 0.0100 | 0.0153 | 0.0032 | 0.0200 | 0 | 0.0019 | 0.0050 | 0 | 0 | 0 | 0 | 0 |
| 2025 pelagics | 0 | 0 | 0.2610 | 0.1579 | 0.0139 | 0.0606 | 0 | 0.0210 | 0.0770 | 0 | 0 | 0 | 0 | 0 |
| 212 L pisc comm | 0 | 0 | 0.0225 | 0.0110 | 0 | 0.0120 | 0.0260 | 0.0004 | 0.0045 | 0 | 0 | 0 | 0 | 0 |
| 222 Mikfish plus | 0 | 0 | 0.0329 | 0.0001 | 0 | 0.0128 | 0 | 0 | 0.0027 | 0 | 0 | 0 | 0 | 0 |
| 232 SM inv | 0 | 0 | 0.1719 | 0.2370 | 0.0188 | 0.1476 | 0.0260 | 0.0078 | 0.0942 | 0 | 0 | 0 | 0 | 0 |
| 242 SM pisc | 0 | 0 | 0.0600 | 0.0837 | 0.0123 | 0.1097 | 0 | 0.0010 | 0.0373 | 0 | 0 | 0 | 0 | 0 |
| 252 Crustaceans | 0 | 0 | 0.1610 | 0.1183 | 0.1000 | 0.2796 | 0.2440 | 0.1500 | 0.3500 | 0.01 | 0 | 0 | 0 | 0 |
| 262 Macrobenthos | 0 | 0 | 0.0229 | 0.0554 | 0.0940 | 0.1837 | 0.6571 | 0.4440 | 0.3500 | 0.04 | 0.03 | 0 | 0 | 0 |
| 272 Meiobenthos | 0 | 0 | 0.0000 | 0.0004 | 0.0101 | 0 | 0 | 0.0473 | 0.0001 | 0.40 | 0.20 | 0 | 0 | 0 |
| 282 Zooplankton | 0 | 0 | 0.0245 | 0.0533 | 0.6810 | 0.0345 | 0.0003 | 0.1307 | 0.0123 | 0.15 | 0.10 | 0 | 0 | 0.2928 |
| 29 2,3 Hilsa | 0 | 0 | 0.0395 | 0.0576 | 0.0155 | 0.0127 | 0.0260 | 0.0025 | 0.0172 | 0 | 0 | 0 | 0 | 0 |
| 30 2,3 Indian mackerel | 0 | 0 | 0.0292 | 0.0711 | 0.0084 | 0.0253 | 0 | 0.0020 | 0.0100 | 0 | 0 | 0 | 0 | 0 |
| 313 L pisc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 323 Carangids | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 333 S pelagics | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 343 L pisc comm | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 353 Milkfish plus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 363 SM inv | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 373 SM pisc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 383 Crustaceans | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 393 Macrobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 403 Meiobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 413 Zooplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0046 |
| 421 Phytoplankton | 0 | 0.9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 432 Phytoplankton | 0 | 0 | 0 | 0.0117 | 0.0340 | 0 | 0 | 0.0345 | 0 | 0 | 0.12 | 0 | 0.90 | 0.4974 |
| 443 Phytoplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0079 |
| 45 Benthic plants | 0 | 0 | 0.0010 | 0 | 0 | 0.0003 | 0 | 0.1161 | 0.0001 | 0 | 0 | 0 | 0 | 0.0160 |
| 46 Bigeye tuna | 0 | 0 | 0.0000 | 0 | 0 | 0.0000 | 0 | 0.0000 | 0.0000 | 0 | 0 | 0 | 0 | 0.0000 |
| 47 Yellowfin tuna | 0 | 0 | 0.0000 | 0 | 0 | 0.0000 | 0 | 0.0000 | 0.0000 | 0 | 0 | 0 | 0 | 0.0000 |
| 48 Marlins | 0 | 0 | 0.0000 | 0 | 0 | 0.0000 | 0 | 0.0000 | 0.0000 | 0 | 0 | 0 | 0 | 0.0000 |
| 49 Detritus | 1 | 0.1 | 0 | 0.0092 | 0.0084 | 0 | 0 | 0.0327 | 0 | 0.40 | 0.55 | 1 | 0.10 | 0 |
| 50 Import | 0 | 0 | 0 | 0 | 0.0001 | 0 | 0 | 0.0010 | 0.0002 | 0 | 0 | 0 | 0 | 0.1814 |


|  | Prey \ predator | 30 | 31 | 32 | 33 | 34 | 35 | 36 | 37 | 38 | 39 | 40 | 41 | 46 | 47 | 48 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Oceanic shark | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | Coastal shark ray | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | Tuna-like | 0 | 0.0030 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0216 | 0 | 0.0313 |
| 4 | Coastal scombrids | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0216 | 0 | 0.07443 |
| 5 | Jellyfish | 0 | 0 | 0.0125 | 0 | 0 | 0 | 0.0053 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | Cephalopods | 0 | 0.0666 | 0.0631 | 0.0002 | 0.0236 | 0.0206 | 0.0016 | 0.0103 | 0 | 0 | 0 | 0 | 0.061 | 0.7741 | 0.19763 |
| 7 | 1 S bathy | 0 | 0.0147 | 0.0760 | 0 | 0.0414 | 0 | 0 | 0.0200 | 0 | 0 | 0 | 0 | 0.1946 | 0.0152 | 0 |
| 8 | 1 ML bathy | 0 | 0.0028 | 0 | 0 | 0.0027 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.3243 | 0 | 0.0588 |
| 9 | 1 L pisc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0021 | 0 | 0.0024 |
| 10 | 1 Carangids | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0021 | 0.0005 | 0.0040 |
| 11 | 1 S pelagics | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0021 | 0.0005 | 0.0075 |
| 12 | 1 L pisc comm | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 1 SM inv | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0021 | 0.0020 | 0.0021 |
| 14 | 1 SM pisc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0011 | 0.0012 | 0.0016 |
| 15 | 1 Macrobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0011 | 0 | 0.0010 |
| 16 | 1 Meiobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 17 | 1 Zooplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 18 | 2 L pisc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0148 | 0 | 0.0170 |
| 19 | 2 Carangids | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0148 | 0.0035 | 0.0277 |
| 20 | 2 S pelagics | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0148 | 0.0035 | 0.0514 |
| 21 | 2 L pisc comm | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 22 | 2 Mikfish plus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0131 | 0 |
| 23 | 2 SM inv | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0148 | 0.0136 | 0.0143 |
| 24 | 2 SM pisc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0074 | 0.0080 | 0.0107 |
| 25 | 2 Crustaceans | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0071 | 0 | 0.0071 |
| 26 | 2 Macrobenthos | 0.1509 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27 | 2 Meiobenthos | 0.0541 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 28 | 2 Zooplankton | 0.1260 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 29 | 2,3 Hilsa | 0.0100 | 0.0395 | 0.0300 | 0.0155 | 0.0127 | 0.0260 | 0.0025 | 0.0172 | 0 | 0 | 0 | 0 | 0.0324 | 0.0152 | 0 |
| 30 | 2,3 Indian mackerel | 0.0277 | 0.0292 | 0.0500 | 0.0084 | 0.0253 | 0 | 0.0020 | 0.0110 | 0 | 0 | 0 | 0 | 0.0216 | 0.0152 | 0.0450 |
| 31 | 3 L pisc | 0 | 0.0200 | 0.0050 | 0 | 0.0121 | 0 | 0 | 0.0013 | 0 | 0 | 0 | 0 | 0.0479 | 0 | 0.0545 |
| 32 | 3 Carangids | 0 | 0.1352 | 0.0153 | 0.0020 | 0.0414 | 0 | 0.0019 | 0.0200 | 0 | 0 | 0 | 0 | 0.0479 | 0.0112 | 0.0896 |
| 33 | 3 S pelagics | 0.018 | 0.2410 | 0.1200 | 0.0139 | 0.0606 | 0 | 0.0180 | 0.0670 | 0 | 0 | 0 | 0 | 0.0479 | 0.0112 | 0.1663 |
| 34 | 3 L pisc comm | 0 | 0.0225 | 0.0110 | 0 | 0.0120 | 0.0260 | 0.0004 | 0.0045 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 35 | 3 Milkfish plus | 0 | 0.0329 | 0.0001 | 0 | 0.0128 | 0 | 0 | 0.0027 | 0 | 0 | 0 | 0 | 0 | 0.0424 | 0 |
| 36 | 3 SM inv | 0 | 0.1220 | 0.2370 | 0.0188 | 0.1476 | 0.0260 | 0.0078 | 0.0942 | 0 | 0 | 0 |  | 0.04798 | 0.0434 | 0.0462 |
| 37 | 3 SM pisc | 0 | 0.0616 | 0.0937 | 0.0123 | 0.1097 | 0 | 0.0010 | 0.0373 | 0 | 0 | 0 | 0 | 0.0240 | 0.0258 | 0.0346 |
| 38 | 3 Crustaceans | 0 | 0.1610 | 0.1183 | 0.1609 | 0.2796 | 0.2440 | 0.1701 | 0.4933 | 0.05 | 0 | 0 | 0 | 0.0229 | 0 | 0.0231 |
| 39 | 3 Macrobenthos | 0.1976 | 0.0229 | 0.0554 | 0.0331 | 0.1837 | 0.6571 | 0.4239 | 0.2080 | 0 | 0.05 | 0 | 0 | 0 | 0 | 0 |
| 40 | 3 Meiobenthos | 0.0709 | 0 | 0.0004 | 0.0101 | 0 | 0 | 0.0473 | 0.0001 | 0.40 | 0.20 | 0 | 0 | 0 | 0 | 0 |
| 41 | 3 Zooplankton | 0.1650 | 0.0245 | 0.0910 | 0.6820 | 0.0345 | 0.0003 | 0.1340 | 0.0123 | 0.15 | 0.10 | 0 | 0 | 0 | 0 | 0 |
| 42 | 1 Phytoplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 43 | 2 Phytoplankton | 0.0541 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 44 | 3 Phytoplankton | 0.0709 | 0 | 0.0117 | 0.0340 | 0 | 0 | 0.0345 | 0 | 0 | 0.1 | 0 | 0.90 | 0 | 0 | 0 |
| 45 | Benthic plants | 0 | 0.0010 | 0 | 0 | 0.0003 | 0 | 0.1161 | 0.0001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 46 | Detritus | 0 | 0.0000 | 0 | 0 | 0.0000 | 0 | 0.0000 | 0.0000 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0156 |
| 47 | Import | 0 | 0.0000 | 0 | 0 | 0.0000 | 0 | 0.0000 | 0.0000 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0156 |

## Uncertainty and pedigree

As detailed above there is a large degree of uncertainty in the data supporting the model. It is possible to quantify the level of uncertainty by attributing level of confidence based on the source of the data and its precision (Christensen and Walters 2004). The pedigree routine describes the data source and assigning a coefficient of variation by comparing the general description of the data to a pre-defined table for each type of input parameters (biomass, $\mathrm{P} / \mathrm{B}, \mathrm{Q} / \mathrm{B}$, diet, catch). The rank (order from the top) of a category is named index. Each category is characterised by associated with an index value between 0 and 1 describing how well the parameter is rooted in local data and a confidence interval expressed as a percentage of the mean.

The pedigree assumes that locally derived data (e.g. field sampling, local diets) represent local conditions than data from elsewhere or values derived from empirical relationships or other models.
Specifying the pedigree of data used to build the Ecopath model is pertinent in the present study, as it:

- Provides a clear overview of how well the Bay of Bengal Ecopath model parameters are based on local, fieldbased data;
- Provides a basis for computing an overall index of model 'quality' using a scale ranging from 0 to 1 (a model has high quality when it is constructed mainly using precise estimates of various parameters, based on data from the system to be represented by the model); and
- Provides parameter ranges used for subsequent Monte Carlo uncertainty evaluation (in Ecosim).

This biomas scale was based on the observation that the functional groups' biomass is difficult to estimate accurately, and that there are different levels of uncertainty depending on the pedigree of data used (Table 18). This also applies to biomass estimates that are obtained from other models, where local conditions may be different.

Table 18. Model Pedigree definitions for biomass

| Index | Description | Index value | Confidence interval (\%) |
| ---: | :--- | :---: | :---: |
| 1 | Estimated by Ecopath | 0.0 | 80 |
| 2 | From other models | 0.0 | 80 |
| 3 | Based on professional judgement (guesstimate) | 0.0 | 80 |
| 4 | Approximate or indirect method | 0.4 | 50 |
| 5 | Sampling based, low precision | 0.7 | 30 |
| 6 | Sampling based, high precision | 1.0 | 10 |

Most functional groups in the ecosystem model were not sampled with precision or not sampled at all. Some biomass estimates were based on a few samples or on extensive sampling in a restricted area (e.g. zooplankton, meibenthos) and were assigned an index value of 0.7 (index=5). Estimates for large pelagics and several other groups were based on a combination of low precision sampling and of jugement on their spatial distribution and densities in deep waters and for this reason were considered as guestimates. The estimate for jellyfish being taken from the FAO 57 model was assigned an index of 2 . In several cases, missing estimates from the Indian coasts were approximated from contiguous territories which always decreased the index assigned to the biomass parameter. Although obtained from a stock assessment the biomass estimate for hilsa is given classified as obtained from an indirect method (index=4), that is a stock assessment model and its inherent uncertainty and the assumptions that had to be made to produce an estimate for 1987-1980.

The scale for $\mathbf{P} / \mathbf{B}$ and $\mathbf{Q} / \mathbf{B}$ ratios (Table 19) is based on the principle that these ratios are highly conservative parameters that are functions of species' size and population dynamics, i.e. characteristics for which there is ample information available (e.g., from empirical models or FishBase; (Froese and Pauly 2013).

In this model, $\mathrm{P} / \mathrm{B}$ values were obtained from the addition of M computed using empirical relationships based on growth information from local studies or other ecosystems, and an estimate of F which was often only a guesstimate. For this reason, the estimates for this parameter was classified as 4, empirical relationships (Table 19)

Table 19. Model Pedigree definitions for Production/Biomass and Consumption/Biomass ratios

| Index | Description | Index value | Confidence Interval (\%) |
| :---: | :--- | :---: | :---: |
| 1 | Estimated by Ecopath | 0.0 | 80 |
| 2 | Professional judgement (guesstimate) | 0.1 | 70 |
| 3 | From other models | 0.2 | 60 |
| 4 | Empirical relationships | 0.5 | 50 |
| 5 | Similar group/species, similar system, low precision | 0.6 | 40 |
| 6 | Similar group/species, same system, low precision | 0.7 | 30 |
| 7 | Same group/species, similar system | 0.8 | 20 |
| 8 | Same group/species, same system | 1.0 | 10 |

Species' diet compositions can be highly variable and thus locally observed diets tend to be more reliable than those derived from other systems and/or species groups. Pedigree definitions were modified from default values for diets (Table 20).

The large pelagics diets were mainly obtained from studies in other ecosystems and assumptions were made about the allocation of the diet in various areas thus, they were classified as 4 . The diet of several most functional groups was assigned an index of 5 because they were a compilation of studies carried out in various locations and varying in quality (food items to local quantitative studies; see Appendix A3). The diet of invertebrates (benthic and pelagics) are typically a compilation of general knowledge for several species and were attributed the lowest index.

Table 20. Model Pedigree definitions for diet compositions

| Index | Description | Index value | Confidence Interval (\%) |
| ---: | :--- | :---: | :---: |
| 1 | General knowledge of related group/species | 0.0 | 80 |
| 2 | From other models | 0.0 | 80 |
| 3 | Qualitative diet composition | 0.2 | 60 |
| 4 | all types of diet, different systems, + allocation | 0.5 | 60 |
| 5 | Quantitative but limited diet composition study | 0.7 | 50 |
| 6 | Quantitative, detailed, diet composition study | 1.0 | 30 |

All catches were given the highest score (6, Local study, High precision/complete) because it was based on extensive studies that included estimates of illegal, unreported and unregulated catches, completing the FAO data base (see the Catch, effort and CPUE section above). However, this may be an over-estimate of the precision given the assumptions that were made for spatial distributions and species compositions.

## Ecosim model

A preliminary Ecosim model was set up to allow exploring interactions between functional groups and the impact of fishing. To run the model, a CSV file (comma delimited) was assembled, containing:

1. Effort time series for each region and four specific functional groups: hilsa, yellowfin and bigeye tunas, and marlins (see the section Catch, effort and CPUE and Appendix 2) (the relative industrial effort for India including the industrial fleet and large trawlers and labeled large vessels in A2.2);
2. Relative biomass time series for hilsa (biomass from assessment) and large pelagics (commercial CPUE);
3. Catch time series for each functional group.

The rules to build this type of file are described in the user's guide. The model was fitted using the automatic fitting procedure and considering only the groups for which the sum of squares (SS) is sensitive to vulnerabilities (v) and excluding benthic invertebrates and bathypelagics.

The fit to catches and relative biomass indices are not very good for either hilsa or the 3 large pelagic groups (Figure 5 and 6). The biomass trend predicted for hilsa shows no sign of decline contrary to the biomass trend from the stock assessment. Catches are either overestimated (e.g. hilsa and marlins) or under estimated by the model (e.g. bigeye and yellowfin tunas). At first sight, the predicted biomass indices seem to match the observed values, but a closer examination shows that although this is true at the end of the simulations, the predicted bigeye biomass does not decrease as much in the 1980s as the observed values suggest. The problem with tunas in this kind of model, in addition to the uncertainty of commercial CPUEs, is the migratory nature of these fish and the changes in spatial allocation of fishing effort which is not captured here. In addition, several of these species were already declining before 1978, the onset of this model (R. Sharma, pers. comm).

The fit to catches is variable among the other functional groups (Figure 6). Predicted catches for sharks, and small fish (SM inv, SM pisc and Spelagics) do not fit the observed catches very well. The reasons for this are numerous. The effort trends are approximates for regions 1 and 2 and could lead to error in predicted catches. However, predicted catches are not necessarily worse in these regions than in region 2. Also, using the same effort for all species implies that relative abundances do not change over time and that there is no variation in which species are targeted by the fishery during the study period which may not be the case.

There are no biomass trends for most species which means there are no constraints in the model and no guidance for what is possible. Trends for 2 Lpisc constitute a good example of these types of problems. The biomass is predicted to decrease abruptly in the 1980s (Figure 7) while fishing mortality increases from about 0.5 in 1978 to reach close to 4 at the end of the time series while predicted catches become lower than the observed. This poses at least two questions beyond that of the validity of the initial biomass: 1. Was F equal to 0.5 in 1978; and 2. Has fishing mortality increased on these species as fast as effort did? These are difficult questions to answer without knowledge on fishing habits during the time period and about biomass trends.

Another example is how the predictions for hilsa biomass trends are dramatically different when the Indian industrial effort is slightly changed by considering only the industrial vessels the large vessels time series (A2.2). Appendix 6 shows a new series of plots similar to figures 5-7 for this alternative scenario. The predicted trend for hilsa biomass would change from a flat line when the large vessels time series is used (Figure 5) to a declining trend when only industrial vessels are considered (Figure A6.1); none of these simulations fit hilsa very well. The sudden change is caused by changes in trends of other functional groups trends in biomass (not shown, but see changes in the trends in catches in A6.2 compared to Figure 6) induced by the simulated fishery in region 2 leading to different estimations of vulnerabilities. Also note the small change in level of catch fo 2 L pisc in Figure 7 and Figure A6.3.

A word of caution: It should be kept in mind that, for instance, the effort time series in region 1 and 3 are approximates, and the abundance indices for large pelagics are based on commercial CPUEs that are often biased. Also, abundance trends are missing for most functional groups, a lack of constraints resulting in very variable trends (e.g. hilsa).

Thus, results from temporal simulations should not be used to give quantitative management advice. Instead, current Ecosim simulations could be used as a tool to explore food web dynamics and effects of fishing. It could also be used as a framework to devise what important piece of information would be useful to gather to answer the most crucial questions.


Figure 5. Observed (dots) and predicted (lines) relative abundance (B or CPUE) and catches (C) for hilsa, marlins, yellowfin tuna (YFT) and bigeye tuna (BET). The numbers besides the name is the sum of squares. The region 2 industrial fleet includes industrial vessels and large trawlers (Large vessels time series in A2.2).


Figure 6. Observed (dots) and predicted (lines) catches for other species. The numbers besides the name is the sum of squares. The region 2 industrial fleet includes industrial vessels and large trawlers (Large vessels time series in A2.2)


Figure 7. Predicted trends in biomass, catches, and mortality (lines) and observed catches (dots) for the functional group 2 L pisc. The solid green areas represent the catch or fishing mortality caused by the each of the fleet in region 2 . The region 2 industrial fleet includes industrial vessels and large trawlers (Large vessels time series in A2.2)

## Acknowledgements

The present work was funded by FAO through the Bay of Bengal Large Marine Ecosystem Project (www.boblme.org). Special thanks to Rudolf Hermes who coordinated this project and helped in various ways. Special thanks to Rishi Sharma for his advice and invaluable contribution to this work. Thank you to Krista Greer, Dirk Zeller, Danielle Knip, Villy Christensen, Duto Nugroho for data, articles, and helpful discussions.

## References

Abdussamad, E. M., Pillai, N. G. K., Mohamed Kasim, H., Habeeb Mohammed, O. M. M. J., and Jeyabalan, K. 2010. Fishery, biology and population characteristics of the Indian mackerel, Rastrelliger kanagurta (Cuvier) exploited along the Tuticorin coast. Indian J. Fish., 57:17-21.
Afsal, V. V., Yousuf, K., Anoop, B., Anoop, A. K., Kannan, P., Rajagopalan, M., and Vivekanandan, E. 2008. A note on cetacean distribution in the Indian EEZ and contiguous seas during 2003-07. J. Cetacean Res. Manage., 10(3):209-215.
Alder, J., Guénette, S., Beblow, J., Cheung, W., and Christensen, V., 2007. Ecosystem-based global fishing policy scenarios. University of British Columbia, Vancouver, BC Canada, Fisheries Centre Research Report, 15 (7). 91 pp.
Alias, M., 2003. Trophic model of the coastal fisheries ecosystem of the west coast of Peninsular Malaysia. In: Assessment, management and future directions for coastal fisheries in Asian countries. pp. 313-332, Edited by G. Silvestre, L. Garces, I. Stobutzki, M. Ahmed, R. A. Valmonte-Santos, C. Luna, L. Lachica-Alino, P. Munro, V. Christensen, and D. Pauly, WorldFish Center, Penang (Malaysia). Vol. 67.

Allen, R. R. 1971. Relation between production and biomass. J. Fish. Res. Bd. Canada, 28:1573-1581.
Amin, S. M. N., Rahman, M. A., Haldar, G. C., Mazid, M. A., and Milton, D. A. 2008. Catch per unit effort, exploitation level, and production of hilsa shad in Bangladesh waters. Asian Fish. Sci., 21:175-187.
Anderson, R. C., and Waheed, Z., 1999. Management of shark fisheries in the Maldives. Case studies of the management of elasmobranch fisheries, Edited. by R. Shotton, FAO, Rome
Anderson, R. C., Waheed, Z., Rasheed, M., and Arif, A., 1992. Reef fish resources survey in the Maldives - Phase II. Madras, India, BOBP/WP/80. 54 pp. ftp://ftp.fao.org/docrep/fao/007/ae459e/ae459e00.pdf
Ansari, Z. A., Furtado, R., Badesab, S., Mehta, P., and Thwin, S. 2012. Benthic macroinvertebrate community structure and distribution in the Ayeyarwady continental shelf, Andaman Sea. Indian journal of Geomarine sciences, 41(3):272-278.
Antony, P. J., Dhanya, S., Lyla, P. S., Kurup, B. M., and Ajmal Khan, S. 2010. Ecological role of stomatopods (mantis shrimps) and potential impacts of trawling in a marine ecosystem of the southeast coast of India. Ecol. Model., 221(21):2604-2614.
Arai, M., 1996. Carnivorous zooplankton, jellies and Velella in the Alaska Gyre. In: Mass-balance models of Northeastern Pacific ecosystems. pp. 18-19, Edited by D. Pauly and V. Christensen, Fisheries Centre, University of British Columbia, Vancouver, BC. Vol. 4 (1) 129 pp.
Aydin, K. Y., MacFarlane, G. A., King, J. R., and Megery, B. A., 2003. PICES-GLOBEC international program on climate change and carrying capacity; The BASS/MODEL report on trophic models of the subarctic Pacific Basin Ecosystems. North Pacific Marine Scienve Organization (PICES), Sidney, BC, PICES Scientific Report, 25. 93 pp. http://www.pices.in
Babcock, E. A., Pikitch, E. K., McAllister, M. K., Apostolaki, P., and Santora, C. 2005. A perspective on the use of spatialized indicators for ecosystem-based fishery management through spatial zoning. ICES J. Mar. Sci., 62:469-476.
Bhathal, B., 2005. Historical reconstruction of Indian marine fisheries catches, 1950-2000, as a basis for testing the "Marine Tophic Index". UBC Fishery Centre Vancouver, Canada, Fisheries Centre Research Report 13(5). 122 pp.

Bhathal, B., 2013. The governement-led development of India's marine fisheries since 1950: Catch and effort trends, and bioeconomic models for exploring to alternative policy. PhD thesis, University of British Columbia, Vancouver BC.
BOBLME, 2012. Report of the hilsa fisheries assessment working group II, 24-25 April 2012, Mumbai, India. Bay of Bengal Large Marine Ecosystem Project, BOBLME-2012-Ecology-10. 29 pp.
Brotz, L., Cheung, W. L., Kleisner, K., Pakhomov, E., and Pauly, D. 2012. Increasing jellyfish populations: trends in Large Marine Ecosystems. In: Springer Netherlands, Hydrobiologia, 690(1):3-20.
Chen, Z. Z., Qiu, Y. S., Jia, X. P., and Xu, S. N. 2008a. Using an Ecosystem Modeling Approach to Explore Possible Ecosystem Impacts of Fishing in the Beibu Gulf, Northern South China Sea. Ecosystems, 11(8):1318-1334.
Chen, Z. Z., Qiu, Y. S., Jia, X. P., and Xu, S. N. 2008b. Simulating fisheries management options for the Beibu Gulf by means of an ecological modelling optimization routine. Fish. Res., 89(3):257-265.
Chen, Z. Z., Xu, S. N., Qiu, Y. S., Lin, Z. J., and Jia, X. P. 2009. Modeling the effects of fishery management and marine protected areas on the Beibu Gulf using spatial ecosystem simulation. Fish. Res., 100(3):222-229.
Christensen, V. 1998. Fishery-induced changes in a marine ecosystem: insight from models of the Gulf of Thailand. J. Fish Biol., 53:128-142.

Christensen, V., and Pauly, D., 1992a. A guide to the Ecopath II software system (version 2.1). ICLARM Software 6. International Centre for Living Aquatic Resources Management, Manila, Philippines. 72 pp.
Christensen, V., and Pauly, D. 1992b. Ecopath II- a software for balancing steady-state ecosystem models and calculating network characteristics. Ecol. Model., 61:169-185.
Christensen, V., and Walters, C. J. 2004. Ecopath with Ecosim: methods, capabilities and limitations. Ecol. Model., 172:109-139.
Christensen, V., and Walters, C. J., 2005. Using ecosystem modelling for fisheries management: Where are we? ICES CM 2005/M:19 (updated).
Christensen, V., Walters, C. J., Ahrens, R., Alder, J., Buszowski, J., Christensen, L. B., Cheung, W. W. L., Dunne, J., Froese, R., Karpouzi, V., Kaschner, K., Kearney, K., Lai, S., Lam, V., Palomares, M. L. D., Peters-Mason, A., Piroddi, C., Sarmiento, J. L., Steenbeek, J., Sumaila, R., Watson, R., Zeller, D., and Pauly, D. 2009. Databasedriven models of the world's Large Marine Ecosystems. Ecol. Model., 220(17):1984-1996.
Darwin Reef Fish Project, 2011. Management plan for the Maldives grouper fishery. Darwin Reef Fish Project, Marine Research Centre, Maldives and Marine Conservation Society, UK, 29 pp. http://www.mcsuk.org/downloads/coral reefs/Maldives Grouper\%20 fishery Management Plan.pdf
Dutta, S., Maity, S., Bhattacharyya, S. B., Sundaray, J. K., and Hazra, S. 2013. Diet composition and intensity of feeding of Tenualosa ilisha (Hamilton, 1822) occuring in the Northern Bay of Bengal, India. Proc. Zool. Soc., DOI 10.1007/s12595-013-0066-3
FAO, 2001 Towards ecosystem-based fisheries management. In: The Reykjavik Conference on Responsible Fisheries in the Marine Ecosystem, 1-4 October 2001, p 11, Reykjavik, Iceland, FAO and Governement of Norway.
Froerman, Y. M. 1984. Feeding spectrum and trophic relationship of short-finned squid (Illex illecebrosus) in the Northwest Atlantic. NAFO Sci. Coun. Studies, 7:67-75.
Froese, R., and Pauly, D., (Editors). 2013. FishBase, World Wide Web electronic publication Version 08/2013. www.fishbase.org,
Garces, L. R., Man, A., Ahmad, A. T., Mohamad-Norizan, M., and Silvestre, G., 2003. A trophic model of the coastal fisheries ecosystems off the west coast of Sabah and Sarawak, Malaysia. In: Assessment, management and future directions for coastal fisheries in Asian countries. pp. 333-352, Edited by G. Silvestre, L. Garces, I. Stobutzki, M. Ahmed, R. A. Valmonte-Santos, C. Luna, L. Lachica-Alino, P. Munro, V. Christensen, and D. Pauly, WorldFish Center, Penang (Malaysia). Vol. 67.
Garcia, S., Sparre, P., and Csirke, J. 1989. Estimating surplus production and maximum yield from biomass data when catch and effort series are not available. Fish. Res., 8:13-23.
Gavaris, S. 2009. Fisheries management planning and support for strategic and tactical decisions in an ecosystem approach context. In: Ecosystem Approach to Fisheries: Improvements on Traditional Management for Declining and Depleted Stocks, Annual Meeting of the North Pacific Marine Science Organization, Fish. Res., 100:6-14.
Gerlach, S. A. 1971. On the importance of marine meiofauna for benthos communities. Oecologia, 6:176-190.

Gjøsaeter, J., 1978. Resource study of mesopelagic fish. PhD thesis thesis, University of Bergen, Bergen, Norway, 203 pp.
Gonçalves, J. M. 1991. The octopoda (Mollusca: Cephalopoda) of the Azores. Arquipel Life Mar Sci, 9:75-81. Guénette, S., 2005. Model of the Southeast Alaska. In: Foodweb models and data for studying fisheries and environmental impact on Eastern Pacific ecosystems. pp. 106-178, Edited by S. Guénette and V. Christensen, Fisheries Centre, University of British Columbia, Vancouver, BC, Canada. Vol. 13 (1).
Guénette, S., Meissa, B., and Gascuel, D. 2014. Assessing the contribution of marine protected areas to the trophic functioning of Ecosystems: A model for the Banc d'Arguin and the Mauritanian shelf PLoS ONE, 9(4):e94742.
Hanlon, R. T., and Messenger, J. B., 1996. Cephalopod behaviour. Cambridge University Press, Cambridge. 232 pp.
Hay, S. 2006. Marine ecology: Gelatinous bells may ring change in marine ecosystems. Curr. Biol., 16:R679-R682.
Heileman, S., Bianchi, G., and Funge-Smith, S., 2009. VII-10 Bay of Bengal: LME \#34. In: The UNEP large marine ecosystems: A perspective on changing conditions in LMEs of the world's regional seas. pp. 237-251, Edited by K. Sherman and G. Hempel, United Nations Environement Programme, Nairobi, Kenya. Vol. 82.
Hemmings, M., Harper, S., and Zeller, D., 2011. Reconstruction of total marine catches for the Maldives, 19502008. In: Fisheries catch reconstructions: Islands, Part II. Edited by S. Harper and D. Zeller, Fisheries Centre. University of British Columbia, Vancouver, Canada. Vol. 19(4).
Hilborn, R., Orensanz, J. M. L., and Parma, A. M. 2005. Institutions, incentives and the future of fisheries. Phil. Trans. R. Soc. B, 360:47-57.
Hoenig. 1983. Empirical use of longevity data to estimate mortality rates. Fish. Bull., 82:898-903.
Holmgren, S., 1994. An environmental assessment of the Bay of Bengal region. Bay of Bengal Programme, Madras, BOBP/REP/67. 240 pp.
Hong, J., He-Qin, C., Hai-Gen, X., Arreguin-Sanchez, F., Zetina-Rejón, M. J., Del Monte Luna, P., and Le Quesne, W. J. F. 2008. Trophic controls of jellyfish blooms and links with fisheries in the East China Sea. Ecol. Model., 212:492-503.
Hossain, M. M., 2004. On sustainable management of the Bay of Bengal Large Marine Ecosystem (BOBLME) National Report of Bangladesh, FAO: GCP/RAS/179/WBG. 121 pp. http://www.boblme.org/documentRepository/Nat Bangladesh.pdf
Jarre-Teichmann, A., and Guénette, S., 1996. Invertebrate benthos. In: Mass-balance models of North-eastern Pacific ecosystems. pp. 38-39, Edited by D. Pauly and V. Christensen, UBC Fish Centre Res Rep, Vancouver, BC. Vol. 4 (1).
Joseph, L., 1999. Management of shark fisheries in Sri Lanka. In: Case studies of the management of elasmobranch fisheries. Edited by R. Shotton, FAO Fisheries technical paper 378/1, Rome.
Kenchington, T. J. 2013. Natural mortality estimators for information-limited fisheries. Fish Fish., doi:10.1111/faf. 12027
Lam, V. W. Y., and Pauly, D. 2005. Mapping the global biomass of mesopelagic fishes. Sea Around Us Project Newsletter.July/August (30)
Lynam, C. P., Gibbons, M. J., Axelsen, B. E., Sparks, C. A. J., Coetzee, J., Heywood, B. G., and Brierley, A. S. 2006. Jellyfish overtake fish in a heavily fished ecosystem. Curr. Biol., 16(13):R492-R493.
Mahmood, N., Chowdhury, M. J. U., Hossain, M. M., Haider, S. M. B., and Chowdbury, S. R., 1994. Bangladesh. In: An environmental assessment of the Bay of Bengal region. pp. 75-94, Edited by S. Holmgren, Bay of Bengal Programme, BOBP/REP/67, Madras.
Mills, C. E. 2001. Jellyfish blooms: are populations increasing globally in response to changing ocean conditions? In: Springer Netherlands, Hydrobiologia, 451(1):55-68.
Milton, D. A., 2010. Status of hilsa (Tenualosa ilisha) management in the Bay of Bengal: An assessment of population risk and data gaps for more effective regional management. FAO Bay of Bengal Large Marine Ecosystem Project BOBLME-2010-Ecology-01. 67 pp.
Mondreti, R., Davidar, P., Péron, C., and Grémillet, D. 2013. Seabirds in the Bay of Bengal large marine ecosystem: Current knowledge and research objectives. Open J. Ecol., 3:172-184.
Mustafa, G., 2003. Trophic model of the coastal ecosystem in the waters of Bangladesh, Bay of Bengal. In: Assessment, amanagement and future directions for coastal fisheries in Asian countries. pp. 263-280, Edited by G. Silvestre, L. Garces, I. Stobutzki, M. Ahmed, R. A. Valmonte-Santos, C. Luna, L. Lachica-Aliño, P. Munro, V. Christensen, and D. Pauly, WorldFish Center Conference Proceedings, Vol. 67.

Noble, A., Gopakumar, G., Pillai, N. G., Kulkarni, G. M., Kurup, K. N., Reuben, S., Sivadas, M., and Yohannan, T. M. 1992. Assessment of mackerel stock along the Indian coast. Indian J. Fish., 39:119-124.

Nootmorn, P., Sumontha, M., Keereerut, P., Jayasingh, R., Jagannath, N., and Sinha, M., 2008. Stomach content of the large pelagic fishes in the bay of bengal. IOTC-2008-WPEB-11. 13 pp . http://www.iotc.org/files/proceedings/2008/wpeb/IOTC-2008-WPEB-11.pdf
Palomares, M. L. D., and Pauly, D. 1998. Predicting food consumption of fish populations as functions of mortality, food type, morphometrics, temperature and salinity. Mar. Freshwater Res., 49(5):447-453.
Panjarat, S., 1999. Preliminary study on the stomach content of yellowfin tuna in the Andaman Sea. In: Preliminary results on the large pelagic fisheries resources survey in the Andaman Sea. pp. 114-122, Southeast Asian Fisheries Development Center, Vol. TD/RES 99.
Pauly, D. 1980. On the interrelationships between natural mortality, growth parameters, and mean environmental temperature in 175 fish stocks. J. Cons. Int. Explor. Mer, 39:175-192.
Pauly, D., and Christensen, V. 1995. Primary production required to sustain global fisheries. Nature, 374:255-257.
Pierce, G. J., Boyle, P. R., Hastie, L. C., and Key, L. 1994. The life history of Loligo forbesi (Cephalopoda: Loliginidae) in Scottish waters. Fish. Res., 21:17-41.
Pikitch, E. K., Santora, C., Babcock, E. A., Bakun, A., Bonfil, R., Conover, D. O., Dayton, P., Doukakis, P., Fluharty, D., Heneman, B., Houde, E. D., Link, J., Livingston, P. A., Mangel, M., McAllister, M. K., Pope, J., and Sainsbury, K. J. 2004. Ecosystem-Based Fishery Management. Science, 305(5682):346-347.

Prasanna Kumar, S., Sardesai, S., Ramaiah, N., Bhosle, N. B., Ramaswamy, V., Ramesh, R., Sharada, S., M.M. , Sarupriya, J. S., and Muraleedharan, U., 2007. Bay of Bengal process studies (BOBPS) Final Report. Final Report Submitted to the Department of Ocean Development New Delhi, 142 pp. http://drs.nio.org/drs/handle/2264/535
Rashed-Un-Nabi, M., and Hadayet Ullah, M. 2012. Effects of Set Bagnet fisheries on the shallow coastal ecosystem of the Bay of Bengal. Ocean \& Coastal Management, 67:75-86.
Rikhter, V. A., and Efanov, V. N. 1976. On one of the approaches to estimation of natural mortality on fish populations. ICNAF Research Documents, 76(VI/8):1-12.
Robert, M., Faraj, A., McAllister, M. K., and Rivot, E. 2010. Bayesian state-space modelling of the De Lury depletion model: strengths and limitations of the method, and application to the Moroccan octopus fishery. ICES J. Mar. Sci., 67(6):1272-1290.
Rost Martins, H. 1982. Biological studies of the exploited stock of Loligo forbesi (Mollusca: Cephalopoda) in the Azores. J. Mar. Biol. Assoc. U.K., 62:799-808.
Sætersdal, G., Bianchi, G., Strømme, T., and Venema, S. C., 1999. The DR. FRIDTJOF NANSEN Programme 19751993. Investigations of fishery resources in developing countries. History of the programme and review of results. FAO Fish. Tech. Pap. 391, 434 pp.
Samarayanke, R. A. D. B., 2003. Review of national fisheries situation in Sri Lanka. In: Assessment, management and future directions for coastal fisheries in Asian countries. pp. 987-1012, Edited by G. Silvestre, L. Garces, I. Stobutzki, M. Ahmed, R. A. Valmonte-Santos, C. Luna, L. Lachica-Alino, P. Munro, V. Christensen, and D. Pauly, WorldFish Center, Penang (Malaysia). Vol. 67.
Smith, B. D., Ahmed, B., Mowgli, R. M., and Strindberg, S. 2008. Species occurrence and distributional ecology of nearshore cetaceans in the Bay of Bengal, Bangladesh, with abundance estimates for Irrawaddy dolphins Orcaella brevirostris and finless porpoises Neophocaena phocaenoides. J. Cetacean Res. Manage., 10(1):45-58.
Strømme, T., 1983. Reports on surveys with the R/V Dr Fridtjof Nansen. Institute of Marine Research of Bergen, Bergen, UNDP/FAO Programme GLO/82/001. 29 pp. ftp://ftp.fao.org/docrep/nonfao/fns/fn134e.pdf
Suman, A., Wudianto, and Bintoro, G. 2006. Species composition, distribution, and potential yield of deep sea shrimp resources in the western Sumatera of the Indian Ocean EEZ of Indonesian waters. Ind. Fish. Res. J., 12(2):159-167.
Ullah, M. H., Rashed-Un-Nabi, M., and Al-Mamun, M. A. 2012. Trophic model of the coastal ecosystem of the Bay of Bengal using mass balance Ecopath model. Ecol. Model., 225(0):82-94.
Vermaat, J. E., Beijer, J. A. J., Gijlstra, R., Hootsmans, M. J. M., Philippart, C. J. M., and van den Brink, N. W. 1993. Leaf dynamics and standing stocks of intertidal Zostera noltii Hornem. and Cymodocea nodosa (Ucria) Ascherson on the Banc d'Arguin (Mauritania). In: Ecological studies in the coastal waters of Mauritania, Leiden, The Netherlands, Hydrobiologia, 258:59-72.

Vibunpant, S., Khongchai, N., Seng-eid, J., Eimsa-ard, M., and Supongpan, M., 2003. Trophic model of the coastal fisheries ecosystem in the Gulf of Thailand. In: Assessment, amanagement and future directions for coastal fisheries in Asian countries. pp. 365-386, Edited by G. Silvestre, L. Garces, I. Stobutzki, M. Ahmed, R. A. Valmonte-Santos, C. Luna, L. Lachica-Aliño, P. Munro, V. Christensen, and D. Pauly, WorldFish Center Conference Proceedings, Vol. 67.
Vovk, A. N. 1985. Feeding spectrum of longfin squid (Loligo pealeis) in the Northwest Atlantic and its position in the ecosystem. NAFO Sci. Coun. Studies, 8:33-38.
Walters, C., and Kitchell, J. F. 2001. Cultivation/depensation effects on juvenile survival and recruitment: implications for the theory of fishing. Can. J. Fish. Aquat. Sci., 58:39-50.
Walters, C. J., Christensen, V., and Pauly, D. 1997. Structuring dynamic models of exploited ecosystems from trophic mass-balance assessments. Rev. Fish Biol. Fish., 7:139-172.
Zeller, D., Knip, D. M., Zylich, K., and Pauly, D., 2013. Reconstructed total fisheries catches for the countries of the Bay of Bengal Large Marine Ecosystem: 1950-2010. Report to the Bay of Bengal Large Marine Ecosystem Project (www.boblme.org) Prepared by the Sea Around Us, Fisheries Centre, Vancouver BC, Canada, 346 pp.
Zischke, M. T., Griffiths, S. P., and Tibbetts, I. R. 2013. Rapid growth of wahoo (Acanthocybium solandri) in the Coral Sea, based on length-at-age estimates using annual and daily increments on sagittal otoliths. ICES J. Mar. Sci., 70(6):1128-1139.

## Appendices

## A1 Functional groups.

A1.1 Allocation of catch species to functional groups.

| Catch Name | Functional group | Catch Name | Functional group |
| :---: | :---: | :---: | :---: |
| Abalistes stellaris | SM pisc | Auxis spp. | coastal scombrid |
| Acanthocybium solandri | tuna-like | Auxis thazard | coastal scombrid |
| Acanthopagrus latus | SM inv | Auxis thazard thazard | coastal scombrid |
| Acanthuridae | SM inv | Batoidea | oceanic, coastal sharks |
| Acanthurus | SM inv | Belonidae | L pisc |
| Acanthurus lineatus | SM inv | Billfishes | tuna-like |
| Acetes | shrimps | Bivalvia | macrobenthos |
| Aethaloperca rogaa | L pisc comm | Bohadschia marmorata | macrobenthos |
| Alectis ciliaris | Carangids | Bothidae | SM inv |
| Alectis indica | Carangids | Brachyura | Crabs |
| Alepes | Carangids | Bramidae | SM pisc |
| Alepisaurus ferox | ML bathy | Bregmaceros mcclellandi | S pelagics |
| Alopias | Oceanic shark | Caesio | SM inv |
| Alopias pelagicus | oceanic sharks | Caesio caerulaurea | SM inv |
| Alopias spp. | Oceanic shark | Caesio lunaris | SM inv |
| Alopias superciliosus | oceanic sharks | Caesionidae | SM inv |
| Alopias vulpinus | oceanic sharks | Carangidae | Carangids |
| Aluterus | SM inv | Carangoides | Carangids |
| Ambassidae | $S$ pisc inv | Carangoides coeruleopinnatus | Carangids |
| Amblygaster sirm | $S$ pelagics | Carangoides ferdau | Carangids |
| Anchoviella | $S$ pelagics | Carangoides malabaricus | Carangids |
| Anguilla | L pisc | Carangoides orthogrammus | Carangids |
| Anguilliformes | L pisc | Caranx | Carangids |
| Anodontostoma chacunda | $S$ pelagics | Caranx hippos | Carangids |
| Aphareus rutilans | L pisc comm | Caranx ignobilis | Carangids |
| Apogonidae | SM inv | Caranx lugubris | Carangids |
| Aprion virescens | L pisc comm | Caranx melampygus | Carangids |
| Aquatic invertebrates | macrobenthos | Caranx sexfasciatus | Carangids |
| Arcidae | macrobenthos | Carcharhinidae | coastal elasmobranch ${ }^{\text {a }}$ |
| Ariidae | Milkfish, Minv Mpisc | Carcharhinidae | oceanic, coastal sharks |
| Ariomma indicum | SM pisc | Carcharhinus | coastal elasmobranch ${ }^{\text {a }}$ |
| Arius | SM inv | Carcharhinus | oceanic, coastal sharks |
| Atherinomorus lacunosus | SM inv | Carcharhinus albimarginatus | coastal elasmobranch |
| Auxis | coastal scombrid | Carcharhinus amblyrhynchos | coastal elasmobranch |
| Auxis rochei | coastal scombrid | Carcharhinus falciformis | oceanic sharks |
| Auxis rochei rochei | coastal scombrid | Carcharhinus limbatus | coastal elasmobranch |
| Carcharhinus melanopterus | coastal elasmobranch | Carcharhinus longimanus | oceanic sharks |


| Catch Name | Functional group | Catch Name | Functional group |
| :---: | :---: | :---: | :---: |
| Carcharhinus obscurus | coastal elasmobranch | Elagatis bipinnulata | Carangids |
| Carcharhinus sorrah | coastal elasmobranch | Elasmobranchii | oceanic, coastal sharks |
| Centrophorus granulosus | oceanic sharks | Eleutheronema tetradactylum | L pisc |
| Centropomidae | L pisc | Encrasicholina heteroloba | S pelagics |
| Cephalopholis argus | L pisc comm | Engraulidae | $S$ pelagics |
| Cephalopholis boenak | L pisc comm | Ephippidae | $M$ inv, L pisc |
| Cephalopholis miniata | L pisc comm | Epinephelus | L pisc comm |
| Cephalopoda | cephalopods | Epinephelus fuscoguttatus | L pisc comm |
| Cephea | jellyfish | Epinephelus polyphekadion | L pisc comm |
| Chanos chanos | Milkfish plus | Epinephelus tauvina | L pisc comm |
| Charybdis | Crabs | Euthynnus affinis | coastal scombrid |
| Chirocentrus | L pisc | Exocoetidae | $S$ pelagics |
| Chirocentrus dorab | L pisc | Fenneropenaeus indicus | shrimps |
| Chirocentrus nudus | L pisc | Fenneropenaeus merguiensis | shrimps |
| Clams or cockles and arkshells | macrobenthos | Fistulariidae | L pisc |
| Clupeidae | $S$ pelagics | Galeocerdo cuvier | coastal elasmobranch |
| Clupeiformes | $S$ pelagics | Gastropoda | macrobenthos |
| Clupeoids | $S$ pelagics | Gazza minuta | SM inv |
| Congresox talabonoides | L pisc | Gerreidae | SM inv |
| Congridae | I bathy, L pisc | Gerres | SM inv |
| Coryphaena | L pisc | Gnathanodon speciosus | Carangids |
| Coryphaena hippurus | L pisc | Gobiidae | S inv, M pisc |
| Crassostrea | macrobenthos | Gymnosarda unicolor | coastal scombrid |
| Crassostrea madrasensis | macrobenthos | Haemulidae | SM pisc |
| Cynoglossidae | SM coast inv | Harpadon nehereus | L pisc comm |
| Cynoglossus | SM coast inv | Harpago chiragra | macrobenthos |
| Dasyatidae | oceanic, coastal sharks | Hemiramphidae | $S$ pelagics |
| Dasyatis | coastal elasmobranch | Hemiramphus | S pelagics |
| Decapoda | crab, shrimp | Hexanchus griseus | oceanic sharks |
| Decapterus | Carangids | Hilsa kelee | Hilsa |
| Decapterus russelli | Carangids | Himantura | coastal elasmobranch |
| Diodon | Milkfish plus | Holocentridae | SM inv, pisc |
| Drepane | SM pisc | Holothuria atra | macrobenthos |
| Drepane punctata | SM pisc | Holothuria edulis | macrobenthos |
| Dussumieria | $S$ pelagics | Holothuriidae | macrobenthos |
| Dussumieria elopsoides | $S$ pelagics | Holothuroidea | macrobenthos |
| Hyporhamphus | $S$ pelagics | Homaridae and Palinuridae | Crabs |
| Ilisha elongata | Hilsa | Lutjanus gibbus | L pisc comm |
| Istiompax indica | tuna-like | Lutjanus johnii | L pisc comm |
| Istiophoridae | tuna-like | Lutjanus lutjanus | L pisc comm |
| Istiophorus | tuna-like | Lutjanus malabaricus | L pisc comm |
| Istiophorus platypterus | tuna-like | Macolor macularis | L pisc comm |


| Catch Name | Functional group | Catch Name | Functional group |
| :---: | :---: | :---: | :---: |
| Isurus | oceanic sharks | Macolor niger | L pisc comm |
| Isurus oxyrinchus | oceanic sharks | Makaira | tuna-like |
| Isurus paucus | oceanic sharks | Makaira indica | tuna-like |
| Isurus spp | oceanic sharks | Makaira mazara | Marlins |
| Kajikia audax | Marlins | Makaira nigricans | tuna-like |
| Katsuwonus pelamis | tuna-like | Marine fishes not identified | Miscellaneous fishes |
| Kawakawa | coastal scombrid | Marine pelagic fishes nei | Miscellaneous fishes |
| Labridae | L pisc | Marsupenaeus japonicus | shrimps |
| Lactarius lactarius | $S$ pelagics | Megalaspis cordyla | Carangids |
| Lambis lambis | macrobenthos | Megalops cyprinoides | L pisc |
| Lamma nasus | oceanic sharks | Melicertus latisulcatus | shrimps |
| Lamnidae | oceanic sharks | Meretrix | macrobenthos |
| Lamniformes | oceanic, coastal sharks | Metapenaeus | shrimps |
| Lates calcarifer | L pisc | Metapenaeus monoceros | shrimps |
| Latidae | L pisc | Miscellaneous aquatic invertebrates | macrobenthos |
| Leiognathidae | SM inv | Miscellaneous crustaceans | crab, shrimp |
| Leiognathus | SM inv | Miscellaneous fishes | Miscellaneous fishes |
| Lethrinidae | L pisc comm | Miscellaneous marine crustaceans | crab, shrimp |
| Lethrinus | L pisc comm | Miscellaneous marine molluscs | macrobenthos |
| Lethrinus harak | L pisc comm | Miscellaneous molluscs | macrobenthos |
| Lethrinus microdon | L pisc comm | Miscellaneous shrimps | shrimps |
| Lethrinus nebulosus | L pisc comm | Modiolus | macrobenthos |
| Lethrinus olivaceus | L pisc comm | Monacanthidae | SM inv |
| Lethrinus rubrioperculatus | L pisc comm | Mugil | SM inv |
| Lethrinus xanthochilus | L pisc comm | Mugilidae | M inv, S inv |
| Liza | SM inv | Mullidae | SM inv |
| Lobotes surinamensis | L pisc | Muraenesocidae | L pisc |
| Loliginidae | cephalopods | Muraenesox | L pisc |
| Loligo | cephalopods | Muraenesox cinereus | L pisc |
| Loligonidae | cephalopods | Myliobatidae | coastal elasmobranch |
| Lutjanidae | L pisc comm | Lutjanus bohar | L pisc comm |
| Lutjanus | L pisc comm | Nemipteridae | SM pisc |
| Lutjanus argentimaculatus | L pisc comm | Nemipterus | SM pisc |
| Netuma thalassina | Milkfish plus | Nemipterus japonicus | SM pisc |
| Octopoda | cephalopods | Plectropomus laevis | L pisc comm |
| Octopodidae | cephalopods | Plectropomus pessuliferus | L pisc comm |
| Octopus | cephalopods | Pleuronectidae | SM pisc |
| Octopus vulgaris | cephalopods | Pleuronectiformes | Pleuronectiformes |
| Otolithoides | L pisc comm | Plotosidae | L pisc |
| Palaemonidae | shrimps | Plotosus | L pisc |
| Palinuridae | Crabs | Polynemidae | L pisc, SM inv |


| Catch Name | Functional group | Catch Name | Functional group |
| :---: | :---: | :---: | :---: |
| Palinurus | Crabs | Polynemus | L pisc, SM inv |
| Pampus | SM inv | Pomacentridae | SM inv |
| Pampus argenteus | SM inv | Pomadasys | SM pisc |
| Pampus chinensis | SM inv | Pomadasys argenteus | SM pisc |
| Panulirus | Crabs | Portunidae | Crabs |
| Panulirus homarus | Crabs | Portunus | Crabs |
| Panulirus longipes | Crabs | Portunus pelagicus | Crabs |
| Panulirus penicillatus | Crabs | Priacanthus | SM pisc |
| Panulirus polyphagus | Crabs | Prionace glauca | oceanic sharks |
| Panulirus versicolor | Crabs | Pristipomoides | L pisc comm |
| Paphia | macrobenthos | Psettodes erumei | SM pisc |
| Parapenaeopsis | shrimps | Psettodidae | SM pisc |
| Parapenaeopsis hardwickii | shrimps | Pseudocarcharias kamoharai | oceanic sharks |
| Parapenaeopsis sculptilis | shrimps | Pseudorhombus | SM pisc |
| Parastromateus niger | Carangids | Pseudotolithus | L pisc comm |
| Pectinidae | macrobenthos | Pseudotriakis microdon | oceanic sharks |
| Pellona | $S$ pelagics | Rachycentron canadum | L pisc |
| Pellona ditchela | $S$ pelagics | Rajiformes | oceanic, coastal sharks |
| Penaeidae | shrimps | Rastrelliger | Indian mackerel |
| Penaeus | shrimps | Rastrelliger brachysoma | Indian mackerel |
| Penaeus monodon | shrimps | Rastrelliger kanagurta | Indian mackerel |
| Penaeus semisulcatus | shrimps | Rhincodon typus | oceanic sharks |
| Pennahia | L pisc comm | Rhinobatidae | coastal elasmobranch |
| Pennahia argentata | L pisc comm | Rhizostomatidae | jellyfish |
| Perciformes | Perciformes | Rhopilema | jellyfish |
| Perna viridis | macrobenthos | Rhynchobatus djiddensis | coastal elasmobranch |
| Pinctada margaritifera | macrobenthos | Saccostrea cuccullata | macrobenthos |
| Platycephalidae | L pisc | Sardinella | $S$ pelagics |
| Platycephalus indicus | L pisc | Sardinella fimbriata | $S$ pelagics |
| Plectorhinchus | SM pisc | Sardinella gibbosa | $S$ pelagics |
| Plectropomus areolatus | L pisc comm | Sardinella lemuru | $S$ pelagics |
| Saurida | L pisc comm | Sardinella longiceps | $S$ pelagics |
| Saurida tumbil | L pisc comm | Selachimorpha | oceanic, coastal sharks |
| Scarus | SM inv | Selachimorpha (Pleurotremata) | oceanic, coastal sharks |
| Scatophagus argus | SM inv | Selar boops | SM inv |
| Sciaenidae | L pisc comm | Selar crumenophthalmus | Carangids |
| Scolopsis | SM pisc | Selaroides leptolepis | SM inv |
| Scomber | Indian mackerel | Sepia | cephalopods |
| Scomberoides | Carangids | Sepiidae | cephalopods |
| Scomberoides commersonnianus | Carangids | Sepioteuthis lessoniana | cephalopods |
| Scomberoides lysan | Carangids | Sergestidae | shrimps |
| Scomberomorini | coastal scombrid | Seriola rivoliana | Carangids |


| Catch Name | Functional group | Catch Name | Functional group |
| :---: | :---: | :---: | :---: |
| Serranidae | L pisc comm | Seriolina nigrofasciata | Carangids |
| Sharks or rays and chimaeras | oceanic, coastal sharks | Thenus orientalis | Crabs |
| Shrimps and prawns | shrimps | Thryssa | $S$ pelagics |
| Shrimps and prawns | shrimps | Thunnini | tuna-like, scombrids ${ }^{\text {b }}$ |
| Siganidae | SM inv | Thunnus | tuna-like |
| Siganus | SM inv | Thunnus alalunga | tuna-like |
| Siganus canaliculatus | SM inv | Thunnus albacares | Yellowfin tuna |
| Sillaginidae | SM inv | Thunnus maccoyii | tuna-like |
| Sillago | SM inv | Thunnus obesus | Bigeye tuna |
| Sillago sihama | SM inv | Thunnus tonggol | coastal scombrid |
| Siluriformes | Milkfish, Minv Mpisc | Trachipterus | ML bathy |
| Soleidae | Sinv, M pisc | Triacanthidae | SM inv |
| Solenocera crassicornis | shrimps | Triaenodon obesus | coastal elasmobranch |
| Sparidae | SM inv | Trichiuridae | L pisc |
| Sphyraena | L pisc | Trichiurus | L pisc |
| Sphyraena jello | L pisc | Trichiurus lepturus | L pisc |
| Sphyraenidae | L pisc | Tridacna | macrobenthos |
| Sphyrna | oceanic sharks | Trochus | macrobenthos |
| Sphyrna lewini | oceanic sharks | Trochus niloticus | macrobenthos |
| Sphyrna mokarran | oceanic sharks | Turbo | macrobenthos |
| Sphyrna | oceanic sharks | Upeneus | SM inv |
| Sphyrna zygaena | oceanic sharks | Upeneus sulphureus | SM inv |
| Sphyrnidae | oceanic sharks | Upeneus vittatus | SM inv |
| Spratelloides delicatulus | $S$ pelagics | Uranoscopus | SM pisc |
| Spratelloides gracilis | S pelagics | Variola louti | L pisc comm |
| Squillidae | shrimps | Veneridae | macrobenthos |
| Stichopus | macrobenthos | Xiphias gladius | tuna-like |
| Stolephorus | $S$ pelagics | Xiphioidei | tuna-like |
| Stomatopoda | shrimps | Zenarchopterus dispar | $S$ pelagics |
| Stromateidae | SM inv |  |  |
| Synodontidae | L pisc comm |  |  |
| Tegillarca granosa | macrobenthos |  |  |
| Tenualosa ilisha | Hilsa |  |  |
| Tenualosa toli | Hilsa |  |  |
| Terapon | SM pisc |  |  |
| Terapon jarbua | SM pisc |  |  |
| Terapontidae | SM pisc |  |  |
| Tetraodontidae | milkfish, SM inv |  |  |
| Tetrapturus angustirostris | tuna-like |  |  |
| Tetrapturus audax | tuna-like |  |  |

[^2]A1.2 Area considered for each functional group.

| Group name | area (km ${ }^{\mathbf{2}}$ ) | proportion | explanation |
| :---: | :---: | :---: | :---: |
| Oceanic sharks | 6,205,051 | 1 | All regions |
| Coastal elasmobranch | 4,275,177 | 0.689 | All regions |
| Tuna-like | 6,205,051 | 1 | All regions |
| Coastal scombrid | 4,275,177 | 0.689 | EEZ without high seas |
| Jellyfish | 6,205,051 | 1 | All regions |
| Cephalopods | 6,205,051 | 1 | All regions |
| $S$ bathy | 5,268,883 | 0.849 | Deep waters only |
| ML bathy | 5,268,883 | 0.849 | Deep waters only |
| 1 L pisc | 30,998 | 0.005 | Shelf of region 1 |
| 1 Carangids | 30,998 | 0.005 | Shelf of region 1 |
| 1 S pelagics | 30,998 | 0.005 | Shelf of region 1 |
| 1 L pisc comm | 30,998 | 0.005 | Shelf of region 1 |
| 1 SM inv | 30,998 | 0.005 | Shelf of region 1 |
| 1 SM pisc | 30,998 | 0.005 | Shelf of region 1 |
| 1 Macrobenthos | 2,845,297 | 0.459 | EEZ of region 1 |
| 1 Meiobenthos | 2,845,297 | 0.459 | EEZ of region 1 |
| 1 Zooplankton | 2,845,297 | 0.459 | EEZ of region 1 |
| 2 L pisc | 213,663 | 0.034 | Shelf of region 2 |
| 2 Carangids | 213,663 | 0.034 | Shelf of region 2 |
| 2 S pelagics | 213,663 | 0.034 | Shelf of region 2 |
| 2 L pisc comm | 213,663 | 0.034 | Shelf of region 2 |
| 2 Milkfish plus | 213,663 | 0.034 | Shelf of region 2 |
| 2 SM inv | 213,663 | 0.034 | Shelf of region 2 |
| 2 SM pisc | 213,663 | 0.034 | Shelf of region 2 |
| 2 Crustaceans | 213,663 | 0.034 | Shelf of region 2 |
| 2 Macrobenthos | 1,282,459 | 0.207 | EEZ of region 2 |
| 2 Meiobenthos | 1,282,459 | 0.207 | EEZ of region 2 |
| 2 Zooplankton | 1,282,459 | 0.207 | EEZ of region 2 |
| 2,3 Hilsa | 905,170 | 0.146 | Shelf of regions 2 and 3 |
| 2,3 Indian mackerel | 905,170 | 0.146 | Shelf of regions 2 and 3 |
| 3 L pisc | 691,507 | 0.111 | Shelf of region 3 |
| 3 Carangids | 691,507 | 0.111 | Shelf of region 3 |
| 3 S pelagics | 691,507 | 0.111 | Shelf of region 3 |
| 3 L pisc comm | 691,507 | 0.111 | Shelf of region 3 |
| 3 Milkfish plus | 691,507 | 0.111 | Shelf of region 3 |
| 3 SM inv | 691,507 | 0.111 | Shelf of region 3 |
| 3 SM pisc | 691,507 | 0.111 | Shelf of region 3 |
| 3 Crustaceans | 2,077,295 | 0.335 | EEZ of region 3 |
| 3 Macrobenthos | 2,077,295 | 0.335 | EEZ of region 3 |
| 3 Meiobenthos | 2,077,295 | 0.335 | EEZ of region 3 |
| 3 Zooplankton | 2,077,295 | 0.335 | EEZ of region 3 |
| 1 Phytoplankton | 2,845,297 | 0.459 | EEZ of region 1 |
| 2 Phytoplankton | 1,282,459 | 0.207 | EEZ of region 2 |
| 3 Phytoplankton | 2,077,295 | 0.335 | EEZ of region 3 |
| 0 plants | 4,275,177 | 0.689 | EEZ without high seas |
| Bigeye tuna | 6,205,051 | 1 | All regions |
| Yellowfin tuna | 6,205,051 | 1 | All regions |
| Marlins | 6,205,051 | 1 | All regions |

## A1.3 Functional groups composition

| Functional group | Species list |
| :---: | :---: |
| Oceanic sharks | Isurus paucus, Isurus oxyrinchus, Alopias vulpinus, Alopias pelagicus, Prionace glauca, Carcharhinus longimanus, Sphyrna lewini, Sphyrna mokarran, Sphyrna zygaena, Carcharhinus falciformis, Alopias superciliosus, Hexanchus griseus, Rhincodon typus, Pteroplatytrygon violacea, Dalatias licha, Echinorhinus brucus, Centrophorus moluccensis Centrophorus uyato, Centroscyllium ornatum, Centrophorus niaukang, Centrophorus squamosus, Centrophorus tessellatus, Centrophorus granulosus, Pseudotriakis microdon, Pseudocarcharias kamoharai, Carcharhinus amblyrhynchoides |
| Coastal elasmobranch | Galeocerdo cuvier, Triaenodon obesus, Carcharhinus melanopterus, Carcharhinus limbatus, Carcharhinus dussumieri, Carcharhinus macloti, Carcharhinus albimarginatus, Rhizoprionodon acutus, Scoliodon laticaudus, Scoliodon walbeehmil, Carcharhinus sorrah, Mustelus manazo, Mustelus mosis, Carcharhinus albimarginatus, Carcharhinus altimus, Stegostoma fasciatum, Nebrius ferrugineus, Chiloscyllium indicum, Chiloscyllium griseum, Eusphyra blochii, Odontaspis ferox, Carcharhinus amblyrhynchos, Loxodon macrorhinus, Rhynchobatus djiddensis, Glaucostegus granulatus, Dasyatis microps,Himantura undulate, Himantura alcockii, Himantura marginata, Himantura imbricate, Himantura uarnak, Himantura jenkinsii, Himantura bleekeri, Aetobatus narinari, Rhinoptera javanica, Mobula japonica, Aetomylaeus nichofii, Mobula eregoodootenkee, Aetomylaeus maculatus, Aetomylaeus vespertilio |
| Tuna-like | Katsuwonus pelamis, Acanthocybium solandri, Thunnus alalunga, Tetrapturus angustirostris, Istiophorus platypterus, <br> Istiompax indica, Makaira nigricans, Xiphias gladius |
| Bigeye tuna | Thunnus obesus |
| Yellowfin tuna | Thunnus albacares |
| Marlins | Kajikia audax, Makaira mazara |
| Coastal scombrids | Auxis thazard thazard, Sarda orientalis, Auxis rochei rochei, Euthynnus affinis, Thunnus tonggol, Scomberomorus commerson, Scomberomorus guttatus, Scomberomorus lineolatus, Gymnosarda unicolor |
| Carangids | Carangoides fulvoguttatus, Caranx sexfasciatus, Gnathanodon speciosus, Caranx melampygus, Seriolina nigrofasciata, Carangoides orthogrammus, Carangoides equula, Carangoides chrysophrys, Carangoides malabaricus, Carangoides hedlandensis, Carangoides ferdau, Carangoides coeruleopinnatus, Carangoides talamparoides, Carangoides armatus, Caranx ignobilis, Alectis indica, Scomberoides lysan, Scomberoides commersonnianus, Alectis ciliaris, Elagatis bipinnulata, Seriola rivoliana, Caranx lugubris, Megalaspis cordyla, Selar crumenophthalmus, Uraspis helvola, Parastromateus niger, Alepes djedaba, Decapterus macrosoma, Decapterus russelli |
| L pisc | Sphyraena jello, Sphyraena barracuda, Sphyraena obtusata, Lates calcarifer, Pristis perotteti, Lepturacanthus savala, Trichiurus lepturus, Coryphaena hippurus, Coryphaena equiselis, Pomatomus saltatrix, Albula vulpes, Ablennes hians, Tylosurus crocodilus crocodilus, Strongylura leiura, Chirocentrus dorab, Chirocentrus nudus, Cheilinus undulatus, Strophidon sathete, Brotula multibarbata, Platycephalus indicus, Rachycentron canadum, Muraenesox bagio, Muraenesox cinereus, Congresox talabonoides, Fistularia petimba, Fistularia commersonii, Lobotes surinamensis, Leptomelanosoma indicum, Eleutheronema tetradactylum, Plotosus canius, Conger cinereus, Megalops cyprinoides |
| L pisc comm | Epinephelus fuscoguttatus, Epinephelus lanceolatus, Epinephelus coioides, Epinephelus flavocaeruleus, Epinephelus malabaricus, Epinephelus latifasciatus, Epinephelus multinotatus, Epinephelus polyphekadion, Plectropomus areolatus, Plectropomus pessuliferus, Plectropomus Iaevis, Variola louti, Epinephelus morrhua, Cephalopholis argus, Aethaloperca rogaa, Epinephelus chlorostigma, Epinephelus undulosus, Epinephelus fasciatus, Epinephelus areolatus, Epinephelus bleekeri, Epinephelus tauvina, Cephalopholis miniata, Cephalopholis boenak, <br> Etelis coruscans, Lutjanus erythropterus, Lutjanus johnii, Lutjanus malabaricus, Lutjanus sanguineus, Lutjanus sebae, Lutjanus argentimaculatus, Aphareus rutilans, Aprion virescens, Etelis carbunculus, Etelis radiosus, Pristipomoides filamentosus, Pristipomoides multidens, Lutjanus ehrenbergii, Lutjanus monostigma, Lutjanus vitta, Lutjanus fulvus, Lutjanus lutjanus, Lutjanus carponotatus, Lutjanus rivulatus, Lutjanus gibbus, Lutjanus kasmira, Lipocheilus carnolabrum, Pristipomoides auricilla, Pristipomoides sieboldii, Pristipomoides zonatus, Lutjanus quinquelineatus, Lutjanus fulviflamma, Lutjanus bohar, Macolor macularis, Macolor niger |


| Functional group | Species list |
| :---: | :---: |
|  | Protonibea diacanthus, Otolithoides biauritus, Otolithoides pama, Pterotolithus maculatus, Otolithes ruber, Otolithes cuvieri, Johnius carutta, Johnius dussumieri, Johnius borneensis, Johnius macrorhynus, Pennahia anea, Pennahia argentata <br> Lethrinus ornatus, Lethrinus nebulosus, Lethrinus microdon, Lethrinus harak, Lethrinus olivaceus, Lethrinus lentjan, Lethrinus erythracanthus, Wattsia mossambica, Gymnocranius grandoculis, Lethrinus xanthochilus, Lethrinus rubrioperculatus, <br> Saurida tumbil, Saurida undosquamis, Harpadon nehereus |
| Milkfish plus | Chamos chanos, Panfasius pangasius, Netuma thalassina, Diodon, hystrix, Arothron stellatus |
| SM inv | Acanthopagrus latus, Acanthopagrus berda, Rhabdosargus sarba, Pampus argenteus, Pampus chinensis, Acanthurus triostegus, Acanthurus leucosternon, |
|  | Siganus canaliculatus, Siganus guttatu, Siganus fuscescens, Siganus argenteus, Siganus corallinus, Siganus lineatus, Siganus puelloides, Myripristis murdjan, Kyphosus cinerascens, Monotaxis grandoculis, Sillago sihama, Sillago aeolus, Sillago chondropus, Sillaginopsis panijus, Gymnocranius griseus, Scatophagus argus, Mugil cephalus, Chelon planiceps, Chelon macrolepis, Moolgarda seheli Liza subviridis, Valamugil cunnesius, Valamugil speigleri, Cynoglossus arel, Cynoglossus puncticeps, Cynoglossus lingua, Cynoglossus bilineatus, Arius maculatus, Arius venosus, Mene maculata, Caesio caerulaurea, Caesio lunaris |
|  | Acanthurus auranticavus, Acanthurus bariene, Acanthurus dussumieri, Acanthurus guttatus, Acanthurus lineatus, Acanthurus mata, Acanthurus nigricans, Acanthurus nigricauda, Acanthurus nigrofuscus, Acanthurus tennentii, Acanthurus thompsoni, Acanthurus xanthopterus, Aluterus monoceros, Aluterus scriptus, Scarus festivus, Scarus ghobban, Scarus prasiognathos, Scarus quoyi Scarus rubroviolaceus, Scarus tricolor, Chelonodon patoca, Arothron immaculatus, Platax orbicularis, Cynoglossus lida |
|  | Chelmon rostratus, Chaetodon plebeius, Chaetodon vagabundus, Polydactylus multiradiatus, Polydactylus multiradiatu, Nemipterus furcosus, Triacanthus biaculeatus, Liza parsia, Plotosus lineatus, <br> Centropyge bicolor, Nuchequula blochii, Nuchequula gerreoides, Secutor ruconius, Nuchequula blochii, Nuchequula gerreoides, Secutor ruconius, Leiognathus daura, Leiognathus brevirostris, Gazza minuta, Equulites leuciscus, Eubleekeria splendens, Leiognathus equulus, Photopectoralis bindus, Secutor insidiator |
|  | Gerres filamentosus, Pentaprion longimanus, Gerres oyena, Selaroides leptolepis, Selar boops, Pterocaesio pisang, Johnius coitor, Grammatobothus polyophthalmus, Engyprosopon grandisquama, Petroscirtes breviceps, Amphiprion ocellaris, Atherinomorus lacunosus, Pseudocheilinus hexataenia, Halichoeres melanurus, Thalassoma amblycephalum, Apogon coccineus, Apogon crassiceps, Apogon doryssa, Apogon ellioti, Fusigobius maximus, Periophthalmodon schlosseri, Oxyurichthys microlepis, Paramonacanthus japonicus, Paramonacanthus curtorhynchos, Aseraggodes umbratilis, Pardachirus pavoninus, Upeneus sulphureus, Upeneus moluccensis, Upeneus vittatus |
| SM pisc | Argyrops spinifer, Saurida gracilis, Dactyloptena orientalis, Atule mate, Priacanthus macracanthus, Priacanthus tayenus, Priacanthus hamrur, Pseudorhombus arsius, Pseudorhombus javanicus, Psettodes erumei, Glossogobius giuris, Abalistes stellaris, Scomberoides tol, Naucrates ductor, Drepane punctata, Drepane longimana, Taractichthys steindachneri, Neoniphon sammara, Sciades sona, Terapon theraps, Terapon jarbua, Pterois miles, Pterois russelii, Pterois volitans, Scorpaenopsis diabolus, Scorpaenopsis oxycephala, Brachirus orientalis, Synaptura albomaculata, |
|  | Diagramma pictum, Pomadasys argenteus, Pomadasys maculatus, Pomadasys argyreus, Pomadasys furcatus, Pomadasys olivaceus, Plectorhinchus pictus, Plectorhinchus albovittatus, Plectorhinchus chaetodonoides, Plectorhinchus gibbosus, Plectorhinchus lineatus, Plectorhinchus picus, Plectorhinchus vittatus, |
|  | Nemipterus peronei, Nemipterus bathybius, Nemipterus bipunctatus, Nemipterus randalli, Nemipterus hexodon, Nemipterus japonicus, Nemipterus nematophorus, Parascolopsis inermis, |


| Functional group | Species list |
| :---: | :---: |
|  | Scolopsis bilineata, Scolopsis vosmeri, Scolopsis xenochrous, Onigocia macrolepis, Synodus hoshinonis Brachypleura novaezeelandiae, Dendrophysa russelii, Eleotris fusca, Ambassis gymnocephalus, Filimanus heptadactyla, Polynemus paradiseus, Syngnathoides biaculeatus, Ephippus orbis, Ariomma indicum, Dendrochirus biocellatus, Dendrochirus brachypterus, Dendrochirus zebra, Pterois antennata, Pterois radiata, Scorpaenodes albaiensis |
| Indian Mackerel | Rastrelliger kanagurta, Rastrelliger brachysoma, Rastrelliger faughni |
| Hilsa | Tenualosa ilisha, Tenualosa toli, Hilsa kelee, Ilisha melastoma, Ilisha filigera, Ilisha megaloptera, Ilisha elongata, Gudusia chapra |
| Small pelagics | Anodontostoma chacunda, Sardinella longiceps, Sardinella fimbriata, Dussumieria elopsoides, Nematalosa nasus, Herklotsichthys quadrimaculatus, Spratelloides delicatulus, Sardinella gibbosa, Sardinella melanura, Sardinella albella, Sardinella lemuru, Opisthopterus tardoore, Spratelloides gracilis, Amblygaster sirm, Dussumieria acuta, Raconda russeliana, Escualosa thoracata, Bregmaceros mcclellandi <br> Coilia reynaldi, Thryssa vitrirostris, Coilia dussumieri, Setipinna taty, Stolephorus commersonnii, Thryssa dussumieri, Thryssa mystax, Encrasicholina devisi, Encrasicholina heteroloba, Stolephorus insularis, Thryssa baelama, Stolephorus waitei, Stolephorus indicus, Thryssa hamiltonii, Coilia ramcarati, Encrasicholina punctifer <br> Atropus atropos, Lactarius lactarius, Pellona ditchela, Alepes melanoptera, Hemiramphus convexus, Rhynchorhamphus georgii, Cheilopogon spilopterus, Parexocoetus mento, Exocoetus volitans, Hirundichthys coromandelensis, Cheilopogon abei, Cheilopogon atrisignis, Cheilopogon cyanopterus, Cheilopogon furcatus, Cheilopogon nigricans, Cheilopogon suttoni, Hyporhamphus limbatus, Hyporhamphus unicuspis, Hyporhamphus balinensis, Zenarchopterus dispar |
| S bathy | Myctophidae, Stomiidae, Diretmidae |
| ML bathy | e.g. Pontinus macrocephalus, Arctozenus risso, Nemichthys scolopaceus, Lampris guttatus, Alepisaurus ferox, Trachipterus jacksonensis |
| jellyfish | e.g. Cephea, Rhizostomatidae, Rhopilema, Scyphozoa (based on catches) |
| cephalopods | Sepia sp, Loligo sp, Octopus vulgaris |
| Macrobenthos | shrimps, prawns, lobsterx, crabs, molluscs, worms, small benthic crustaceans |
| Meiobenthos | e,g, benthic copepods |
| zooplankton | copepods, fish and invertebrate eggs, euphausiids, chaetognats, siphonophora |

## A2 Catch, effort and CPUEs

A2.1 Total catch (t/km2/year) per fleet and functional group as input in Ecopath for year 1978.
$\mathrm{BET}=$ bigeye tuna, $\mathrm{YFT}=$ yellowfin tuna. The total catch in t is divided by the area of the entire study area.

|  | Group name | Region1 | Region3 | Region 2 industrial | Region 2 small-scale | Hilsa | BET | YFT | Marlins | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Oceanic sharks | 0.001 | 0.0035 | 0.008082 | 0.00391 | 0 | 0 | 0 | 0 | 0.016493 |
| 2 | Coastal elasmobranch | 0.001198 | 0.006461 | 0.008798 | 0.009074 | 0 | 0 | 0 | 0 | 0.025531 |
| 3 | Tuna-like | 0.005722 | 0.004945 | 0.001569 | 0.004825 | 0 | 0 | 0 | 0 | 0.01706 |
| 4 | Coastal scombrids | 0.001283 | 0.012144 | 0.002552 | 0.01328 | 0 | 0 | 0 | 0 | 0.02926 |
| 5 | Jellyfish | 0 | 0.000112 | 0.000134 | 0 | 0 | 0 | 0 | 0 | 0.000246 |
| 6 | Cephalopods | 0 | 0.013 | 0.003117 | 0.003597 | 0 | 0 | 0 | 0 | 0.019715 |
| 7 | S bathy | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8 | ML bathy | 0.000251 | 0 | 0.000632 | 0.00258 | 0 | 0 | 0 | 0 | 0.003463 |
| 9 | 1 L pisc | 0.000448 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.000448 |
| 10 | 1 Carangids | 0.000877 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.000877 |
| 11 | 1 S pelagics | 0.000341 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.000341 |
| 12 | 1 L pisc comm | 0.00075 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00075 |
| 13 | 1 SM inv | 0.000348 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.000348 |
| 14 | 1 SM pisc | 0.000223 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.000223 |
| 15 | 1 Macrobenthos | $3.07 \mathrm{E}-07$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3.07E-07 |
| 16 | 1 Meiobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 17 | 1 Zooplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 18 | 2 Lpisc | 0 | 0 | 0.01281 | 0.049802 | 0 | 0 | 0 | 0 | 0.062613 |
| 19 | 2 Carangids | 0 | 0 | 0.004471 | 0.01644 | 0 | 0 | 0 | 0 | 0.020911 |
| 20 | 2 S pelagics | 0 | 0 | 0.023344 | 0.08478 | 0 | 0 | 0 | 0 | 0.108124 |
| 21 | 2 L pisc comm | 0 | 0 | 0.011675 | 0.065067 | 0 | 0 | 0 | 0 | 0.076742 |
| 22 | 2 Milkfish plus | 0 | 0 | 0.002048 | 0.012956 | 0 | 0 | 0 | 0 | 0.015004 |
| 23 | 2 SM inv | 0 | 0 | 0.025159 | 0.067588 | 0 | 0 | 0 | 0 | 0.092747 |
| 24 | 2 SM pisc | 0 | 0 | 0.005138 | 0.026823 | 0 | 0 | 0 | 0 | 0.031961 |
| 25 | 2 Crustaceans | 0 | 0 | 0.041264 | 0.035892 | 0 | 0 | 0 | 0 | 0.077156 |
| 26 | 2 Macrobenthos | 0 | 0 | 0.001875 | 0.014206 | 0 | 0 | 0 | 0 | 0.016081 |
| 27 | 2 Meiobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 28 | 2 Zooplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 29 | 2,3 Hilsa | 0 | 0 | 0 | 0 | 0.03662 | 0 | 0 | 0 | 0.036623 |
| 30 | 2,3 Indian mackerel | 0 | 0.026994 | 0.016928 | 0.004008 | 0 | 0 | 0 | 0 | 0.04793 |
| 31 | 3 Lpisc | 0 | 0.029437 | 0 | 0 | 0 | 0 | 0 | 0 | 0.029437 |
| 32 | 3 Carangids | 0 | 0.031914 | 0 | 0 | 0 | 0 | 0 | 0 | 0.031914 |
| 33 | 3 S pelagics | 0 | 0.046784 | 0 | 0 | 0 | 0 | 0 | 0 | 0.046784 |
| 34 | 3 L pisc comm | 0 | 0.042999 | 0 | 0 | 0 | 0 | 0 | 0 | 0.042999 |
| 35 | 3 Milkfish plus | 0 | 0.006744 | 0 | 0 | 0 | 0 | 0 | 0 | 0.006744 |
| 36 | 3 SM inv | 0 | 0.043341 | 0 | 0 | 0 | 0 | 0 | 0 | 0.043341 |
| 37 | 3 SM pisc | 0 | 0.031763 | 0 | 0 | 0 | 0 | 0 | 0 | 0.031763 |
| 38 | 3 Crustaceans | 0 | 0.070085 | 0 | 0 | 0 | 0 | 0 | 0 | 0.070085 |
| 39 | 3 Macrobenthos | 0 | 0.010223 | 0 | 0 | 0 | 0 | 0 | 0 | 0.010223 |
| 40 | 3 Meiobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 41 | 3 Zooplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 42 | 1 Phytoplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 43 | 2 Phytoplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 44 | 3 Phytoplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 45 | Benthic plants | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 46 | Bigeye tuna | 0 | 0 | 0 | 0 | 0 | 0.004 | 0 | 0 | 0.004 |
| 47 | Yellowfin tuna | 0 | 0 | 0 | 0 | 0 | 0 | 0.0065 | 0 | 0.0065 |
| 48 | Marlins | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0011 | 0.0011 |
| 49 | Detritus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 50 | Sum | 0.012441 | 0.380447 | 0.169596 | 0.414828 | 0.036623 | 0.004 | 0.0065 | 0.0011 | 1.025534 |

A2.2. Effort in horse power days for the east coast of India (Bhathal 2013) and relative effort as used for region 2 in Ecosim.

| Marine fishing effort in horse power days |  |  |  |  |  | Relative effort for region 2 |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
|  | Small-scale | Industrial | Large trawlers | Small-scale | All industrial $a$ | Industrial only |  |  |
| 1978 | $95,482,790$ | $53,782,186$ | $2,315,460$ | 1 | 1 | 1 |  |  |
| 1979 | $102,335,975$ | $57,186,311$ | $2,970,688$ | 1.072 | 1.072 | 1.063 |  |  |
| 1980 | $108,636,178$ | $61,228,736$ | $4,283,750$ | 1.138 | 1.168 | 1.138 |  |  |
| 1981 | $112,113,803$ | $64,684,161$ | $4,307,393$ | 1.174 | 1.230 | 1.203 |  |  |
| 1982 | $117,647,129$ | $70,187,123$ | $4,497,938$ | 1.232 | 1.331 | 1.305 |  |  |
| 1983 | $124,238,927$ | $76,740,390$ | $3,256,759$ | 1.301 | 1.426 | 1.427 |  |  |
| 1984 | $130,951,700$ | $83,406,468$ | $3,084,113$ | 1.371 | 1.542 | 1.551 |  |  |
| 1985 | $143,445,594$ | $9,821,044$ | $4,101,788$ | 1.502 | 1.781 | 1.782 |  |  |
| 1986 | $150,505,299$ | $102,733,189$ | $4,496,053$ | 1.576 | 1.911 | 1.910 |  |  |
| 1987 | $154,487,496$ | $108,202,947$ | $2,053,794$ | 1.618 | 1.965 | 2.012 |  |  |
| 1988 | $181,710,800$ | $136,865,740$ | $2,451,666$ | 1.903 | 2.483 | 2.545 |  |  |
| 1989 | $212,923,659$ | $166,580,613$ | $2,758,574$ | 2.230 | 3.019 | 3.097 |  |  |
| 1990 | $246,342,990$ | $196,892,566$ | $3,010,675$ | 2.580 | 3.563 | 3.661 |  |  |
| 1991 | $293,766,283$ | $235,343,022$ | $2,909,193$ | 3.077 | 4.247 | 4.376 |  |  |
| 1992 | $320,745,757$ | $257,653,290$ | $2,780,838$ | 3.359 | 4.643 | 4.791 |  |  |
| 1993 | $347,515,519$ | $283,497,874$ | $2,596,934$ | 3.640 | 5.100 | 5.271 |  |  |
| 1994 | $401,001,786$ | $335,259,000$ | $2,355,627$ | 4.200 | 6.018 | 6.234 |  |  |
| 1995 | $440,934,984$ | $370,550,807$ | $1,910,840$ | 4.618 | 6.640 | 6.890 |  |  |
| 1996 | $481,578,826$ | $406,553,257$ | $1,544,876$ | 5.044 | 7.275 | 7.559 |  |  |
| 1997 | $522,935,045$ | $443,268,084$ | $1,203,612$ | 5.477 | 7.923 | 8.242 |  |  |
| 1998 | $565,479,216$ | $480,696,473$ | 896,642 | 5.922 | 8.585 | 8.938 |  |  |
| 1999 | $608,737,772$ | $518,839,247$ | 929,483 | 6.375 | 9.265 | 9.647 |  |  |
| 2000 | $608,737,772$ | $518,839,247$ | 889,852 | 6.375 | 9.265 | 9.647 |  |  |
| 2001 | $608,737,772$ | $518,839,247$ | 894,039 | 6.375 | 9.265 | 9.647 |  |  |
| 2002 | $608,737,772$ | $518,839,247$ | 891,919 | 6.375 | 9.265 | 9.647 |  |  |
| 2003 | $608,737,772$ | $518,839,247$ | 905,565 | 6.375 | 9.265 | 9.647 |  |  |
| 2004 | $608,737,772$ | $518,839,247$ | 919,428 | 6.375 | 9.265 | 9.647 |  |  |
| 2005 | $608,737,772$ | $518,839,247$ | 918,256 | 6.375 | 9.265 | 9.647 |  |  |

$a$ includes large trawlers; used in the Ecosim model section
b excludes large trawlers; used to generate Ecosim results found in Appendix A6.

A2.3. Biomass, catch, and fishing effort directed to hilsa compiled for Bangladesh, the extrapolation for 1978-1986, and the resulting fishing effort time series used in Ecosim (relative value).
Both catches and biomass densities are for the entire study area.

|  | Effort in number of boats a |  | Extrapolation back to 1978 |  | Total | Relative effort $b$ | Biomass$\mathrm{t} / \mathrm{km}^{2} \mathrm{c}$ | Catches$\mathrm{t} / \mathrm{km}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Nonmotorized | motorized | Nonmotorized | Motorized |  |  |  |  |
| 1978 |  |  | 3633 | 3639 | 7272 | 1 |  | 0.037 |
| 1979 |  |  | 3678 | 3684 | 7362 | 1.012 |  | 0.036 |
| 1980 |  |  | 3723 | 3729 | 7451 | 1.025 |  | 0.035 |
| 1981 |  |  | 3768 | 3774 | 7541 | 1.037 |  | 0.035 |
| 1982 |  |  | 3812 | 3818 | 7631 | 1.049 |  | 0.035 |
| 1983 |  |  | 3857 | 3863 | 7721 | 1.062 |  | 0.036 |
| 1984 |  |  | 3902 | 3908 | 7810 | 1.074 |  | 0.037 |
| 1985 |  |  | 3947 | 3953 | 7900 | 1.086 |  | 0.037 |
| 1986 |  |  | 3992 | 3998 | 7990 | 1.099 |  | 0.043 |
| 1987 | 4037 | 4043 |  |  | 8080 | 1.111 | 0.219 | 0.046 |
| 1988 | 5186 | 4284 |  |  | 9470 | 1.302 | 0.200 | 0.044 |
| 1989 | 6336 | 4525 |  |  | 10861 | 1.494 | 0.219 | 0.046 |
| 1990 | 7486 | 4766 |  |  | 12252 | 1.685 | 0.253 | 0.047 |
| 1991 | 8635 | 5007 |  |  | 13643 | 1.876 | 0.230 | 0.049 |
| 1992 | 9785 | 5248 |  |  | 15033 | 2.067 | 0.223 | 0.055 |
| 1993 | 10935 | 5490 |  |  | 16424 | 2.259 | 0.212 | 0.057 |
| 1994 | 12084 | 5731 |  |  | 17815 | 2.450 | 0.195 | 0.049 |
| 1995 | 13234 | 5972 |  |  | 19205 | 2.641 | 0.191 | 0.055 |
| 1996 | 14383 | 6213 |  |  | 20596 | 2.832 | 0.181 | 0.058 |
| 1997 | 15533 | 6454 |  |  | 21987 | 3.024 | 0.163 | 0.056 |
| 1998 | 16683 | 6695 |  |  | 23378 | 3.215 | 0.126 | 0.055 |
| 1999 | 17832 | 6936 |  |  | 24768 | 3.406 | 0.175 | 0.054 |
| 2000 | 18982 | 7177 |  |  | 26159 | 3.597 | 0.122 | 0.055 |
| 2001 | 20132 | 6377 |  |  | 26509 | 3.645 | 0.152 | 0.061 |
| 2002 | 21281 | 6618 |  |  | 27899 | 3.837 | 0.130 | 0.056 |
| 2003 | 22431 | 6859 |  |  | 29290 | 4.028 | 0.132 | 0.051 |
| 2004 | 23581 | 7100 |  |  | 30681 | 4.219 | 0.159 | 0.059 |
| 2005 | 24730 | 7341 |  |  | 32072 | 4.410 | 0.156 | 0.061 |
| 2006 | 25880 | 7582 |  |  | 33462 | 4.602 | 0.153 | 0.062 |
| 2007 |  |  |  |  |  | 5.522 | 0.165 | 0.060 |
| 2008 |  |  |  |  |  | 5.522 | 0.178 | 0.062 |
| 2009 |  |  |  |  |  | 5.522 | 0.193 | 0.062 |
| 2010 |  |  |  |  |  | 5.522 | 0.213 | 0.073 |

a from (BOBLME 2012)
$b$ the time series was continued to 2010 by assuming constant effort from 2006 to 2010 (Rishi Sharma, pers.
comm., IOTC)
c Rishi Sharma, pers. comm., IOTC, based on results of single species model described in (BOBLME 2012)

A2.4 Effort by large pelagic fleet contributed by Rishi Sharma, IOTC.

| Year | Striped Marlin |  | Blue Marlin |  | $\begin{gathered} \hline \text { marlins } \\ \hline a \end{gathered}$ | Bigeye | Yellowfin | Kawakawa | Skipjack |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Taiwan | Japan | Taiwan | Japan |  |  |  |  |  |
| 1960 |  |  |  |  |  | 1.42 |  |  |  |
| 1961 |  |  |  |  |  | 1.44 |  |  |  |
| 1962 |  |  |  |  |  | 1.47 |  |  |  |
| 1963 |  |  |  |  |  | 1.28 | 0.54 |  |  |
| 1964 |  |  |  |  |  | 1.35 | 0.63 |  |  |
| 1965 |  |  |  |  |  | 1.15 | 0.49 |  |  |
| 1966 |  |  |  |  |  | 1.27 | 0.67 |  |  |
| 1967 |  |  |  |  |  | 1.25 | 0.53 |  |  |
| 1968 |  |  |  |  |  | 1.13 | 0.53 |  |  |
| 1969 |  |  |  |  |  | 1.18 | 0.53 |  |  |
| 1970 |  |  |  |  |  | 1.11 | 0.63 |  |  |
| 1971 |  | 1.38 |  | 1.49 | 1.44 | 0.86 | 0.48 |  |  |
| 1972 |  | 1.43 |  | 2.04 | 1.74 | 1.03 | 0.42 |  |  |
| 1973 |  | 2.62 |  | 2.34 | 2.48 | 0.97 | 0.51 |  |  |
| 1974 |  | 2.39 |  | 2.10 | 2.24 | 1.02 | 0.33 |  |  |
| 1975 |  | 2.68 |  | 1.66 | 2.17 | 1.02 | 0.29 |  |  |
| 1976 |  | 2.36 |  | 1.90 | 2.13 | 1.46 | 0.37 |  |  |
| 1977 |  | 5.64 |  | 2.69 | 4.17 | 2.07 | 0.42 |  |  |
| 1978 |  | 4.49 |  | 2.06 | 3.28 | 2.08 | 0.38 |  |  |
| 1979 |  | 3.99 |  | 1.58 | 2.79 | 1.61 | 0.24 |  |  |
| 1980 | 2.93 | 4.41 | 1.05 | 2.32 | 3.37 | 1.45 | 0.29 |  |  |
| 1981 | 2.47 | 2.76 | 1.25 | 1.80 | 2.28 | 0.99 | 0.28 |  |  |
| 1982 | 1.54 | 2.47 | 1.03 | 1.69 | 2.08 | 1.26 | 0.24 |  |  |
| 1983 | 0.99 | 1.39 | 1.15 | 2.17 | 1.78 | 1.30 | 0.34 |  |  |
| 1984 | 1.35 | 2.39 | 1.42 | 1.98 | 2.18 | 0.94 | 0.33 |  |  |
| 1985 | 1.63 | 2.49 | 1.15 | 2.29 | 2.39 | 0.85 | 0.34 |  | 1.46 |
| 1986 | 2.12 | 2.44 | 1.49 | 1.65 | 2.04 | 0.96 | 0.34 |  | 1.37 |
| 1987 | 1.37 | 1.10 | 1.26 | 1.54 | 1.32 | 1.12 | 0.30 |  | 1.37 |
| 1988 | 1.06 | 0.83 | 1.03 | 1.31 | 1.07 | 0.95 | 0.38 |  | 1.43 |
| 1989 | 0.79 | 0.60 | 0.77 | 0.93 | 0.77 | 1.02 | 0.26 |  | 1.34 |
| 1990 | 0.46 | 0.50 | 0.57 | 0.83 | 0.67 | 0.83 | 0.39 |  | 1.37 |
| 1991 | 1.10 | 0.88 | 0.70 | 0.81 | 0.85 | 1.02 | 0.19 |  | 1.37 |
| 1992 | 1.02 | 0.84 | 0.88 | 0.95 | 0.90 | 0.65 | 0.19 |  | 1.34 |
| 1993 | 0.83 | 0.76 | 0.78 | 0.98 | 0.87 | 0.92 | 0.20 |  | 1.28 |
| 1994 | 1.78 | 0.92 | 0.76 | 1.23 | 1.07 | 0.96 | 0.15 |  | 1.31 |
| 1995 | 1.43 | 0.99 | 0.86 | 0.80 | 0.90 | 0.75 | 0.14 |  | 1.23 |
| 1996 | 1.21 | 0.77 | 0.89 | 0.66 | 0.71 | 0.82 | 0.12 |  | 1.11 |
| 1997 | 1.05 | 0.57 | 1.15 | 0.94 | 0.76 | 0.68 | 0.14 |  | 1.07 |
| 1998 | 0.68 | 0.29 | 1.04 | 0.82 | 0.55 | 0.64 | 0.11 |  | 0.97 |
| 1999 | 0.88 | 0.50 | 1.33 | 0.77 | 0.64 | 0.73 | 0.14 |  | 0.93 |
| 2000 | 0.62 | 0.35 | 1.15 | 0.75 | 0.55 | 0.58 | 0.18 |  | 0.83 |
| 2001 | 0.73 | 0.31 | 1.02 | 0.50 | 0.41 | 0.63 | 0.10 |  | 0.93 |
| 2002 | 0.65 | 0.25 | 1.06 | 0.42 | 0.33 | 0.51 | 0.08 |  | 0.88 |
| 2003 | 0.58 | 0.17 | 0.96 | 0.37 | 0.27 | 0.59 | 0.07 |  | 1.00 |
| 2004 | 0.61 | 0.11 | 0.86 | 0.35 | 0.23 | 0.66 | 0.10 | 0.86 | 0.68 |
| 2005 | 0.37 | 0.11 | 0.83 | 0.30 | 0.20 | 0.73 | 0.06 | 1.09 | 0.80 |
| 2006 | 0.30 | 0.17 | 0.78 | 0.43 | 0.30 | 0.67 | 0.10 | 0.75 | 0.78 |
| 2007 | 0.15 | 0.14 | 0.62 | 0.44 | 0.29 | 0.57 | 0.08 | 1.41 | 0.55 |
| 2008 | 0.31 | 0.22 | 0.79 | 0.35 | 0.29 | 0.59 | 0.04 | 0.82 | 0.51 |
| 2009 | 0.16 | 0.21 | 0.87 | 0.36 | 0.29 | 0.45 | 0.04 | 1.16 | 0.46 |
| 2010 | 0.41 | 0.93 | 1.11 | 0.54 | 0.74 | 0.48 | 0.03 | 1.01 | 0.37 |
| 2011 | 0.42 | 1.12 | 1.40 | 0.70 | 0.91 | 0.82 | 0.03 | 0.91 | 0.28 |

[^3]
## A3 Diet compositions

## A3.1 Source of fish diets.

References labeled FB paired with a number correspond to Fishbase document number; FB, items are species for which there is only a list of food items.

| Species | Location | Source |
| :---: | :---: | :---: |
| Indian mackerel |  |  |
| Rastrelliger kanagurta |  | FB |
| Rastrelliger kanagurta |  | FB |
| Hilsa |  |  |
| Tenualosa ilisha | Bangladesh, freshwater to marine | FB 4837 |
| Tenualosa ilisha | marine, Bangladesh | (Dutta et al. 2013) |
| Oceanic sharks |  |  |
| Pteroplatytrygon violacea |  | FB, items |
| Carcharhinus falciformis | NW Atlantic USA | FB 37512 |
| Alopias superciliosus | NW Atlantic USA | FB 37512 |
| Alopias vulpinus | NW Atlantic USA | FB 37512 |
| Alopias vulpinus |  | FB, items |
| Isurus oxyrinchus | NW Atlantic | FB 37512 |
| Isurus oxyrinchus | NW Atlantic | FB 37512 |
| Prionace glauca | NW Atlantic | FB 37512 |
| Prionace glauca | Monterey Bay California | FB 28071 |
| Isurus paucus | NW Atlantic | FB 37512 |
| Centrophorus squamosus | S Africa | FB 12473 |
| Coastal elasmobranch |  |  |
| Rhizoprionodon acutus | Australia | FB 13356 |
| Carcharhinus limbatus | South Africa 1978-91 | FB 26970 |
| Galeocerdo cuvier | NW Atlantic | FB 37512 |
| Himantura uarnak | Kuwait | food items +FB 37858 |
| Mustelus manazo | Japan | FB 47252 |
| Tuna-like |  |  |
| Thunnus albacares | Andaman sea | (Panjarat 1999) |
| Katsuwonus pelamis | Mozambique 1989 | FB 9035 |
| Thunnus obesus | Solomon Is. June 1993 | FB 28765 |
| Xiphias gladius | North BOB | (Nootmorn et al. 2008) |
| Xiphias gladius | Algeria, Annaba Gulf 1986-87 | FB 419911 |
| Xiphias gladius | North +Tropical Atlantic | FB 76866 |
| Makaira nigricans | SW equatorial Atlantiqc 1992-1993 | FB 51769 |
| Istiompax indica | Malaysia east coast 93-94 | FB 53850 |
| Istiophorus platypterus | Malaysia 1993-94 | FB 53850 |
| Acanthocybium solandri | G Mexico USA | FB 28119 |
| Coastal scombrids |  |  |
| Scomberomorus commerson | Solomon Is | FB 30531 |
| Scomberomorus commerson | Malaysia east coast 1993-94 | FB 53850 |
| Euthynnus affinis | Solomon Is | FB 30531 |
| Euthynnus affinis | Taiwan 2000-2001 | FB 53677 |
| Sarda orientalis |  | FB, items |
| Thunnus tonggol | Malaysia | FB 53850 |
| ML bathy |  |  |
| Alepisaurus ferox | Hawaii may 1990 | FB 12036 |
| Arctozenus risso | Kuril Is 87-92 | FB 41668 |
| Nemichthys scolopaceus | USA Newfoundland | FB 37512 |
| Carangids |  |  |
| Carangoides chrysophrys | New Caledonia 1985-98 | FB 55797 |
| Alectis indica |  | FB, items |


| Selar crumenophthalmus | Thailand 1985 | FB 26908 |
| :---: | :---: | :---: |
| Decapterus russelli | Manila Bay, Philippines | FB 761 |
| Megalaspis cordyla | Solomon Is | FB 30531 |
| Caranx sexfasciatus | Malaysia | FB 53850 |
| Caranx melampygus | Hawaii | FB 6057 |
| Alectis ciliaris | Columbia | FB 56479 |
| Elagatis bipinnulata | Brazil | FB 89206 |
| Carangoides ferdau | Malaysia 93-94 | FB 53850 |
| Carangoides ferdau | New Caledonia 85-97 | FB 55797 |
| L pisc |  |  |
| Albula vulpes | Florida | FB 30204 |
| Tylosurus crocodilus crocodilus | Solomon Is | FB 30531 |
| Chirocentrus dorab | Solomon Is | FB 30531 |
| Sphyraena jello | Malaysia 1993-94 | FB 53850 |
| Sphyraena jello |  | FB, items |
| Sphyraena barracuda | Puerto Rico 1958-1961 | FB 33 |
| Sphyraena obtusata | Malaysia 1993-94 | FB 53850 |
| Sphyraena obtusata | Solomon Is | FB 30531 |
| Brotula multibarbata | Hawaii | FB 13550 |
| Rachycentron canadum | Malaysia | FB 53850 |
| Trichiurus lepturus | W India, 1978 | FB 4424 |
| Lepturacanthus savala |  | FB, items |
| Coryphaena hippurus | Malaysia 1993-94 | FB 53850 |
| Pomatomus saltatrix | Brazil | FB 42756 |
| Megalops cyprinoides | New Caledonia | FB 55797 |
| L pisc comm |  |  |
| Lethrinus rubrioperculatus | New Caledonia 85-97 | FB 55797 |
| Lethrinus xanthochilus | New Caledonia 85-97 | FB 55797 |
| Lutjanus bohar | New Caledonia 85-97 | FB 55797 |
| Epinephelus polyphekadion | New Caledonia 85-97 | FB 55797 |
| Variola louti | New Caledonia 85-97 | FB 55797 |
| Etelis coruscans | Hawaii | FB 8925 |
| Lutjanus johnii | Gulf Carpentaria, Australia | FB 6932 |
| Lutjanus malabaricus | Malaysia | FB 53850 |
| Lutjanus sebae | Australia | FB 6932 |
| Aprion virescens | Hawaii | FB 8925 |
| Etelis carbunculus | Hawaii | FB 8926 |
| Pristipomoides filamentosus | East coast Malaysia | FB 53850 |
| Pristipomoides auricilla | N Marianas 1984 | FB 13792 |
| Lutjanus vitta | Australia | FB 6932 |
| Lutjanus fulvus | New Caledonia | FB 55797 |
| Lutjanus carponotatus | Australia | FB 26866 |
| Lutjanus quinquelineatus | New Caledonia | FB 55797 |
| Saurida tumbil | India, 1964-1968 | FB 6931 |
| Cephalopholis miniata | Egypt, Red Sea, 1978-83 | FB 6775 |
| Epinephelus areolatus | New Caledonia 1985-97 | FB 55797 |
| Epinephelus areolatus | Australia, Gulf Carpentaria 1990 | FB 6932 |
| Epinephelus coioides | New Caledonia 1985-97 | FB 55797 |
| Epinephelus malabaricus | New Caledonia 1985-97 | FB 55797 |
| Lethrinus nebulosus | New Caledonia 1985-97 | FB 55797 |
| Lethrinus nebulosus | Australia, Gulf Carpentaria 1990 | FB 6932 |
| Lutjanus argentimaculatus | New Caledonia 1985-1997 | FB 55797 |
| Pristipomoides sieboldii | Hawaii 1987-89 | FB 8925 |
| Pristipomoides zonatus | Hawaii 1987-90 | FB 8926 |
| Pristipomoides zonatus | N Marianas Pathfinder reef | FB 13792 |
| Lethrinus harak | New Caledonia 1985-97 | FB 55797 |


| Saurida undosquamis | India, NW BOB | FB 6931 |
| :---: | :---: | :---: |
| Gymnocranius grandoculis | New Caledonia 85-87 | FB 55797 |
| Lutjanus gibbus | Malaysia 85-97 | FB 55797 |
| Lutjanus kasmira | New Caledonia 85-97 | FB 55797 |
| Lutjanus fulviflamma | New Caledonia | FB, items |
| $S$ pelagics |  |  |
| Sardinella gibbosa |  | FB, items |
| Stolephorus indicus | Singapore | FB 51145 |
| Stolephorus insularis | Solomon Is | FB 32754 |
| Thryssa mystax | Kuwait | items + FB 37858 |
| Herklotsichthys | Kiribati 89-91 | FB 9004 |
| quadrimaculatus |  |  |
| Encrasicholina devisi | Solomon Is | FB 32754 |
| Spratelloides delicatulus | Solomon Is | FB 32754 |
| Spratelloides gracilis | Solomon Is | FB 32754 |
| Amblygaster sirm | Indonesia | FB 823 |
| SM inv |  |  |
| Upeneus sulphureus | Red Sea 1984 | FB 6292 |
| Pampus argenteus | Orissa, India BOB 1972 | FB 37087 |
| Pampus argenteus | Orissa, India BOB 1973 | FB 37087 |
| Pampus chinensis |  | FB, items |
| Acanthopagrus berda |  | FB, items |
| Acanthopagrus latus | Kuwait NW Arabian sea | FB 37858 |
| Mugil cephalus | Spain Valencia, summer | FB 50467 |
| Gerres filamentosus | New Caledonia | FB 55797 |
| Myripristis murdjan | Hawaii | FB 13550 |
| Myripristis murdjan | Madagascar | FB 78108 |
| Monotaxis grandoculis | Hawaii 69-70 | FB 13550 |
| Siganus fuscescens |  | FB, items |
| Siganus guttatus |  | FB, items |
| Sillago sihama | Australia 78 | FB 9638 |
| Moolgarda seheli |  | FB, items |
| Cynoglossus arel | India | FB 5260 |
| Chaetodon plebeius | Ryukyu Is | FB 6110 |
| Chaetodon vagabundus | Ryukyu Is | FB 6110 |
| Nuchequula gerreoides | Singapore | FB 51145 |
| Secutor ruconius | China | FB 26569 |
| Grammatobothus polyophthalmus | New Caledonia 85-97 | FB 55797 |
| Gazza minuta | Solomon Is | FB 30531 |
| Petroscirtes breviceps |  | FB, items |
| Centropyge bicolor |  | FB, items |
| Atherinomorus lacunosus | Marshall Is 1972 | FB 13784 |
| Pentaprion longimanus | Thailand | FB 26908 |
| Pseudocheilinus hexataenia | Ryukyu Is | FB 6110 |
| Halichoeres melanurus | Ryukyu Is | FB 6110 |
| Thalassoma amblycephalum | Ryukyu Is | FB 6110 |
| Scatophagus argus | New Caledonia 85-97 | FB 55797 |
| Selaroides leptolepis | Thailand 1985 | FB 26908 |
| Pterocaesio pisang | Japan | FB 36318 |
| Upeneus moluccensis | New Caledonia 85-97 | FB 55797 |
| Upeneus vittatus | New Caledonia 85-97 | FB 55797 |
| Nemipterus furcosus | Australia 1990 | FB 6932 |
| Equulites leuciscus | Thailand | FB 26908 |
| Photopectoralis bindus | Thailand | FB 26908 |
| Pentaprion longimanus | Thailand | FB 26908 |
| Acanthurus lineatus | Guam | FB 6155 |


| Acanthurus nigrofuscus | Ryuku Is | FB 6110 |
| :---: | :---: | :---: |
| Acanthurus xanthopterus | NA | FB, items |
| Apogon crassiceps | NA | FB, items |
| Apogon ellioti | G. Thailand 1985 | FB 26908 |
| Scarus ghobban | Ryulu Is, Philippines | FB6110, FB43166 |
| Scarus quoyi | Philippines 1991 | FB 43166 |
| Aluterus monoceros | Colombia | FB 46593 |
| Aluterus scriptus | Puerto Rico | FB 33 |
| SM pisc |  |  |
| Argyrops spinifer | Australia Gulf Carpentaria 1990 | FB 6932 |
| Dendrochirus brachypterus | New Caledonia 85-97 | FB 55797 |
| Dendrochirus zebra | Ryuku Is | FB 6110 |
| Plectorhinchus gibbosus | New Caledonia 85-97 | FB 55797 |
| Onigocia macrolepis | New Caledonia 85-97 | FB 55797 |
| Abalistes stellaris | New Caledonia 85-97 | FB 55797 |
| Eleotris fusca | Japan 1998 | FB 54520 |
| Nemipterus hexodon | Australia 1990 | FB 6932 |
| Atule mate | Thailand | FB 26908 |
| Neoniphon sammara | Madagascar | FB 78108 |
| Pomadasys argenteus | New Caledonia 85-97 | FB 55797 |
| Nemipterus peronii | Australia, Gulf Carpentaria 1990 | FB 6932 |
| Diagramma pictum | New Caledonia 1985-98 | FB 55797 |
| Diagramma pictum | Australia Gulf Carpentaria | FB 6932 |
| Priacanthus tayenus | Gulf Thailand | FB 26908 |
| Priacanthus hamrur | India east coast 1993 | FB 30941 |
| Psettodes erumei | India Porto novo 1972-73 | FB 6003 |
| Plectorhinchus pictus | East coast Peninsular Malaysia 93-94 | FB 53850 |
| Milkfish plus |  |  |
| Netuma thalassina | Kuwait Arabian Gulf | FB 37858, |
| Netuma thalassina | Australia, G Carpentaria 1990 | FB 6932 |
| Diodon hystrix | Hawaii | FB 3921 |
| Arothron stellatus | New Caledonia 1985-97 | FB 55797 |

## A3.2 Original diet matrix

| Prey | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 Oceanic shark | 0.0066 | 0.0295 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 Coastal shark ray | 0.0129 | 0.0581 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 Tuna-like | 0.0834 | 0.0018 | 0.0306 | 0 | 0 | 0 | 0 | 0.0019 | 0.0030 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 Coastal scombrids | 0.0111 | 0.0018 | 0.0426 | 0 | 0 | 0 | 0 | 0.0019 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 Jellyfish | 0.0100 | 0.0000 | 0.0000 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0125 | 0 | 0 | 0.0053 | 0 | 0 |
| 6 Cephalopods | 0.1605 | 0.0110 | 0.1124 | 0.0322 | 0 | 0.075 | 0 | 0.0640 | 0.0666 | 0.0631 | 0.0002 | 0.0236 | 0.0016 | 0.0103 | 0 |
| 71 Sbathy | 0.0379 | 0.0000 | 0.0561 | 0.0177 | 0 | 0.017 | 0 | 0.0691 | 0.0147 | 0.0271 | 0 | 0.0414 | 0 | 0.0010 | 0 |
| 81 ML bathy | 0.0658 | 0.0053 | 0.0246 | 0 | 0 | 0 | 0 | 0.0290 | 0.0028 | 0 | 0 | 0.0027 | 0 | 0 | 0 |
| 91 Lpisc | 0.0027 | 0.0032 | 0.0005 | 0.0003 | 0 | 0 | 0 | 0.0093 | 0.0316 | 0.0152 | 0 | 0.0121 | 0 | 0.0013 | 0 |
| 101 Carangids | 0.0007 | 0.0008 | 0.0084 | 0.0059 | 0 | 0.0025 | 0 | 0 | 0.1747 | 0.0729 | 0.0188 | 0.0541 | 0.0044 | 0.0645 | 0 |
| 111 S pelagics | 0.0012 | 0.0010 | 0.0083 | 0.0093 | 0 | 0.0019 | 0 | 0.0066 | 0.2702 | 0.2290 | 0.0223 | 0.0859 | 0.0170 | 0.0800 | 0 |
| 121 L pisc comm | 0.0006 | 0.0005 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0225 | 0.0110 | 0 | 0.0120 | 0.0004 | 0.0045 | 0 |
| 131 SM inv | 0.0019 | 0.0038 | 0.0018 | 0.0084 | 0 | 0.0037 | 0 | 0.0053 | 0.1548 | 0.2371 | 0.0188 | 0.1604 | 0.0078 | 0.0869 | 0 |
| 141 SM pisc | 0.0030 | 0.0032 | 0.0019 | 0.0010 | 0 | 0.0037 | 0 | 0.0106 | 0.0496 | 0.0837 | 0.0123 | 0.1097 | 0.0072 | 0.0373 | 0 |
| 151 Macrobenthos | 0.0012 | 0.0000 | 0.0000 | 0.0037 | 0 | 0.0460 | 0 | 0.1961 | 0.1840 | 0.1737 | 0.1940 | 0.4633 | 0.5940 | 0.7013 | 0.05 |
| 161 Meiobenthos | 0.0004 | 0.0000 | 0.0000 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0004 | 0.0101 | 0 | 0.0473 | 0.0001 | 0.20 |
| 171 Zooplankton | 0.0024 | 0.0000 | 0.0005 | 0.0004 | 0 | 0 | 0.4585 | 0.1509 | 0.0245 | 0.0533 | 0.6810 | 0.0345 | 0.1307 | 0.0123 | 0.10 |
| 182 L pisc | 0.0188 | 0.0222 | 0.0033 | 0.0023 | 0 | 0 | 0 | 0.0041 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 192 Carangids | 0.0050 | 0.0058 | 0.0582 | 0.0409 | 0 | 0.0174 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2025 pelagics | 0.0082 | 0.0072 | 0.0569 | 0.0641 | 0 | 0.0086 | 0 | 0.0029 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 212 L pisc comm | 0.0041 | 0.0037 | 0.0000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 222 Mikfish plus | 0.0013 | 0.0010 | 0.0068 | 1.5E-06 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 232 SM inv | 0.0128 | 0.0261 | 0.0125 | 0.0580 | 0 | 0.0254 | 0 | 0.0024 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 242 SM pisc | 0.0210 | 0.0224 | 0.0133 | 0.0071 | 0 | 0.0254 | 0 | 0.0048 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 252 Crustaceans | 0.0000 | 0.0646 | 0.0000 | 0.0198 | 0 | 0.1076 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 262 Macrobenthos | 0.0080 | 0.0233 | 0.0000 | 0.0064 | 0 | 0 | 0 | 0.0884 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 272 Meiobenthos | 0.0029 | 0.0000 | 0.0000 | 0 | 0 | 0.0104 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 282 Zooplankton | 0.0166 | 0.0000 | 0.0038 | 0.0002 | 0 | 0 | 0.2067 | 0.0680 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 29 2,3 Hilsa | 0.0413 | 0.0208 | 0.0226 | 0.0168 | 0 | $\begin{aligned} & 0.0100 \\ & 33128 \\ & 0.0250 \end{aligned}$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 302,3 Indian mackerel | 0.0290 | 0.0226 | 0.0327 | 0.0619 | 0 | 82821 | 0 | 0.0019 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 313 L pisc | 0.0607 | 0.0718 | 0.0106 | 0.0075 | 0 | 0 | 0 | 0.0068 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 323 Carangids | 0.0163 | 0.0188 | 0.1884 | 0.1324 | 0 | 0.0563 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 333 S pelagics | 0.0264 | 0.0232 | 0.1843 | 0.2075 | 0 | 0.0240 | 0 | 0.0049 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 343 L pisc comm | 0.0133 | 0.0121 | 0.0000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  |  |  | 4.7752 |  |  |  |  |  |  |  |  |  |  |  |
| 353 Milkfish plus | 0.0041 | 0.0034 | 0.0221 | 6E-06 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 363 SM inv | 0.0415 | 0.0844 | 0.0403 | 0.1876 | 0 | 0.0821 | 0 | 0.0039 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 373 SM pisc | 0.0679 | 0.0725 | 0.0432 | 0.0229 | 0 | 0.0821 | 0 | 0.0078 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 383 Crustaceans | 0.0001 | 0.2091 | 0.0000 | 0.0640 | 0 | 0.3483 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 393 Macrobenthos | 0.0257 | 0.0755 | 0.0000 | 0.0207 | 0 | 0.0268 | 0 | 0.1432 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 403 Meiobenthos | 0.0092 | 0 | 0.0000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 413 Zooplankton | 0.0538 | 0 | 0.0122 | 0.0007 | 1 | 0 | 0.3348 | 0.1101 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 421 Phytoplankton | 0.0021 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0117 | 0.0340 | 0 | 0.0345 | 0 | 0.10 |
| 432 Phytoplankton | 0.0144 | 0 | 0.0008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 443 Phytoplankton | 0.0466 | 0 | 0.0000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 45 Benthic plants | 0.0016 | 0 | 0.0002 | 0 | 0 | 0 | 0 | 0 | 0.0010 | 0 | 0 | 0.0003 | 0.1161 | 0.0001 | 0 |
| 46 Bigeye tuna | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 47 Yellowfin tuna | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 48 Marlins | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 49 Detritus | 0.0405 | 0.0086 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0092 | 0.0084 | 0 | 0.0327 | 0 | 0.55 |
| 50 Import | 0.0043 | 0.0809 | 0 | 0 | 0 | 0 | 0 | 0.0058 | 0 | 0 | 0.0001 | 0 | 0.0010 | 0.0002 | 0 |


| Prey \ predator | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 Oceanic shark | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 Coastal shark ray | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 Tuna-like | 0 | 0 | 0.0030 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 Coastal scombrids | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 Jellyfish | 0 | 0 | 0 | 0.0125 | 0 | 0 |  | 0.0053 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 Cephalopods | 0 | 0 | 0.0666 | 0.0631 | 0.0002 | 0.0236 | 0.0206 | 0.0016 | 0.0103 | 0 | 0 | 0 | 0 | 0 |
| 71 S bathy | 0 | 0 | 0.0147 | 0.0271 | 0 | 0.0414 | 0 | 0 | 0.0010 | 0 | 0 | 0 | 0 | 0 |
| 81 ML bathy | 0 | 0 | 0.0028 | 0 | 0 | 0.0027 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 91 Lpisc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 101 Carangids | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 111 S pelagics | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 121 L pisc comm | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 131 SM inv | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 141 SM pisc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 151 Macrobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 161 Meiobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 171 Zooplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 182 L pisc | 0 | 0 | 0.0316 | 0.0152 | 0 | 0.0121 | 0 | 0 | 0.0013 | 0 | 0 | 0 | 0 | 0 |
| 192 Carangids | 0 | 0 | 0.1352 | 0.0153 | 0.0032 | 0.0414 |  | 0.0019 | 0.0473 | 0 | 0 | 0 | 0 | 0 |
| 202 S pelagics | 0 | 0 | 0.2410 | 0.1579 | 0.0139 | 0.0606 |  | 0.0150 | 0.0590 | 0 | 0 | 0 | 0 | 0 |
| 212 L pisc comm | 0 | 0 | 0.0225 | 0.0110 | 0 | 0.0120 | 0.0260 | 0.0004 | 0.0045 | 0 | 0 | 0 | 0 | 0 |
| 222 Mikfish plus | 0 | 0 | 0.0329 | 0.0001 | 0 | 0.0128 | 0 | 0 | 0.0027 | 0 | 0 | 0 | 0 | 0 |
| 232 SM inv | 0 | 0 | 0.1220 | 0.2370 | 0.0188 | 0.1476 | 0.0260 | 0.0078 | 0.0842 | 0 | 0 | 0 | 0 | 0 |
| 242 SM pisc | 0 | 0 | 0.0496 | 0.0837 | 0.0123 | 0.1097 |  | 0.0072 | 0.0373 | 0 | 0 | 0 | 0 | 0 |
| 252 Crustaceans | 0 | 0 | 0.1610 | 0.1183 | 0.1609 | 0.2796 | 0.2440 | 0.1701 | 0.4933 | 0.05 | 0 | 0 | 0 | 0 |
| 262 Macrobenthos | 0 | 0 | 0.0229 | 0.0554 | 0.0331 | 0.1837 | 0.6571 | 0.4239 | 0.2080 | 0 | 0.05 | 0 | 0 | 0 |
| 272 Meiobenthos | 0 | 0 | 0 | 0.0004 | 0.0101 | 0 |  | 0.0473 | 0.0001 | 0.40 | 0.20 | 0 | 0 | 0 |
| 282 Zooplankton | 0 | 0 | 0.0245 | 0.0533 | 0.6810 | 0.0345 | 0.0003 | 0.1307 | 0.0123 | 0.15 | 0.10 | 0 | 0 | 0.2928 |
| 29 2,3 Hilsa | 0 | 0 | 0.0395 | 0.0576 | 0.0155 | 0.0127 | 0.0260 | 0.0025 | 0.0172 | 0 | 0 | 0 | 0 | 0 |
| 302,3 Indian mackerel | 0 | 0 | 0.0292 | 0.0711 | 0.0084 | 0.0253 |  | 0.0020 | 0.0210 | 0 | 0 | 0 | 0 | 0 |
| 313 L pisc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 323 Carangids | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 333 S pelagics | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 343 L pisc comm | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 353 Milkfish plus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 363 SM inv | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 373 SM pisc | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 383 Crustaceans | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 393 Macrobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 403 Meiobenthos | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 413 Zooplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0046 |
| 421 Phytoplankton | 0 | 0.9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 432 Phytoplankton | 0 | 0 | 0 | 0.0117 | 0.0340 | 0 |  | 0.0345 | 0 | 0 | 0.10 | 0 |  | 0.4974 |
| 443 Phytoplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0079 |
| 45 Benthic plants | 0 | 0 | 0.0010 | 0 | 0 | 0.0003 |  | 0.1161 | 0.0001 | 0 | 0 | 0 |  | 0.0160 |
| 46 Bigeye tuna | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 47 Yellowfin tuna | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 48 Marlins | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 49 Detritus | 1 | 0.1 | 0 | 0.0092 | 0.0084 | 0 |  | 0.0327 | 0 | 0.40 | 0.55 | 1 | 0.1 | 0 |
| 50 Import | 0 | 0 | 0 | 0 | 0.0001 | 0 |  | 0.0010 | 0.0002 | 0 | 0 | 0 | 0 | 0.1814 |


|  | Prey \predator | 30 | 30 | 32 | 33 | 34 |  | 35 |  | 36 | 37 | 38 | 39 | 40 | 41 | 46 | 47 | 48 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Oceanic shark | 0 | 00 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | Coastal shark ray | 0 | 00 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | Tuna-like |  | 00.0030 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0216 | 0 | 0.0313 |
| 4 | Coastal scombrids | 0 | 00 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0216 | 0 | 0.0744 |
| 5 | Jellyfish | 0 | 00 | 0.0125 |  | 0 | 0 |  |  | 0.0053 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | Cephalopods |  | 00.0666 | 0.0631 | 0.0002 | 20.0236 |  | 0.0206 |  | 0.0016 | 0.0103 | 0 | 0 | 0 |  | 0.0610 | 0.7741 | 0.1976 |
| 7 | 1 S bathy |  | 00.0147 | 0.0271 |  | 00.0414 |  |  | 0 |  | 0.0010 | 0 | 0 | 0 |  | 0.1946 | 0.0152 | 0 |
| 8 | 1 ML bathy |  | 00.0028 | 0 |  | 00.0027 |  |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.3243 | 0 | 0.0588 |
| 9 | 1 L pisc |  | 00 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0021 | 0 | 0.0025 |
| 10 | 1 Carangids |  | 00 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0021 | 0.0005 | 0.0040 |
| 11 | 1 S pelagics |  | 00 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0021 | 0.0005 | 0.0075 |
| 12 | 1 L pisc comm |  | 0 0 |  |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 13 | 1 SM inv |  | 00 |  |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0021 | 0.0020 | 0.0021 |
| 14 | 1 SM pisc |  | 00 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0011 | 0.0012 | 0.0016 |
| 15 | 1 Macrobenthos |  | $0 \quad 0$ |  |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0010 | 0 | 0.0010 |
| 16 | 1 Meiobenthos |  | $0 \quad 0$ | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 17 | 1 Zooplankton |  | $0 \quad 0$ |  |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 18 | 2 L pisc |  | $0 \quad 0$ | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0148 | 0 | 0.0170 |
| 19 | 2 Carangids |  | $0 \quad 0$ | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0148 | 0.0035 | 0.0277 |
| 20 | 2 S pelagics |  | 00 |  | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0148 | 0.0035 | 0.0514 |
| 21 | 2 L pisc comm |  | 0 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 22 | 2 Mikfish plus |  | 00 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0131 | 0 |
| 23 | 2 SM inv |  | $0 \quad 0$ | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0148 | 0.0136 | 0.0143 |
| 24 | 2 SM pisc |  | 00 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0074 | 0.0080 | 0.0107 |
| 25 | 2 Crustaceans |  | 00 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.0071 | 0 | 0.0071 |
| 26 | 2 Macrobenthos | 0.1509 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27 | 2 Meiobenthos | 0.0541 | 10 | 0 |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 28 | 2 Zooplankton | 0.1260 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 29 | 2,3 Hilsa | 0.0277 | 70.0395 | 0.0576 | 0.0155 | 50.0127 |  | 0.0260 |  | 0.0025 | 0.0172 | 0 | 0 | 0 |  | 0.0324 | 0.0152 | 0 |
| 30 | 2,3 Indian mackerel | 0.0277 | 70.0292 | 0.0711 | 0.0084 | 40.0253 |  |  |  | 0.0020 | 0.0210 | 0 | 0 | 0 |  | 0.0216 | 0.0152 | 0.0450 |
| 31 | 3 L pisc |  | 00.0316 | 0.0152 |  | 00.0121 |  |  | 0 |  | 0.0013 | 0 | 0 | 0 |  | 0.0479 | 0 | 0.0550 |
| 32 | 3 Carangids |  | 00.1352 | 0.0153 | 0.0032 | 20.0414 |  |  |  | 0.0019 | 0.0473 | 0 | 0 | 0 |  | 0.0479 | 0.0112 | 0.0896 |
| 33 | 3 S pelagics |  | 00.2410 | 0.1579 | 0.0139 | 90.0606 |  |  |  | 0.0150 | 0.0590 | 0 | 0 | 0 |  | 0.0479 | 0.0112 | 0.1663 |
| 34 | 3 L pisc comm |  | 00.0225 | 0.0110 |  | 00.0120 |  | 0.0260 |  | 0.0004 | 0.0045 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 35 | 3 Milkfish plus |  | 00.0329 | 0.0001 |  | 00.0128 |  |  | 0 |  | 0.0027 | 0 | 0 | 0 | 0 |  | 0.0424 | 0 |
| 36 | 3 SM inv |  | 00.1220 | 0.2370 | 0.0188 | 80.1476 |  | 0.0260 |  | 0.0078 | 0.0842 | 0 | 0 | 0 |  | 0.0479 | 0.0440 | 0.0462 |
| 37 | 3 SM pisc |  | 00.0496 | 0.0837 | 0.0123 | 30.1097 |  |  |  | 0.0072 | 0.0373 | 0 | 0 | 0 |  | 0.0240 | 0.0258 | 0.0346 |
| 38 | 3 Crustaceans |  | 00.1610 | 0.1183 | 0.1609 | 9.2796 |  | 0.2440 |  | 0.1701 | 0.4933 | 0.05 | 0 | 0 |  | 0.0229 | 0 | 0.0231 |
| 39 | 3 Macrobenthos | 0.1976 | 60.0229 | 0.0554 | 0.0331 | 10.1837 |  | 0.6571 |  | 0.4239 | 0.2080 | 0 | 0.05 | 0 | 0 | 0 | 0 | 0 |
| 40 | 3 Meiobenthos | 0.0709 |  | 0.0004 | 0.0101 |  | 0 |  |  | 0.0473 | 0.0001 | 0.40 | 0.20 | 0 | 0 | 0 | 0 | 0 |
| 41 | 3 Zooplankton | 0.1650 | 0.0245 | 0.0533 | 0.6810 | 0.0345 |  | 0.0003 |  | 0.1307 | 0.0123 | 0.15 | 0.10 | 0 | 0 | 0 | 0 | 0 |
| 42 | 1 Phytoplankton |  | $0 \quad 0$ | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 43 | 2 Phytoplankton | 0.0541 | 1 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 44 | 3 Phytoplankton | 0.0709 |  | 0.0117 | 0.0340 |  | 0 |  |  | 0.0345 | 5 | 0 | 0.10 | 0 | 0.9 | 0 | 0 | 0 |
| 45 | Benthic plants |  | 00.0010 |  |  | 00.0003 |  |  |  | 0.1161 | 0.0001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 46 | Detritus |  | 00 |  |  | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0156 |
| 47 | Import |  | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0156 |

## A4. Calculation of natural mortality.

The calculation is based on growth parameters and Pauly's equation ( Mp ) and maximum age and Hoenig's equation (Mh). See Excel spreadsheet Appendix 4.xls.

## A5 Marine and inland catches by country for hilsa.

Table A6. Catches of hilsa as compiled by the UBC Fisheries Centre Sea Around Us Project Team (SAUP) and the inland catch from FAO statistics, compared with the catches for Bangladesh (BGD) compiled for the BOBLME stock assessement (BOB BGD). The shaded cells indicate extrapolation.

|  |  | SAUP, marine catch |  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |


| 2007 | 231,689 | 29,024 | 20,332 | 281,045 | 82,445 | 11,721 | 375,211 | 314,134 | 280,328 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 235,630 | 26,702 | 20,536 | 282,868 | 89,900 | 14,233 | 387,001 | 325,530 | 290,000 |
| 2009 | 238,999 | 24,127 | 16,315 | 279,441 | 95,970 | 12,381 | 387,792 | 334,969 | 298,458 |
| 2010 | 233,860 | 81,609 | 16,399 | 331,868 | 115,179 | 7,260 | 454,307 | 349,039 | 312,612 |

[^4]
## A6. Alternative fitting results with Ecosim

Fitting of data using the Indian relative industrial effort excluding large trawlers (region 2) labeled "industrial only" in A2.2. The most remarkable difference with the other fitting exercise is the trend in Hilsa biomass.


Figure A6.1. Observed (dots) and predicted (lines) relative abundance (B or CPUE) and catches (C) for hilsa, marlins, yellowfin tuna (YFT) and bigeye tuna (BET). The numbers besides the name is the sum of squares. The region 2 industrial fleet includes industrial vessels only (Industrial only time series in A2.2).


Figure A6.2. Observed (dots) and predicted (lines) catches for other species. The numbers besides the name is the sum of squares. The region 2 industrial fleet includes industrial vessels only (Industrial only time series in A2.2).

## 2 L pisc






Figure A6.3. Predicted trends in biomass, catches, and mortality (lines) and observed catches (dots) for the functional group 2 L pisc. The solid green areas represent the catch or fishing mortality caused by the each of the fleet in region 2 . The region 2 industrial fleet includes industrial vessels only (Industrial only time series in A2.2).


[^0]:    ${ }^{\text {a }}$ computed using the MSY estimate and Cadima's equation
    ${ }^{\mathrm{b}} 16 \%$ of total biomass of sampled groups
    ${ }^{\text {c }} 2010$ only from SAUP compared with the estimate of 1140 t in (Darwin Reef Fish Project 2011).

[^1]:    ${ }^{1}$ FishStatJ 2010, http://www.fao.org/fishery/statistics/software/fishstat/en

[^2]:    a. based on list of species in the Andaman catch data base
    b. for the catch from the tuna database only

[^3]:    $a$ based on the Japanese index

[^4]:    ${ }^{\text {a }}$ shaded area are extrapolated from the earlier year of estimation: 1984 in Bangladesh and 1992 in India
    ${ }^{\mathrm{b}}$ catches for inland and marine catches estimated for Bangladesh (BOBLME 2012)

